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A Quantitative Structure-Property Relationship (QSPR) for Estimating Solid Material-Air Partition Coefficients of Organic Compounds

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29 Abstract

30 The material-air partition coefficient (K_{ma}) is a key parameter to estimate the release of
31 chemicals incorporated in solid materials and resulting human exposures. Existing correlations to
32 estimate K_{ma} are applicable for a limited number of chemical-material combinations without
33 considering the effect of temperature. The present study develops a quantitative property-
34 property relationship (QSPR) to predict K_{ma} for a large number of chemical-material
35 combinations. We compiled a dataset of 991 measured K_{ma} for 179 chemicals in 22 consolidated
36 material types. A multiple linear regression model predicts K_{ma} as a function of chemical's K_{oa} ,
37 enthalpy of vaporization (ΔH_v), temperature and material type. The model shows good fitting of
38 the experimental dataset with adjusted R^2 of 0.93 and has been verified by internal and external
39 validations to be robust, stable and has good predicting ability ($R^2_{ext} > 0.78$). A generic QSPR is
40 also developed to predict K_{ma} from chemical properties and temperature only (adjusted $R^2 =$
41 0.84), without the need to assign a specific material type. These QSPRs provide correlation
42 methods to estimate K_{ma} for a wide range of organic chemicals and materials, which will
43 facilitate high-throughput estimates of human exposures for chemicals in solid materials,
44 particularly building materials and furniture.

45 Keywords

46 Partitioning, Indoor release, Solid materials, Organic chemicals, Consumer exposure, Correlation

47 Practical implications

48 The developed QSPRs provide a comprehensive correlation method to estimate K_{ma} , covering a
49 much wider range of organic chemicals and solid materials compared to previous studies, and
50 include a still accurate generic correlation without the need to assign a material type. Combined
51 with the QSPR estimating the internal diffusion coefficient¹, these QSPRs facilitate high-
52 throughput estimates of indoor human exposures to chemicals incorporated in solid materials.
53 This is highly relevant for multiple science-policy fields, including chemical alternatives
54 assessment (CAA), risk assessment (RA) and life cycle assessment (LCA).

56 1. Introduction

57 Chemicals incorporated in solid materials have been identified as a major source of passive
58 emissions to indoor air and of transfers into house dust and skin. Typical examples include

59 chemicals used as plasticizers in building materials and flame retardants in furniture. To estimate
60 the release of these chemicals from solid materials and subsequent consumer exposures, the
61 dimensionless solid material-air partition coefficient (K_{ma}), defined as the ratio of the
62 concentration in the material to the concentration in the air at equilibrium, is one of the key
63 parameters². The K_{ma} is essential in determining the chemical transfer from solid material to air
64 and to house dust, as well as the chemical concentration at the material surface, which further
65 determines the inhalation, dermal and dust ingestion exposures. K_{ma} is specific to a chemical-
66 material combination and is also influenced by ambient temperature. Experimental techniques
67 such as chamber tests for building materials³, and sorption experiments for polymer materials⁴⁻⁶
68 have enabled measurement of a limited number of K_{ma} values for building materials such as
69 vinyl flooring, gypsum board, plywood and cement, as well as polymer materials used for
70 passive samplers including polyurethane foams (PUF), polyethylene (PE), and polypropylene
71 (PP). Recently, studies have also been conducted to measure the K_{ma} for clothing and fabrics^{7,8}.
72 However, since experiments are costly and time-consuming, measured K_{ma} values are only
73 available for a limited number of chemical-material combinations. Thus, quantitative
74 relationships are needed to predict this partition coefficient from known physiochemical
75 properties for chemicals without experimental data, which is especially important for high-
76 throughput approaches, for which a large number of chemical-material combinations need to be
77 evaluated.

78 Several correlation methods have been developed to estimate K_{ma} from physiochemical
79 properties of chemicals. For example, several studies have correlated K_{ma} to the chemical's
80 vapor pressure using data on volatile organic compounds (VOCs) in building materials^{4,9-11}.
81 Other studies which focused on semi-volatile organic compounds (SVOCs) in passive sampling
82 devices have found correlation between K_{ma} and the octanol-air partition coefficient (K_{oa})^{5,6,12,}
83¹³. Furthermore, Holmgren et al. estimated K_{ma} as a function of five Abraham solvation
84 parameters for six groups of materials¹⁴, but these parameters are not readily available. For the
85 aforementioned approaches, the main limitation is that the correlations are specific to certain
86 chemical classes and materials; for example polycyclic aromatic hydrocarbons (PAHs) in low-
87 density polyethylene (LDPE), which limits their application for other chemical-material
88 combinations. Addressing this research gap to facilitate wider applicability, Guo developed a
89 method which estimates the K_{ma} as a function of the chemical's vapor pressure for all materials

90 and chemical classes ¹¹. However, this approach is developed based on a small dataset which
91 mainly includes VOCs in building materials which limits its applicability to also address SVOCs.
92 Another limitation of the previous studies is that the effect of temperature was not well
93 considered in the correlation. Some studies provided different correlation coefficients for certain
94 discrete temperatures ¹⁵, while others corrected the predictors for temperature ¹⁶. However, since
95 the known physiochemical properties such as vapor pressure and K_{oa} are often only given as
96 values at 25 °C, correcting them for temperature may not always be practical as the
97 corresponding enthalpies of phase change are not available for all chemicals. Several studies did
98 establish correlations between K_{ma} and temperature, but the correlations were only verified using
99 experimental data on limited chemicals such as formaldehyde and other aldehydes ^{17, 18}.

100 In all, the currently available correlation methods to estimate K_{ma} do not provide sufficient
101 coverage of chemicals incorporated in solid materials at different ambient temperatures. A recent
102 research hotspot in exposure sciences is to develop low tier, high-throughput methods to estimate
103 exposure to chemical in consumer products across a variety of chemical-material combinations,
104 which requires high-throughput estimates of K_{ma} for a wide range of material-chemicals
105 combinations. Thus, the present study aims to develop a more comprehensive correlation method
106 to estimate K_{ma} for a wide range of organic compounds in multiple solid materials, addressing
107 the need for high-throughput exposure assessments. More specifically, we aim to:

- 108 1) Carry out a comprehensive literature review to collect experimental K_{ma} data on a wide range
109 of materials and chemicals.
- 110 2) Use multiple linear regression techniques to establish the relationship between K_{ma} and
111 various predictor variables including physiochemical properties, material type and temperature.
- 112 3) Perform internal and external validations to characterize the validity and predictive power of
113 the developed correlation.

114 This QSPR provides a more advanced correlation method to estimate the K_{ma} of organic
115 compounds compared to previous studies, as it covers a wide range of solid materials and
116 chemicals, and consistently incorporates the effect of temperature. A similar QSPR has been
117 developed by our group for the internal diffusion coefficient in solid materials ¹. By providing
118 reliable estimates of the key partition and diffusion parameters for a large number of material-
119 chemical combinations, these QSPRs will facilitate high-throughput assessments of chemical
120 emissions and human exposures for chemicals incorporated in solid materials relevant for

121 various science-policy fields such as chemical alternatives assessment (CAA), risk assessment
122 and life-cycle assessment (LCA).

