ORIGINAL RESEARCH

Interspecific variation and elevated CO₂ influence the relationship between plant chemical resistance and regrowth tolerance

Leslie E. Decker¹ Mark D. Hunter²

Correspondence

Leslie E. Decker, Department of Biology, Stanford University, 327 Campus Drive, Stanford, CA 48109-1085, USA. Email: lesldeck@stanford.edu

Funding information

Division of Environmental Biology, Grant/ Award Number: 1256115 and 1257160

Abstract

To understand how comprehensive plant defense phenotypes will respond to global change, we investigated the legacy effects of elevated CO2 on the relationships between chemical resistance (constitutive and induced via mechanical damage) and regrowth tolerance in four milkweed species (Asclepias). We quantified potential resistance and tolerance trade-offs at the physiological level following simulated mowing, which are relevant to milkweed ecology and conservation. We examined the legacy effects of elevated CO2 on four hypothesized trade-offs between the following: (a) plant growth rate and constitutive chemical resistance (foliar cardenolide concentrations), (b) plant growth rate and mechanically induced chemical resistance, (c) constitutive resistance and regrowth tolerance, and (d) regrowth tolerance and mechanically induced resistance. We observed support for one trade-off between plant regrowth tolerance and mechanically induced resistance traits that was, surprisingly, independent of CO2 exposure. Across milkweed species, mechanically induced resistance increased by 28% in those plants previously exposed to elevated CO2 In contrast, constitutive resistance and the diversity of mechanically induced chemical resistance traits declined in response to elevated CO2 in two out of four milkweed species. Finally, previous exposure to elevated CO₂ uncoupled the positive relationship between plant growth rate and regrowth tolerance following damage. Our data highlight the complex and dynamic nature of plant defense phenotypes under environmental change and question the generality of physiologically based defense trade-offs.

KEYWORDS

Asclepias, cardenolides, global change ecology, plant-herbivore interactions, resistance to herbivory

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2020 The Authors. Ecology and Evolution published by John Wiley & Sons Ltd.

¹Department of Biology, Stanford University, Stanford, CA, USA

²Department of Ecology and Evolutionary Biology, University of Michigan, Biological Sciences Building, Ann Arbor, MI, USA

1 | INTRODUCTION

Plants employ a suite of defensive traits to deter and minimize the impacts of herbivory (Agrawal & Fishbein, 2006). As a result, critical factors that contribute to plant fitness in the context of damage may be overlooked when defenses are studied in isolation (Baucom & De Roode, 2011). Regrowth tolerance and chemical resistance are two major strategies of defense that plants employ against herbivory and that jointly influence fitness (Agrawal, 2011; Fornoni, 2011; Núñez-Farfán, Fornoni, & Valverde, 2007; Stamp, 2003; Strauss & Agrawal, 1999; Zas, Moreira, & Sampedro, 2011). One major form of defense, resistance to herbivory, occurs through physical and chemical traits such as trichomes, latex exudation, thorns, and toxic secondary metabolites that together reduce herbivore performance (Rhoades, 1985). Multiple resistance traits can be both constitutively expressed before damage and induced following damage: however, chemical resistance is perhaps best known for this temporal strategy (Agrawal & Karban, 1999; Karban & Baldwin, 2013). Because resistance traits are themselves metabolically costly (Bekaert, Edger, Hudson, Pires, & Conant, 2012; Gershenzon, 19 94; Strauss, Rudgers, Lau, & Irwin, 2002), it is thought that these costs manifest in the form of trade-offs with other plant functions like growth rates (Agrawal, 2011; Fineblum, Rausher, & D., 1995; van der Meijden, Wijn, & Verkaar, 1988; Stamp, 2003). Plant tolerance to herbivory, or compensatory growth following damage, minimizes fitness losses through simultaneous shifts in physiology and resource allocation (Fornoni, Ez-Farfán, & J, Valverde PL, 2003; Rosenthal & Kotanen, 1994; Strauss & Agrawal, 1999).