123

124 2. Materials and Methods

125 2.1 Dataset

126 2.1.1 Data collection

127 Experimental material-air partition coefficient data were compiled from 43 references from the
128 peer-reviewed scientific literature (provided in Supporting Information (SI), Section S1).

129 Dimensionless partition coefficients were collected. If the partition coefficients were expressed
130 in mL/g or m³/g, they were converted to dimensionless values by multiplying these by the
131 density of the solid material. If the partition coefficients were expressed in the unit of m, they
132 were converted to dimensionless values by dividing these by the thickness of the material. The
133 initial dataset of K_{ma} contained a total of 1008 records covering 179 unique chemicals and 75
134 distinct solid materials.

135 2.1.2 Data curation

136 For the 179 unique chemicals of the initial K_{ma} dataset, molecular weight, vapor pressure, water
137 solubility and $\log K_{ow}$ at 25 °C were obtained from EPISuite¹⁹. For these physiochemical
138 properties, experimental values were used when available, otherwise the software-estimated
139 values were used. The enthalpy of vaporization (ΔH_v , J/mol) of each chemical was obtained
140 from ChemSpider estimated values (www.chemspider.com).

141 For the octanol-air partition coefficient ($\log K_{oa}$) at 25 °C, experimental values are only available
142 for part of the 179 chemicals in the dataset. To avoid inconsistency, we used the $\log K_{oa}$ values
143 estimated by EPISuite¹⁹ for all of the 179 chemicals. In EPISuite, $\log K_{oa}$ is estimated by
144 subtracting $\log K_{aw}$ (dimensionless log air-water partition coefficient) from $\log K_{ow}$, $\log K_{aw}$ and
145 $\log K_{ow}$ being estimated by the HenryWin and KowWin functions, respectively¹⁹. Experimental
146 $\log K_{oa}$ values were also collected and their impacts on the QSPR were assessed, as presented in
147 SI Section S6.

148 To avoid over-fitting of the QSPR model, the 75 original materials for K_{ma} were grouped into 22
149 consolidated material types, based on the name of the materials and the similarity of the
150 regression coefficients (see SI Section S1), ensuring a minimum of 5 data points and 3 different

151 chemicals per consolidated material type. The data points with materials that cannot be grouped
152 according to the above criteria were excluded from further analyses.

153 The final K_{ma} dataset contains 991 data points with 179 unique chemicals in 22 consolidated
154 material types. The temperature at which the K_{ma} was measured ranges from 15 to 100 °C. The
155 final dataset is provided in SI.

156

157 2.2 Modeling methods

158 2.2.1 Multiple linear regression model

159 A multiple linear regression (MLR) analysis was performed to identify and quantify the effect of
160 different parameters on the partition coefficient, with details described in our previous paper on
161 the QSPR for diffusion coefficient¹. Briefly, the MLR model takes the following general form:

$$162 \quad \log_{10}K_{ma} = \alpha + \beta_1 \cdot X_1 + \dots + \beta_n \cdot X_n + b_1 \cdot M_1 + \dots + b_m \cdot M_m \quad (1)$$

163 where $\log_{10}K_{ma}$ is the logarithm of the dimensionless K_{ma} , α is the intercept; X_1 to X_n are
164 independent variables related to the properties of the chemical or the environment; β_1 to β_n are
165 regression coefficients for the respective independent variables X_1 to X_n . M_1 to M_m are dummy
166 variables for the packaging materials, with one dummy variable per type of material. A dummy
167 variable equals 1 for the material type it represents, and equals 0 for all other materials; for
168 example, $M_1 = 1$ for material type 1, $M_1 = 0$ for material types 2 to m . b_1 to b_m are regression
169 coefficients for the respective dummy variables M_1 to M_m . The number of m is equal to the
170 number of material types considered minus 1, since PU-ether - the material type with the highest
171 number of measured K_{ma} data - is used as the reference material type and does not require a
172 dummy available in the MLR. Regression coefficients were estimated by the least squares (LS)
173 method. All regression analyses were performed using IBM SPSS Statistics version 23 (IBM
174 corporation, Armonk, New York).

175 In previous studies, either the chemical's vapor pressure^{4, 9-11} or $\log K_{oa}$ ^{5, 6, 12, 13} has been used as
176 predictor of the K_{ma} in a given material. Abraham solvation parameters were also used as
177 predictors by Holmgren et al.¹⁴, but these parameters are not considered here since they are not
178 readily available. Initial regressions (SI Section S2) suggest that $\log K_{oa}$ is a better predictor of
179 K_{ma} compared to vapor pressure. Thus, the chemical's $\log K_{oa}$ at 25 °C was used as the
180 independent variable for chemical properties in Eq. 1.

181 Thus, the MLR model takes the following form:

182 $\log_{10}K_{ma} = \alpha + \beta_{\log K_{oa}} \cdot \log_{10}K_{oa} + \beta_T \cdot T_term + b_1 \cdot M_1 + \dots + b_{21} \cdot M_{21}$ (2)

183 where T_term is a term representing the effect of temperature and will be described in the next
 184 section (Section 2.2.2).

185 2.2.2 Temperature dependence

186 In thermodynamics, the temperature dependence of equilibrium constant, K_{eq} , can be described
 187 by the van't Hoff equation:

188
$$\ln \frac{K_2}{K_1} = \frac{\Delta H_{phase\ change}}{R} \left(\frac{1}{T_2} - \frac{1}{T_1} \right)$$
 (3a)

189 where K_1 and K_2 are the equilibrium constants at temperature T_1 and T_2 , respectively, T_1 and T_2
 190 are absolute temperatures (K), R is ideal gas constant (8.314 J/(K·mol)), and $\Delta H_{phase\ change}$ is the
 191 enthalpy of phase change (J/mol).

192 Since K_{ma} is an equilibrium constant by definition and the chemical's $\log_{10}K_{oa}$ at 25 °C or
 193 298.15 K is used as an independent variable in the MLR model (Eq. 2), we assume that the
 194 temperature dependence of K_{ma} also follows the van't Hoff equation:

195
$$T_term = \log_{10} \frac{K_{ma,2}}{K_{ma,1}} = \frac{\Delta H_{ma}}{2.303 \cdot R} \left(\frac{1}{T_2} - \frac{1}{298.15} \right)$$
 (3b)

196 where ΔH_{ma} is the enthalpy of the partitioning between material and air (J/mol), and 2.303 is a
 197 conversion factor between $\log_{10}K$ and $\ln K$.

198 Ideally, the enthalpy ΔH_{ma} should be different for different chemical-material combinations.

199 Kamprad and Goss have determined the ΔH_{ma} values for 54 unique chemicals in PU-ether using
 200 measured K_{ma} data from 15 °C to 95°C⁴, so we were able to develop a linear correlation to

201 estimate ΔH_{ma} from chemical properties (results shown in Section 3.1). Since no experimental

202 ΔH_{ma} values are available for materials other than PU-ether, we use the ΔH_{ma} correlation

203 developed above across all materials. Therefore, in our regression model of K_{ma} , the ΔH_{ma} is

204 chemical-specific, but not material-specific. The final MLR model thus takes the following form:

205
$$\log_{10}K_{ma} = \alpha + \beta_{\log K_{oa}} \cdot \log_{10}K_{oa} + \beta_T \cdot \frac{\Delta H_{ma}}{2.303 \cdot R} \left(\frac{1}{T} - \frac{1}{298.15} \right) + b_1 \cdot M_1 + \dots + b_{20} \cdot M_{20}$$
 (4)

206

207 2.3 Model validation

208 Validation of the final MLR model (Eq. 4) was performed using the QSARINS software, version

209 2.2.1 (www.qsar.it) which is developed by Gramatica et al.^{20, 21}.

210 2.3.1 Internal validation

211 The MLR model's capacity to predict portions of the training dataset was evaluated in an internal
212 validation process, using two techniques in QSARINS: the leave more out (LMO) cross-
213 validation and the Y-scrambling, which have been described previously^{1,21}. 1000 iterations were
214 used for the LMO cross-validation and the percentage of the excluded elements was set as 20%,
215 and 1000 iterations for Y-scrambling.

216 2.3.2 External validation

217 We also evaluated the model's ability to provide reliable predictions on new datasets by external
218 validation, using the splitting approach, which split the existing dataset (991 data points) into one
219 training dataset and one prediction dataset. The training dataset was used to generate regression
220 coefficients of the MLR model, and then the MLR model was applied to the prediction set to
221 examine the prediction performances of the model. Three kinds of splitting were performed
222 using existing options in the QSARINS software (see SI Section S4.1 for details) by random
223 percentage, by ordered response and by structure. We introduced a fourth kind of splitting by
224 studies, where all data points from certain studies were manually selected as the training set and
225 data points from remaining studies as the prediction set. If a consolidated material type only
226 includes data points from one study, all of these data points were assigned into the training set in
227 order to ensure that the MLR model constructed using the training set includes all consolidated
228 material types. The four types of splitting yielded similar sample sizes of approximately 800 data
229 points for the training set and 200 data points for the prediction set (SI Table S3).

231 3. Results and Discussions

232 3.1 Temperature dependence

233 As described in Section 2.2.2, the temperature dependence of K_{ma} is determined by the enthalpy
234 of the partitioning between material and air, ΔH_{ma} (J/mol). Using the measured K_{ma} data for 54
235 chemicals in PU-ether from 15 °C to 95 °C⁴ (data are provided in SI Section S3), we obtained the
236 following correlation to estimate ΔH_{ma} :

$$237 \quad \Delta H_{ma} = 1.37 \cdot \Delta H_v - 14.0 \quad (5)$$

$$238 \quad N = 54, R^2 = 0.786, R^2_{adj} = 0.782, SE = 2.85, RMSE = 2.80$$

$$239 \quad \text{ANOVA: } F = 191, df = 1, p < 0.0001$$

240 where ΔH_v is the chemical's enthalpy of vaporization (J/mol) obtained from ChemSpider
241 (www.chemspider.com).

242 This simple linear model shows good fitting of the experimental ΔH_{ma} data, with an adjusted R-
243 squared of 0.782, and the model fit is highly significant with an ANOVA p-value < 0.0001.

244 Figure 1 shows the scatter plot of predicted vs measured ΔH_{ma} and the residual plot, which
245 indicate good agreement with the 1:1 line and random distribution of residuals throughout the
246 dataset. These results suggest that there is indeed a linear relationship between ΔH_{ma} and ΔH_v in
247 PU-ether, and Eq.5 was also used as default to estimate ΔH_{ma} for all other materials.

248

249 3.2 Final QSPR and model fitting

250 Using the full dataset (991 data points) and Eq. 4, the final MLR model for predicting the solid
251 material-air partition coefficient is as follows:

$$252 \log_{10}K_{ma} = -0.38 + 0.63 \cdot \log_{10}K_{oa} + 0.96 \cdot \frac{\Delta H_{ma}}{2.303 \cdot R} \left(\frac{1}{T} - \frac{1}{298.15} \right) + b \quad (6)$$

$$253 N = 991, R^2 = 0.934, R^2_{adj} = 0.933, SE = 0.62, RMSE = 0.62$$

$$254 ANOVA: F = 597, df = 23, p < 0.0001$$

255 where K_{ma} is the dimensionless solid material-air partition coefficient, K_{oa} is the chemical's
256 dimensionless octanol-air partition coefficient at 25 °C, ΔH_{ma} is the enthalpy of the partitioning
257 between material and air (J/mol) which is given by Eq. 5, T is absolute temperature (K), and b is
258 the material-specific coefficients presented in Table 1. This model is provided as an excel model
259 in SI to facilitate application. The standard errors for the coefficients are also presented in Table
260 1. An SE of 0.63 of the final model (Eq. 6) indicates that the 95% confidence interval (CI) of the
261 predicted $\log K_{ma}$ is the predicted value ± 1.22 , indicating that most of the predicted K_{ma} are
262 within a factor of 16 from the measured K_{ma} .

263 This MLR model shows excellent fitting of the experimental data, with an adjusted R-squared of
264 0.93 and a root mean square error (RMSE) of 0.62. The model fit is highly significant with an
265 ANOVA p-value smaller than 0.0001. Figure 2A shows the scatter plot of predicted vs measured
266 $\log K_{ma}$, which aligns well with the 1:1 line. The residual plot (Figure 1B) shows that the
267 residuals are distributed evenly throughout the dataset, and most residuals have absolute values
268 smaller than 2, again indicating the good fit of the linear model for the data.

269 This MLR model assumes that the correlation between $\log K_{ma}$ and the chemical's $\log K_{oa}$ is the
270 same across material types, which seems reasonable given the excellent model fitting. Plotting

271 the $\log K_{ma}$ against chemical's $\log K_{oa}$ for selected material types (Figure 3) confirmed that the
272 correlation between $\log K_{ma}$ and the chemical's $\log K_{oa}$ (i.e., the slopes of the fitted straight lines
273 in Figure 3) is similar but with slight differences across material types, indicating that a single
274 coefficient for $\log K_{oa}$, as in the present QSPR model, might not be perfect. This could have been
275 accounted for by including interaction terms between $\log K_{oa}$ and material types, but this would
276 introduce 21 more terms in the model without greatly improving the model fitting (SI Section
277 S5), so the interaction terms were not retained in the final QSPR model.