Numerous hypotheses have been developed to predict resource allocation to the competing plant functions of chemical resistance and regrowth tolerance, often with variable generality and empirical support (reviewed in Stamp, 2003). On a macroevolutionary level, the resource availability hypothesis (RAH) predicts that high-resource environments select for fast-growing species that invest in regrowth tolerance following damage rather than chemical defense production (Coley & Chapin, 1985; Endara & Coley, 2011). But fewer hypotheses address the interplay of tolerance and chemical defense within populations (Hahn, Agrawal, Sussman, & Maron, 2019; Hahn & Maron, 2016) or physiologically within the lifetime of individuals. At the cellular and tissue level, the growth-differentiation balance hypothesis (GDB) posits that plants in high-resource environments will not be limited by photosynthesis and will allocate more energy into regrowth rather than into cellular differentiation-related processes such as secondary metabolism (Herms & Mattson, 1992). At intermediate resource conditions, the GDB predicts that plants will be limited in growth but not photosynthetic capacity and will produce more secondary metabolites relatively cheaply. Due to the need to test at multiple resource levels, and measure not only growth rate, but net assimilation and secondary metabolism, the GDB has proven difficult to test but still provides a useful framework of plant defense at the physiological level (Stamp, 2004). In general, trade-offs between tolerance and chemical resistance arise as a result of plant allocation strategies meant to optimize fitness in a variable environment (Züst & Agrawal, 2017). Therefore, understanding the environmental conditions under which trade-offs manifest is of critical importance.

The rapidly rising concentration of atmospheric carbon dioxide is well-known to influence chemical resistance to herbivores and plant growth rates. Both the composition and concentration of constitutively expressed and induced plant secondary metabolites change in response to elevated CO₂ depending on the class of compounds considered (Bidart-Bouzat, Mithen, & Berenbaum, 2005; Hunter, 2001; Jia, Zhao, Liu, & Huang, 2016; Klaiber, Dorn, & Najar-Rodriguez, 2013; Robinson, Ryan, & Newman, 2012; Ryan, Rasmussen, & Newman, 2010; Zavala, Nabity, & DeLucia, 2013). Further, elevated CO₂ suppresses the synthesis of jasmonic acid and stimulates the production of salicylic acid, compromising the plant's ability to mount an induced resistance response (Ode, Johnson, & Moore, 2014).

Changes in phytohormonal signaling pathways also mediate plant growth and regrowth tolerance following damage under elevated CO_2 (Guo et al., 2012). In general, by increasing photosynthesis and water use efficiency, elevated CO_2 positively affects plant growth rates (Drake, Gonzalez-Meler, & Long, 1997; Ainsworth & Long, 2005; Robinson et al., 2012; Bazzaz, Ackerly, Woodward, & Rochefort, 2013). However, the direct effects of elevated CO_2 on plant regrowth tolerance following damage can be negative (Guo et al., 2012;Lau & Tiffin, 2009;Marshall, Avila-Sakar, & Reekie, 2008;Wilsey, 2001) partially because of increased nutrient limitation under elevated CO_2 paired with phytohormonal suppression. Studies that explore the integrated influence of elevated CO_2 on the relationships between resistance and tolerance are sorely lacking

Even less is known about the lingering effects of past CO2 enrichment on plants. Though not ecologically plausible, the modulation of exposure to environmental change drivers such as elevated CO₂ partially reveals energetic allocation decisions made by plants under future conditions, and the persistence of those responses. Extrapolations based on the substantial below-ground carbon sink and increased soil microbial turnover that develops in response to elevated CO₂ predict mixed but lingering effects of elevated CO₂ on plant regrowth tolerance (Hungate, Johnson, & Dijkstra, 2006; Stiling, Moon, & Rossi, 2013). To our knowledge, only two studies have examined plant responses to elevated CO2 beyond the cessation of enrichment and found lasting effects on aspects of root morphology such as fine root hairs (Stiling et al., 2013) and increases in regrowth tolerance following fire (Bain & Day, 2019). These studies follow plant and arthropod communities in the years following enrichment cessation, yet how plant physiological properties will respond to abrupt changes in CO2 enrichment over the course of a growing season remains to be tested.

Here, we investigate the legacy effects of elevated CO_2 on the chemical resistance traits and regrowth tolerance of four milkweed species (Asclepias). Specifically, we examined the effects of elevated CO_2 on four hypothesized trade-offs between the following: (a) initial growth rate and constitutive chemical resistance, (b) initial

growth rate and mechanically induced chemical resistance, (c) constitutive chemical resistance and regrowth tolerance following damage, and (d) regrowth tolerance and mechanically induced chemical resistance. To our knowledge, no theory exists to predict the interaction between resistance and regrowth tolerance strategies under changing carbon supplementation. Nevertheless, we predicted that elevated CO₂ would induce higher growth rates and regrowth rates and depress constitutive secondary metabolites following the GDB hypothesis and mitigate, in part, any trade-off between chemical resistance traits and regrowth tolerance in milkweed. By analyzing changes in plant tolerance and resistance chemistry, we aimed to improve our understanding of how future environmental conditions may influence the defensive phenotype of plants, with implications for the herbivore communities that damage them.