278 As described in the methods, this final MLR model uses EPISuite-estimated $\log K_{oa}$ values as
279 predictors, since experimental $\log K_{oa}$ are not available for all chemicals in the dataset. MLR
280 models developed using mixed $\log K_{oa}$ values (i.e., for a chemical experimental $\log K_{oa}$ is used
281 when available, otherwise EPISuite-estimated $\log K_{oa}$ is used) also yielded similar results as the
282 final MLR model (adjusted R^2 ranged from 0.930 to 0.931, for details see SI Section S6),
283 indicating that the impact of experimental $\log K_{oa}$ on the model is minimal.

284

285 3.3 Impact of each predictor

286 As shown in Eq. 6, the key predictors of the solid material-air partition coefficient are the
287 chemical's $\log K_{oa}$, ΔH_v , temperature, and the solid material type. The regression coefficient for
288 $\log K_{ow}$ is 0.63 and is highly significant ($p < 0.0001$), indicating that the material-air partition
289 coefficient increases with increasing $\log K_{oa}$, which is consistent with findings from previous
290 studies^{5, 6, 13}.

291 The regression coefficient of the temperature term is 0.96 and is also highly significant ($p <$
292 0.0001), indicating that the K_{ma} decreases with higher temperature. Experimental data from
293 Kamprad et al. did show reduced K_{ma} with increased temperature, and it also makes intuitive
294 sense that at higher temperature the K_{ma} is lower leading to faster chemical migration from solid
295 material to air. As discussed in Section 3.1, the effect of temperature on K_{ma} also depends on the
296 ΔH_{ma} , which increases linearly with the chemical's enthalpy of vaporization ΔH_v .

297 The 21 dummy variables for the material types reflect the material dependency of the K_{ma} . As
298 "PU-ether" (polyurethane-ether) was used as the reference material in the regression, the value of
299 its coefficient b is zero (Table 1). For each of the other material types, the coefficient b ,
300 determines the difference in $\log K_{ma}$ between that material type and PU-ether. Chemicals in solid
301 material types with high values of b are more difficult to migrate to air than in those with low

302 values of b . The three material types with highest b coefficients are ethylene vinyl acetate (EVA),
303 latex and solvent-based paint and polyether ether ketone (PEEK) which are dense materials,
304 while the three types with lowest b coefficients are PU-ester (polyurethane-ester), PU-ether and
305 paper which tend to be porous materials. It should be noted that the data for a given consolidated
306 material type were gathered from different studies, and the composition and properties of the
307 material type may vary between studies, so the material type coefficients in Table 1 only
308 represent an average composition and partition behavior for the specific material types.
309 The significance of the material type coefficient only indicates that the coefficient b s of these
310 material types are significantly different from the reference material type, PU-ether, but if
311 another material type was selected as the reference material, the regression coefficients and
312 statistical significance of all materials would change. Thus, the insignificance of the regression
313 coefficient for “paper” (Table 1) does not indicate that this material type does not have a relevant
314 influence on the K_{ma} . As a result, we keep all 21 material type dummy variables in the final
315 regression to retain as much information as possible.

316 To better illustrate the impact of each predictor on the material-air partition coefficient, we
317 varied each predictor from the minimum to the maximum value in the entire dataset (991 data
318 points) while keeping the other predictors constant, and calculated the change in $\log K_{ma}$ using
319 the regression coefficients in the final QSPR (Eq. 6). Since the chemical’s ΔH_v determines the
320 ΔH_{ma} which modifies the relationship between $\log K_{ma}$ and temperature, the impact of
321 temperature was calculated as two extremes using the minimum and maximum values of ΔH_v in
322 the entire dataset. As shown in Figure 4, the chemical’s $\log K_{oa}$ has the highest impact on $\log K_{ma}$
323 among predictors. The impact of temperature on $\log K_{ma}$ is very low with the lowest value of
324 ΔH_v (22.3 kJ/mol), but the impact become moderate with the highest value of ΔH_v (75.6 kJ/mol).
325 This indicates that for a chemical with low enthalpy of vaporization, the $\log K_{ma}$ only changes
326 slightly with temperature, and vice versa. The material type also has a moderate impact on the
327 $\log K_{ma}$, which is similar to the impact of temperature with the highest value of ΔH_v . Overall, the
328 impact of material type is relatively small compared to the impact of chemical’s $\log K_{oa}$,
329 indicating that the variation in $\log K_{ma}$ does not strongly depend on the solid material type, which
330 suggests the possibility of developing a generic QSPR to predict $\log K_{ma}$ in absence of material-
331 specific data.

332

333 3.4 Model validation results

334 3.4.1 Internal validation

335 The correlation coefficient for the LMO cross validation, Q^2_{LMO} , averages 0.93 (range: 0.90 –
336 0.95) for the 1000 iterations, and the root mean square error for cross validation ($RMSE_{cv}$)
337 averages 0.63. Both the Q^2_{LMO} and $RMSE_{cv}$ are similar to the R^2 and $RMSE$ computed using the
338 full dataset, which is 0.93 and 0.62, respectively, indicating that the model is internally stable.
339 For Y-scrambling, the R^2_{Yscr} , Q^2_{Yscr} and $RMSE_{Yscr}$ for the 1000 iterations average 0.023, -0.028,
340 and 2.37, respectively, which are substantially different from the R^2 , Q^2_{LMO} and $RMSE$ of the
341 original model, indicating that that no correlation exists between the scrambled responses and the
342 predictors. Thus, the internal validation overall demonstrates that the final QSPR model (Eq. 6)
343 is robust and stable, and is not a result of chance correlation.

344 3.4.2 External validation

345 As described in Section 2.3.2, four types of splitting were used for external validation, including
346 splitting by random 20%, by ordered response, by structure, and by studies. Six criteria for
347 external validation, described in detail previously^{1, 22, 23}, were computed and are presented in
348 Table 2. For the first three types of splitting, the R^2_{ext} are higher than 0.9, and the other five
349 criteria all pass the threshold values and are higher than 0.9, indicating good predictive ability of
350 the models constructed from training set data. This is expected because the prediction sets
351 resulted from these three types of splitting are generally well within the applicability domain
352 (described in detail below) defined by the training sets (SI, Figures S1-S6), since the data points
353 were drawn either randomly or alternately.

354 For the splitting by studies, data from 22 studies were selected as the prediction set, while data
355 from 20 studies constituted the training set. This splitting can better represent a truly “external”
356 validation, since all data from one study were either be in the training or the prediction set. the
357 prediction ability of the model constructed from the training set is apparently reduced, as the
358 R^2_{ext} of this splitting dropped to 0.79, and the values of the other five criteria are lower than
359 those for the above three types of splitting. This is reasonable since the data variability is higher
360 between studies than within studies, so the prediction set might not be well within the AD
361 defined by the training set (SI, Figures S7-S10). Nonetheless, all validation criteria for this
362 splitting still pass the thresholds, indicating acceptable prediction ability (Table 2).

363 3.4.3 Applicability domain (AD)

364 It is important to define the AD of our QSPR model, as it can provide information on the
365 reliability of the model predictions²⁴ for future users who would like to use the model on new
366 chemicals. If the new chemicals are inside the AD, the model predictions are interpolated and are
367 more reliable. However, if the chemicals are outside the AD, the predictions are extrapolated and
368 less reliable²⁴.