2 | MATERIALS AND METHODS

2.1 | Study system

The four milkweed, Asclepias, species used in our study (A. syriaca, A. speciosa, A. incarnata, and A. curassavica) originate from North and Central America (Woodson, 1954) and support herbivores that range from phloem-feeding insects such as oleander aphids (Aphis nerii) to chewing insects capable of removing large amounts of tissue, like monarch caterpillars (Danaus plexippus), and long horn beetles (Tetraopes spp.). Most milkweed herbivores specialize within the genus because Asclepias produce a well-characterized suite of defenses against herbivory.

To physically deter feeding by arthropod herbivores, milkweed plants exude latex, produce trichomes, and increase leaf toughness (Agrawal & Fishbein, 2006; Agrawal & Konno, 2009; Hochwender, Marquis, & Stowe, 2000; Zalucki, Brower, & Alonso-M, 2001). However, milkweeds are best known for synthesizing a class of toxic steroids known as cardenolides that disrupt Na⁺/K⁺-ATPase in the Na⁺/K⁺-channels of animal cells (Agrawal, Petschenka, Bingham, Weber, & Rasmann, 2012). The composition and concentration of cardenolides produced constitutively by milkweed plants vary substantially within and among milkweed species (Agrawal et al., 2012; Rasmann & Agrawal, 2011). Damage induces quick increases in cardenolide concentrations and changes in cardenolide composition (Malcolm & Zalucki, 1996). Regrowth following damage also plays a prominent role in the defensive phenotype of milkweeds (Agrawal & Fishbein, 2008; Tao, Ahmad, Roode, & Hunter, 2016). Despite a growing body of work illustrating the effects of environmental change on milkweed chemistry and milkweed growth (Matiella, 2012; Tao, Berns, & Hunter, 2014; Vannette & Hunter, 2011), no study to date has explored the interplay between milkweed chemical resistance traits (both constitutive and induced) and regrowth tolerance under future environmental conditions.

We grew four species of milkweed under ambient (400 ppm) and elevated (760 ppm) concentrations of atmospheric CO_2 at the University of Michigan Biological Station (UMBS). To manipulate

atmospheric CO_2 concentrations, we used an outdoor array consisting of 40 open-top chambers, with 20 chambers maintained at ambient CO_2 , and 20 chambers maintained at elevated CO_2 from May through August of 2015. Chambers were 1 m high cubes with an octagonal top of diameter of 0.8 m composed of a PVC frame and clear plastic walls following a modified design of Drake, Leadley, Arp, Nassiry, and Curtis (1989).

We chose Asclepias species that vary in foliar cardenolide concentrations. Specifically, we included A. incarnata (low cardenolide), A. speciosa, A. syriaca (both medium cardenolide), and A. curassavica (high cardenolide). Seeds of A. speciosa and A. curassavica were obtained from commercial sources (Prairie Moon Nurseries, Winona, USA), and seeds of A. incarnata and A. syriaca were collected locally (Cheboygan county, MI). We surface-sterilized all seeds following a six-week cold stratification period (for all but tropical A. curassavica) and germinated seeds on moist filter paper for 1 week. We planted seedlings in 983 cm³ DeepotsTM (6.9 cm diameter by 35.6 cm height) containing Metromix 360 (Sun Gro Horticulture, Vancouver, BC) and Osmocote Controlled Release Fertilizer [N:P:K:16:16:16 ppm N (g/g)] (ICL Specialty Fertilizers, Dublin, USA) on May 5, 15. Germinated seedlings were watered daily and grown in the UMBS greenhouse for two weeks before they were moved to randomly assigned chambers in the CO2 array. Once in the array, potted plants were maintained under their CO₂ treatments for three months. To minimize the entrance of herbivores into the chambers, we placed fine mesh coverings over the openings of each chamber and physically removed any herbivores that we observed during daily visual inspections.

Within each chamber, we grew as many as seven plants of each milkweed species. Low germination success limited the number of A. speciosa and A. syriaca used in this study, and not all milkweed species were represented in every chamber. Overall, our eight treatments ($2~\mathrm{CO}_2$ treatments \times 4 milkweed species) combined for a total of 442 plants, with exact replicate numbers reported in Table 1.