369 For definition of the AD, the model being evaluated is the final QSPR model presented in Eq. 6,
370 and the training dataset thus refers to the full dataset including 991 data points. Three
371 complementary methods were applied to define the AD of the K_{ma} QSPR: the range of model
372 predictors, the leverage approach, and the PCA of the model predictors, which have been
373 described in detail previously²⁵.

374 For the range of predictors, the model has four predictors: $\log K_{oa}$, ΔH_v , temperature and material
375 type. The $\log K_{oa}$, ΔH_v , temperature of the training dataset range from 1.4 to 14.6, from 22.3 to
376 75.6 kJ/mol, and from 15 to 100 °C, respectively, defining the AD of the model. It is noteworthy
377 that the material type is a categorical variable, and the training set contains 22 consolidated
378 materials types, so the model's AD is also restricted to these 22 material types. For the leverage
379 approach, the critical value h^* for the diagonal values of the hat (h) matrix of the model was
380 calculated to be 0.0727, and the AD is defined as the h values less than h^* ^{21, 25}. For the PCA
381 approach, the AD is defined as the space between the minimum and maximum values of the PC1
382 and PC2 scores of the training dataset^{21, 25}, which range from -4.39 to 2.04 and from -4.52 to
383 2.22, respectively. For future model users, a new chemical should be considered "inside AD" if
384 viewed inside AD by all three methods, and be considered "outside AD" if viewed out of AD by
385 all three methods, otherwise it should be considered "borderline"²⁵.

386 387 3.5 Generic QSPR

388 In order to predict the K_{ma} without assigning material properties, we built a generic QSPR model
389 which does not include any material-specific variables using the same dataset. This model only
390 uses the chemical properties and temperature as predictors and is as follows:

$$391 \quad \log_{10} K_{ma} = -0.37 + 0.75 \cdot \log_{10} K_{oa} + 1.29 \cdot \frac{\Delta H_{ma}}{2.303 \cdot R} \left(\frac{1}{T} - \frac{1}{298.15} \right) \quad (7)$$

$$392 \quad N = 991, R^2 = 0.80, R^2_{adj} = 0.80, SE = 1.08, RMSE = 1.08$$

$$393 \quad \text{ANOVA: } F = 1943, df = 2, p < 0.0001$$

394 This model has a still relatively high adjusted R-squared of 0.80 compared to the 0.93 of the
395 regression with material coefficient (Eq. 6), indicating a good fit of experimental data (Figure 5).
396 As discussed in Section 3.2, the impact of the solid material type on $\log K_{ma}$ is relatively small
397 compared to the impact of chemical properties, so $\log K_{ma}$ can be predicted with reasonably high
398 accuracy without the material type as a predictor. This generic QSPR thus provides a relatively
399 reliable method to estimate the K_{ma} for various solid materials that may be difficult to assign a
400 material type listed in Table 1, which provides a more comprehensive and flexible coverage,
401 although with a slightly lower accuracy, for different chemical-material combinations than the
402 material-specific QSPR and can therefore greatly facilitate high-throughput evaluations of a
403 large variety of chemical-material combinations. However, it should be noted that although
404 without the material type as a predictor, this generic model was still developed using the
405 experimental data of our collection of 22 material types. Thus, this generic model best applies to
406 materials listed in Table 1 and similar materials, but may cause a large error for materials with
407 special properties, e.g. in presence of strong ionic forces, or of strong pseudo-solvation such that
408 some of the target adsorbate molecules take on a different structure within the material itself,
409 either due to ionization or tautomerization.

410 3.6 Limitations and future work

411 While the coverage of 22 consolidated materials and possibly any solid material as well as
412 inclusion of the effect of temperature are major advantages, the present model has several
413 limitations. First, the model does not consider chemical ionization or interaction with other
414 chemicals within a solid material, which may affect the chemical's partitioning between the
415 material and air. Second, the present model assumes that the relationship between ΔH_{ma} and
416 chemical's ΔH_v , derived from experimental ΔH_{ma} data for one material type "PU-ether", is the
417 same across different material types. Ideally, more experimental ΔH_{ma} data for different material
418 types are needed to verify this assumption or to develop unique ΔH_{ma} - ΔH_v relationships for
419 different material types.

420 Third, since for most K_{ma} datasets the material properties are not well characterized or provided
421 in the original publications, the classification of the consolidated material types is qualitative and
422 is simply based on material names, which may result in considerable variations in material
423 properties within one consolidated material type. In addition, even with the same composition,
424 different material structure may affect the material-air partitioning. Ideally, quantitative,

425 continuous properties of the solid materials, such as descriptors of the material's composition
426 and molecular structure, could be measured and entered into the model as numerical predictors,
427 so that the model can be more accurate for particular materials and can be extrapolated to other
428 material types outside the training dataset. In addition, if quantitative variables for material types
429 are used, interaction terms between chemical's $\log K_{oa}$ and material type variables can be added
430 to the model without introducing too many additional terms, which can improve model fitting, as
431 discussed in Section 3.2.

432 Fourth, many materials that appear in indoor environments are inhomogeneous, such as plywood,
433 gypsum board, carpet, concrete, and paper, which may have layers or portions with distinctive
434 properties. Thus, the K_{ma} values measured in experiments and the QSPR built on these
435 measurements likely only represent the material properties across the experiments. As a result,
436 one needs to use caution when applying the present QSPR to predict K_{ma} , especially for highly
437 inhomogeneous materials. Another important aspect related to heterogeneity is surface
438 partitioning versus bulk partitioning. Since the partitioning between solid material and air
439 happens mainly at the material surface, the surface properties may have an unusually large
440 influence on the apparent partitioning behavior. Therefore, for materials with a surface layer of
441 distinct properties, or materials with the same composition but different surface/bulk structures,
442 the present QSPR may not give a correct estimate of the K_{ma} . The distinct surface layer may be a
443 result of oxidative aging and soiling, which may change with time, or intrinsic features that are
444 time invariant. These problems again highlight the importance of using quantitative descriptors
445 of material compositions and structures as predictors in the QSPR.

446 Finally, the functional mechanisms of other influence factors such as relative humidity are
447 unclear, so they are not included in the QSPR. The effect of relative humidity on K_{ma} is likely
448 both chemical and material dependent^{4,9}, which will require more in-depth research.

449

450 4. Conclusions

451 A multiple linear regression model has been developed to predict the solid material-air partition
452 coefficients (K_{ma}) of organic compounds in various solid materials. Experimental K_{ma} data
453 collected from 43 studies were used to construct the regression model. The model uses three
454 continuous variables, chemical's $\log K_{oa}$, ΔH_v , and absolute temperature, as well as one
455 categorical variable, material type, as predictors. The model has been validated internally and

456 externally to be robust and stable, and have good predicting ability. The applicability domain of
457 the model, in terms of the range of predictors, includes chemical's logK_{oa} between 1.4 to 14.6,
458 ΔH_v from 22.3 to 75.6 kJ/mol, temperature between 15 and 100 °C, and material type belonging
459 to the 22 consolidated types.