Using a LI-COR 320 IRGA (LI-COR, Lincoln, USA), we monitored atmospheric CO_2 concentrations daily in the 20 elevated CO_2 chambers and in one randomly selected ambient CO_2 chamber. Concentrations of CO_2 were adjusted throughout the day to maintain the target of 760 ppm in each elevated chamber. The ambient temperature inside each chamber was recorded every hour using a thermochron datalogger (Thermochron, Baulkham Hills, Australia). Elevated CO_2 chambers averaged 21.03 (±0.034) °C, and ambient CO_2 chambers averaged 21.24 (±0.038) °C, roughly 2°C higher than the outside average temperature of 18.93 (±0.039) °C.

2.2 | Simulated damage and growth measures

Three months following the initial transfer of plants into the array, we simulated clipping/mowing by cutting all plants at the soil line. Many milkweed habitats important to the specialist herbivores associated with milkweed are located near roadways and agricultural fields that are regularly mowed. Properly timed mowing can improve reproduction and decrease predator abundance of certain milkweed

TABLE 1 Sample sizes of 442 milkweed plants grown under either ambient (400 ppm) or elevated (760 ppm) CO₂ grouped (a) by species and (b) by their distribution in 40 open-top chambers. Species codes are as follows: CUR = A. curassavica, SYR = A. syriaca, SPE = A. speciosa, INC = A. incarnata

CO ₂ treatment					specie	s					N
Ambient					A. cura	ssavica					84
					A. inca	rnata					105
					A. spec	iosa					22
					A. syric	іса					25
Elevated					A. cura	ssavica					81
					A. inca	rnata					91
					A. spec	iosa					23
					A. syric	іса					11
CO ₂ Treatment	Chamber	CUR	INC	SPE	SYR	Chamber	CUR	INC	SPE	SYR	
Elevated	1	4	3	0	2	21	4	4	2	1	
Ambient	2	6	6	1	2	22	5	6	2	0	
Elevated	3	6	7	1	0	23	2	5	3	0	
Ambient	4	3	6	2	3	24	5	5	0	1	
Elevated	5	4	4	3	1	25	6	6	2	1	
Ambient	6	4	4	0	1	26	4	6	1	0	
Elevated	7	5	6	2	0	27	2	4	0	1	
Ambient	8	4	6	1	1	28	4	6	0	1	
Elevated	9	1	6	0	0	29	5	5	2	0	
Ambient	10	4	5	3	1	30	3	5	0	2	
Elevated	11	4	5	1	0	31	5	2	0	1	
Ambient	12	4	6	2	1	32	4	3	0	0	
Elevated	13	3	1	1	0	33	4	6	3	0	
Ambient	14	2	5	2	3	34	4	5	0	1	
Elevated	15	6	3	1	1	35	2	3	1	2	
Ambient	16	5	5	0	3	36	6	6	3	2	
Elevated	17	3	4	1	0	37	4	6	1	1	
Ambient	18	5	4	0	0	38	5	6	2	1	
Elevated	19	5	6	0	0	39	6	5	0	0	
Ambient	20	5	6	2	1	40	2	4	0	1	

specialists, including the monarch butterfly (Haan & Landis, 2019). Thus, our simulated mowing represents an ecologically relevant stress regularly experienced by many milkweed plants. Moreover, at our field site in northern Michigan, we have observed chipmunks (Tamius striatus), milkweed stem weevils (Rhyssomatus lineaticollis), and porcupines (Erethizon dorsatum) all remove the entire aboveground tissues of milkweed plants. Other herbivores such as monarch caterpillars, and milkweed tussock moths (Euchaetes egle), have also been observed to remove large amounts of foliage. Thus, our clipping treatment also represents severe but not infrequent levels of herbivore damage experienced by milkweed plants. We recognize that mechanical damage does not completely mimic actual herbivory because oral secretions and regurgitant released from the herbivore at the time of feeding can enter wounded plant tissue, inducing the release of jasmonic acid, a phytohormone critical to the production of defensive secondary metabolites (McCloud & Baldwin, 1997).

The aboveground biomass that we removed was dried at 60° C, weighed, and used to calculate growth rate prior to damage (below). Cut plants were watered, moved to the UMBS greenhouse, and maintained under identical (ambient CO_2) conditions for three weeks due to external limitations on use of the chambers. However, by re-growing clipped plants under ambient CO_2 we are able to isolate the legacy effects of altered carbon availability prior to damage on regrowth tolerance, and potential trade-offs between growth and resistance. Thus, we can examine the repercussions of previous energetic allocation decisions made by plants under carbon-enriched conditions in comparison to those under ambient conditions. After a three-week period, the aboveground regrowth plant material was harvested, dried at 60° C, and weighed as a measure of regrowth tolerance.