460 The main advantage of the present model is that it is applicable for a wide range of chemical-
461 material-temperature combinations, which is more comprehensive than the correlation methods
462 developed in previous studies which were specific for one solid material and often at room
463 temperature. Moreover, a generic model is also developed which is able to give relatively
464 accurate estimates of K_{ma} without assigning a particular material type, making it suitable for
465 high-throughput assessments of the chemical releases from solid materials and subsequent
466 consumer exposures.

467

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470 American Chemistry Council and the US EPA (contract EP-16-C-000070).

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477 Tables and Figures

478 Table 1. Regression coefficients for Eq. 6

Variable	Coefficient	SE ^a	P-value
Intercept	-0.38	0.06	< 0.001
$\log_{10}K_{oa}$	0.63	0.01	< 0.001
$\Delta H_{ma}(1/T - 1/298.15)/2.303R$	0.96	0.04	< 0.001
<i>Consolidated material types (coefficient b)</i>			
Carpet	1.97	0.14	< 0.001
Cellulose fabric (cotton, linen)	0.72	0.12	< 0.001
Cement, Calcium silicate	1.11	0.10	< 0.001
Concrete	2.20	0.29	< 0.001
Ethylene Vinyl Acetate (EVA)	3.50	0.32	< 0.001
Glass	1.11	0.29	< 0.001
Gypsum board	1.28	0.18	< 0.001
Latex and solvent-based paint	2.92	0.19	< 0.001
Paper	0.14	0.10	0.16
Plywood	1.36	0.18	< 0.001
Polyester fabric	0.60	0.14	< 0.001
Polyether ether ketone (PEEK)	2.73	0.29	< 0.001
Polyethylene (PE)	2.45	0.17	< 0.001
Polypropylene (PP)	2.06	0.29	< 0.001
Polytetrafluoroethylene (PTFE)	2.08	0.29	< 0.001
PU-ester	-0.72	0.07	< 0.001
PU-ether^b	0.00	0.19	n/a
PUF-undefined	1.06	0.15	< 0.001
Rayon fabric	0.97	0.18	< 0.001
Stainless steel	2.07	0.29	< 0.001
Vinyl flooring	2.26	0.11	< 0.001
Wooden boards ^c	2.01	0.09	< 0.001

^aStandard error.

^bReference material.

^cIncludes oriented strand board (OSB), particleboard, medium-density board and high-density board.

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482 Table 2. External validation results

External validation criteria	R^2_{ext}	Q^2_{F1}	Q^2_{F2}	Q^2_{F3}	$\overline{r^2_m}$	CCC
Threshold		> 0.70	> 0.70	> 0.70	> 0.65	> 0.85
Splitting by random percentage	0.93	0.93	0.93	0.92	0.90	0.96
Splitting by ordered response	0.93	0.93	0.93	0.93	0.90	0.96
Splitting by ordered structure	0.94	0.94	0.94	0.94	0.91	0.97
Splitting by studies	0.79	0.86	0.78	0.86	0.71	0.89

R^2_{ext} : determination coefficient of the prediction set external data.

Q^2_{F1} : correlation coefficient proposed by Shi et al.

Q^2_{F2} : correlation coefficient proposed by Schuurmann et al.

Q^2_{F3} : correlation coefficient proposed by Consonni et al.

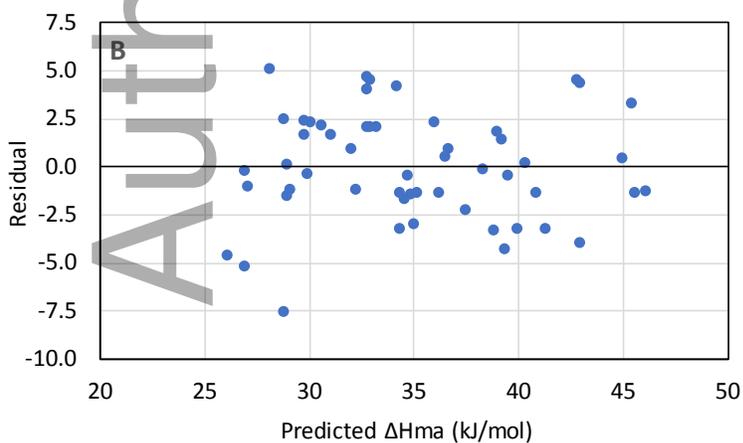
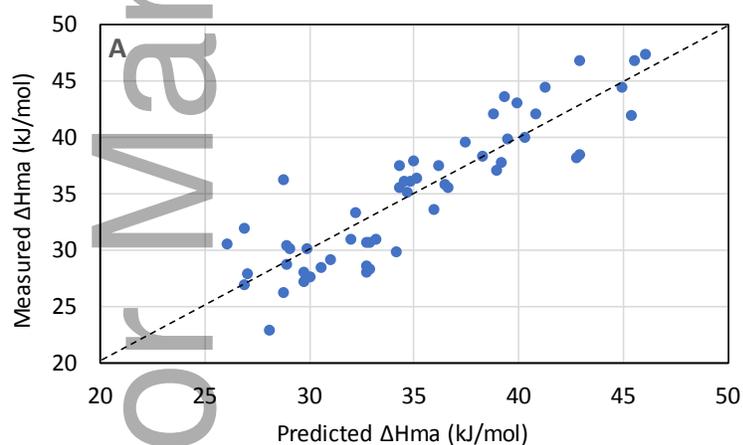
$\overline{r^2_m}$: determination coefficient proposed by Ojha et al.

CCC: concordance correlation coefficient proposed by Chirico and Gramatica.