For a measure of growth rate prior to damage, we divided the aboveground dry biomass of the plant by 64 days (the number of

days since the seedling had been transferred to soil) following Agrawal and Fishbein (2008). Similarly, to calculate plant regrowth rate following mechanical damage, we divided the mass of the regrowth material by 21 days (the length of time plants were allowed to regrow following damage). Differences in regrowth rate following damage are important for the competitive success and ultimate fitness of plants (Züst & Agrawal, 2017).

2.3 | Chemical analyses and resistance classifications

We collected samples of the original aboveground foliage and the regrowth foliage of each plant for cardenolide analysis using established methods (Tao & Hunter, 2012; Vannette & Hunter, 2011; Zehnder & Hunter, 2009). Roughly, 20 mg of dried plant material was ground in a ball mill, deposited in 1 ml methanol, and stored at -10°C prior to analysis. Cardenolides were extracted, separated, and quantified with a 0.15mg/ml digitoxin internal standard (Sigma Chemical Company), by reverse-phase high-performance liquid chromatography (HPLC) on a Waters Acquity UPLC with PDA detector (Waters Corporation, Milford). Peaks with symmetrical absorbance between 217-222 nm were identified as cardenolides. Cardenolide concentrations were calculated as the sums of all separated peak areas, corrected by the concentration of the internal digitoxin standard and sample dry mass. We used digitoxin as an internal standard because it is absent from Asclepias and because purified standards remain unavailable for a majority of milkweed cardenolides. We recognize that cardenolides may differ in their concentration-area relationships, and our estimates of cardenolide concentration should be considered as measured in digitoxin-equivalents. Because milkweed plants were grown in field mesocosms which excluded herbivores all season, the foliar cardenolides measured from plants prior to simulated damage represent natural levels of constitutive resistance. Conversely, the foliar cardenolide concentrations of regrown tissue following clipping represent mechanically induced resistance.

2.4 | Statistical analyses

In all analyses that follow, we used either linear mixed models (LMMs; Lme4 package) or generalized linear mixed models (GLMMs; Lme4 package). To account for variation among chambers and the nonindependence of plants grown within the same individual chamber, we included chamber identity as a random effect in all of our models described below. This design allows us to test our hypotheses at the level of plant individuals to capture relevant variation in our analyses, while accounting for multiple plants within chambers. We performed all statistical tests in R version 3.6.0 (R Core Team, 2018) and selected models using likelihood ratio tests (Burnham & Anderson, 2002). Variables were

transformed to best achieve normality of error as tested by the Shapiro-Wilk normality test. Homogeneity of variance and distribution of residuals were inspected using quantile-quantile and residuals fitted value plots to check for conformation to model assumptions (Crawley, 2012).

2.5 | Testing for trade-offs among milkweed growth, regrowth tolerance, and resistance chemistry

2.5.1 | Plant growth rate and chemical resistance before damage

We used an LMM with log-transformed initial foliar cardenolide concentrations as the dependent variable and square root-transformed growth rate prior to clipping, CO_2 treatment, and milkweed species as fixed effects. An interaction between growth rate prior to clipping and CO_2 indicates a difference between the CO_2 treatments in the extent to which growth rate correlates with the production of cardenolides.

2.5.2 | Plant growth rate before damage and mechanically induced resistance of regrowth tissues

We used an LMM with log-transformed foliar cardenolide concentrations of the regrowth foliage as the dependent variable and square root-transformed growth rate prior to clipping, CO_2 treatment, and milkweed species as fixed effects. An interaction between initial growth rate and CO_2 indicates a difference between CO_2 treatments in the potential trade-off between plant growth rate before damage and chemical resistance after damage.

2.5.3 | Chemical resistance before damage and regrowth tolerance

Likewise, we ran an LMM with square root-transformed regrowth rate as the response variable and log-transformed initial foliar cardenolide concentrations, CO_2 treatment, and milkweed species as fixed effects. An interaction between initial foliar cardenolide concentration and CO_2 indicates a difference between atmospheres in the relationship between initial plant chemical resistance and regrowth.

2.5.4 | Regrowth tolerance and the mechanically induced resistance of regrowth tissues

Lastly, we ran an LMM with log-transformed regrowth foliar cardenolide concentrations as the response variable and square roottransformed regrowth rate, CO₂ treatment, and milkweed species as fixed effects. A significant interaction between CO₂ treatment and regrowth rate would signify a difference between the two atmospheres in any correlation between the two defense traits.