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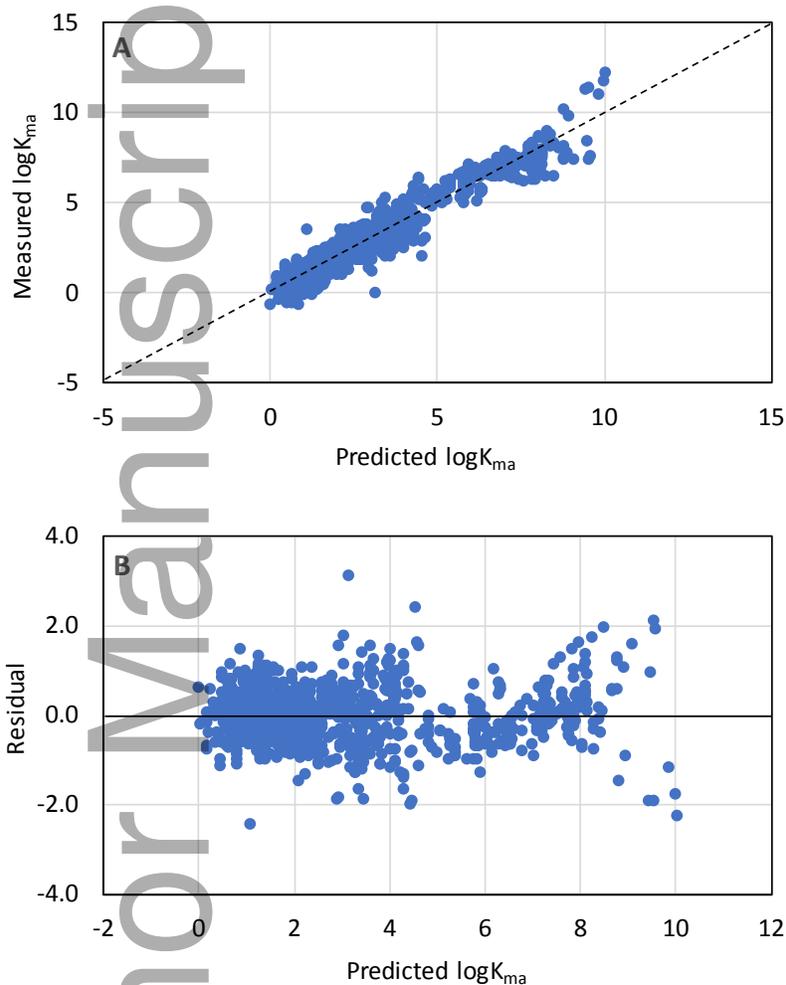
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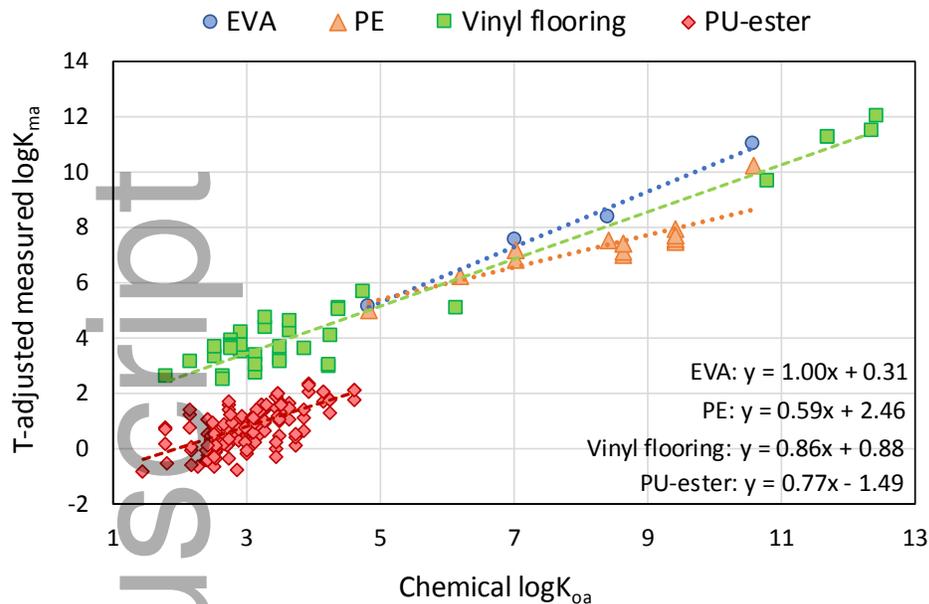


486

487 Figure 1. (A) Measured Enthalpy of material-air partitioning (ΔH_{ma}) and (B) residuals as a
488 function of the (ΔH_{ma}) predicted from chemical enthalpy of vaporization (ΔH_v - Eq. 5). The
489 dotted line in (A) indicates the 1:1 line.
490

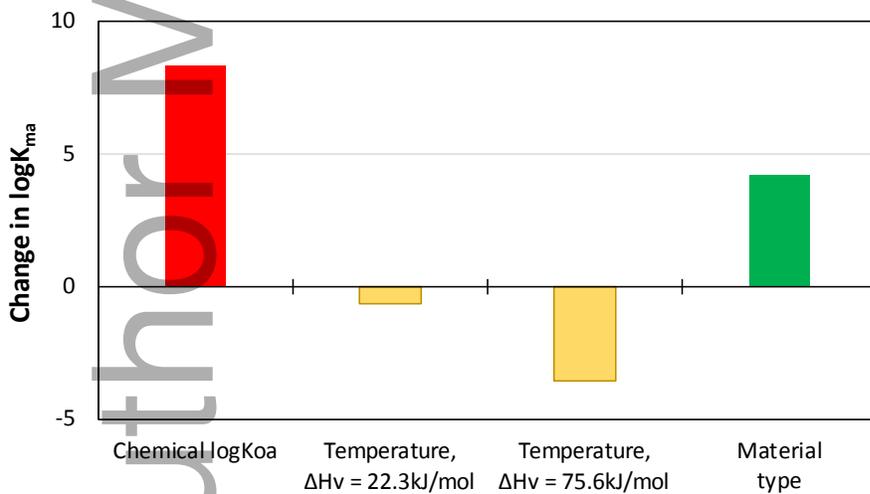


491
492 Figure 2. (A) measured $\log K_{ma}$ and (B) residuals as a function of $\log K_{ma}$ predicted by the final
493 QSPR (Eq. 6). The dotted line in (A) indicates the 1:1 line.
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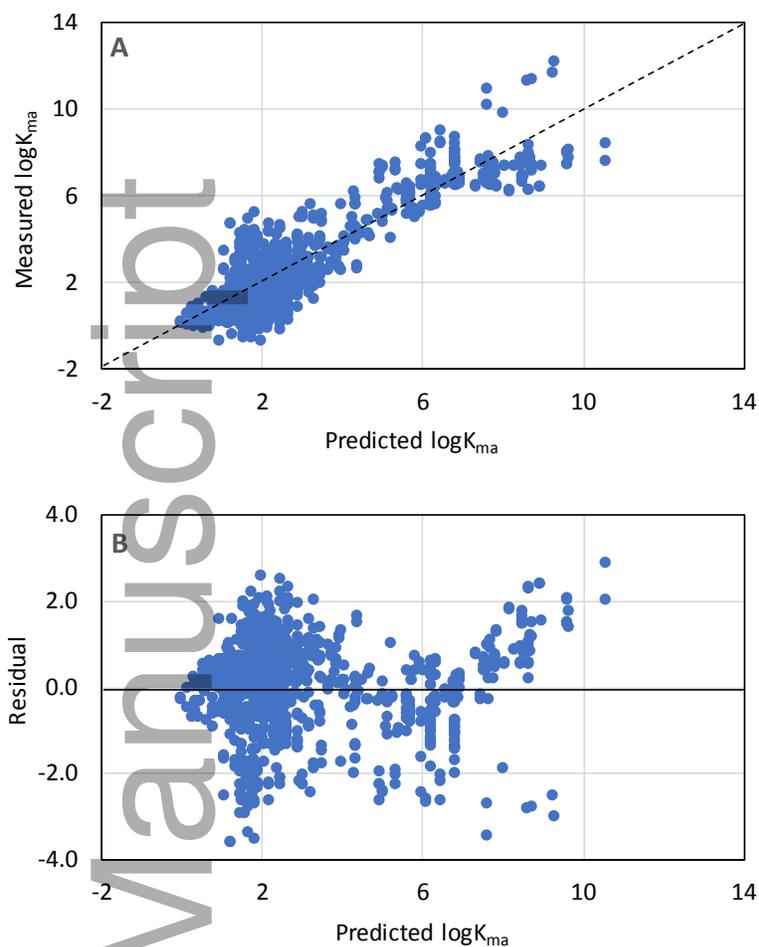
496
 497 Figure 3. temperature adjusted measured $\log K_{ma}$ as a function of $\log K_{oa}$ for selected material
 498 types including EVA, PE, vinyl flooring, and PU-ester.

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 502 Figure 4. Change in $\log K_{ma}$ with respect to the change in each predictor, from minimum to
 503 maximum values within the entire dataset.

504



505
506 Figure 5. (A) measured $\log K_{ma}$ and (B) residuals as a function of $\log K_{ma}$ predicted by the generic
507 QSPR (Eq. 7). The dotted line in (A) indicates the 1:1 line.

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514 References

- 515 1. Huang L, Fantke P, Ernstoff A, et al. A Quantitative Property-Property Relationship for
516 the Internal Diffusion Coefficients of Organic Compounds in Solid Materials. *Indoor Air*.
517 2017.

- 518 2. Huang L and Jolliet O. A parsimonious model for the release of chemicals encapsulated
519 in products. *Atmos Environ.* 2016; 127: p. 223-235.
- 520 3. Liu Z, Ye W, Little JC. Predicting emissions of volatile and semivolatile organic
521 compounds from building materials: a review. *Build Environ.* 2013; 64: p. 7-25.
- 522 4. Kamprad I and Goss K-U. Systematic investigation of the sorption properties of
523 polyurethane foams for organic vapors. *Analytical chemistry.* 2007; 79(11): p. 4222-4227.
- 524 5. Bartkow ME, Hawker DW, Kennedy KE, et al. Characterizing uptake kinetics of PAHs
525 from the air using polyethylene-based passive air samplers of multiple surface area-to-
526 volume ratios. *Environ Sci Technol.* 2004; 38(9): p. 2701-2706.
- 527 6. Kennedy KE, Hawker DW, Müller JF, et al. A field comparison of ethylene vinyl acetate
528 and low-density polyethylene thin films for equilibrium phase passive air sampling of
529 polycyclic aromatic hydrocarbons. *Atmos Environ.* 2007; 41(27): p. 5778-5787.
- 530 7. Morrison G, Li H, Mishra S, et al. Airborne phthalate partitioning to cotton clothing.
531 *Atmos Environ.* 2015; 115: p. 149-152.
- 532 8. Morrison G, Shakila N, Parker K. Accumulation of gas-phase methamphetamine on
533 clothing, toy fabrics, and skin oil. *Indoor air.* 2015; 25(4): p. 405-414.
- 534 9. Zhao D, Little JC, Cox SS. Characterizing polyurethane foam as a sink for or source of
535 volatile organic compounds in indoor air. *Journal of environmental engineering.* 2004;
536 130(9): p. 983-989.
- 537 10. Bodalal A, Zhang J, Plett E, et al. Correlations between the internal diffusion and
538 equilibrium partition coefficients of volatile organic compounds (VOCs) in building
539 materials and the VOC properties. *ASHRAE Trans.* 2001; 107: p. 789.
- 540 11. Guo Z. Review of indoor emission source models. Part 2. Parameter estimation. *Environ*
541 *Pollut.* 2002; 120(3): p. 551-564.
- 542 12. Chaemfa C, Barber JL, Gocht T, et al. Field calibration of polyurethane foam (PUF) disk
543 passive air samplers for PCBs and OC pesticides. *Environ Pollut.* 2008; 156(3): p. 1290-
544 1297.
- 545 13. Shoeib M and Harner T. Characterization and comparison of three passive air samplers
546 for persistent organic pollutants. *Environ Sci Technol.* 2002; 36(19): p. 4142-4151.

- 547 14. Holmgren T, Persson L, Andersson PL, et al. A generic emission model to predict release
548 of organic substances from materials in consumer goods. *Science of the Total*
549 *Environment*. 2012; 437: p. 306-314.
- 550 15. Booij K, Hofmans HE, Fischer CV, et al. Temperature-dependent uptake rates of
551 nonpolar organic compounds by semipermeable membrane devices and low-density
552 polyethylene membranes. *Environ Sci Technol*. 2003; 37(2): p. 361-366.
- 553 16. Adams RG, Lohmann R, Fernandez LA, et al. Polyethylene devices: Passive samplers for
554 measuring dissolved hydrophobic organic compounds in aquatic environments. *Environ*
555 *Sci Technol*. 2007; 41(4): p. 1317-1323.
- 556 17. Liu Y, Zhou X, Wang D, et al. A prediction model of VOC partition coefficient in porous
557 building materials based on adsorption potential theory. *Build Environ*. 2015; 93: p. 221-
558 233.
- 559 18. Zhang Y, Luo X, Wang X, et al. Influence of temperature on formaldehyde emission
560 parameters of dry building materials. *Atmos Environ*. 2007; 41(15): p. 3203-3216.
- 561 19. USEPA, *Estimation Programs Interface Suite™ for Microsoft® Windows, v 4.11*, 2012,
562 United States Environmental Protection Agency: Washington, DC.
- 563 20. Gramatica P, Cassani S, Chirico N. QSARINS-chem: Insubria datasets and new
564 QSAR/QSPR models for environmental pollutants in QSARINS. *J Comp Chem*. 2014;
565 35(13): p. 1036-1044.
- 566 21. Gramatica P, Chirico N, Papa E, et al. QSARINS: A new software for the development,
567 analysis, and validation of QSAR MLR models. *J Comp Chem*. 2013; 34(24): p. 2121-
568 2132.
- 569 22. Chirico N and Gramatica P. Real external predictivity of QSAR models. Part 2. New
570 intercomparable thresholds for different validation criteria and the need for scatter plot
571 inspection. *J Chem Inf Model*. 2012; 52(8): p. 2044-2058.
- 572 23. Chirico N and Gramatica P. Real external predictivity of QSAR models: how to evaluate
573 it? Comparison of different validation criteria and proposal of using the concordance
574 correlation coefficient. *J Chem Inf Model*. 2011; 51(9): p. 2320-2335.
- 575 24. Gramatica P. Principles of QSAR models validation: internal and external. *QSAR &*
576 *combinatorial science*. 2007; 26(5): p. 694-701.

577 25. Cassani S and Gramatica P. Identification of potential PBT behavior of personal care
578 products by structural approaches. *Sustain Chem Pharm.* 2015; 1: p. 19-27.

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