2.6 | Elevated CO₂, milkweed species, and plant growth and resistance profiles

While the trade-off model framework described above provided some information on how growth rates and chemical resistance responded to our treatments, we also performed the following additional analyses to ask further questions about defense phenotypes. To determine the effects of our treatments on plant growth rate prior to damage and regrowth rate after damage, we used CO_2 treatment, the probability of regrowth, and milkweed species as fixed effects and square root-transformed growth rates (mg/day) as response variables. Not all milkweed individuals regrew following damage. We therefore used generalized linear mixed models with binomial error distributions and logit link functions to assess the effects of plant species and CO_2 treatment on the proportion of milkweed plants that regrew following damage.

We then examined how CO_2 treatment and species influenced the relationship between growth rate prior to damage and regrowth rate following damage, using an LMM with square root-transformed regrowth rate as the response variable and square root-transformed initial growth rate, CO_2 treatment, and species as fixed effects.

Plant chemical defense encompasses not only the total concentration of defense compounds but also the diversity of chemical species produced. We therefore examined the relationships between cardenolide community diversity and growth rates. We calculated cardenolide diversity using the Shannon diversity index borrowed from the biodiversity literature: $H = -\text{sum}(P_i \log [P_i])$ where P_i is the relative amount of a cardenolide peak compared to the total amount of cardenolides in an individual plant (Rasmann & Agrawal, 2011). Similar to above, we selected simplified models from two starting LMMs: (a) with constitutive foliar cardenolide concentrations as the dependent variable and square root-transformed growth rate prior to clipping, CO_2 treatment, and milkweed species as fixed effects; and (b) with mechanically induced foliar cardenolide concentrations as the dependent variable and square root-transformed regrowth rate, CO_2 treatment, and milkweed species as fixed effects.

To compare the effects of CO_2 treatment, and milkweed species on the community of cardenolide compounds produced in the plants before and after damage, we used permutational multivariate analysis of variance (PerMANOVA; Anderson, 2001). The model included CO_2 treatment, milkweed species, tissue type, and their interactions as fixed effects, and Bray–Curtis distance of percentage weight of each foliar cardenolide peak as dependent variables. To visualize these differences, we used nonmetric multidimensional scaling (NMDS) with 999 permutations per model run and a maximum of 500 runs per dimension (model stress = 0.200). PerMANOVA and NMDS scaling were performed using the VEGAN package in R (Oksanen & Friendly, 2017).

3 | RESULTS

3.1 | Only regrowth tolerance and induced resistance traded off among individuals following mechanical damage

3.1.1 | Plant growth rate and constitutive resistance

Milkweed growth rate prior to damage was unrelated to foliar constitutive cardenolide concentrations prior to damage (initial growth rate: $F_{1.195} = 2.72$, p = .100, Figure 1a; Table 2). Elevated CO_2 had no effect on this nonsignificant relationship (CO_2^* initial growth rate: $F_{1.195} = 0.46$, p = .499).

3.1.2 | Plant growth rate before damage and mechanically induced resistance

Instead of a trade-off between growth rate prior to damage and the mechanically induced chemical resistance of regrown tissues following damage, we found a positive relationship that weakened (became less steep) under elevated CO_2 (CO_2^* initial growth rate: $\mathrm{F}_{1,215} = 5.33$, p = .022, Figure 1b; Table 2).

3.1.3 | Constitutive resistance before damage and regrowth tolerance after damage

Similarly, we observed a weak positive relationship between constitutive chemical resistance and regrowth tolerance (constitutive resistance: $F_{1,208} = 3.66$, p = .057, Figure 1c; Table 2). Model selection eliminated models containing the influence of CO_2 on this relationship.

3.1.4 | Regrowth tolerance and mechanically induced resistance of regrown tissues

In contrast to the first three potential trade-offs, we observed a significant trade-off between regrowth tolerance and the mechanically induced chemical resistance of regrown foliage (Regrowth rate*milkweed species: $F_{1,215} = 7.18$, p = .0001, Figure 1d; Table 2). The trade-off was determined by two of the four milkweed species (A. *incarnata* and A. *speciosa*). As above, our selection process eliminated models containing the influence of CO_2 on this relationship.

3.2 | Elevated CO₂ eliminated the positive relationship between initial growth rate and regrowth tolerance following damage

Across all milkweed species, elevated CO_2 induced an average 24% increase in growth rate (CO_2 : $F_{1,151}$ = 9.71, p = .002, Figure 2a)

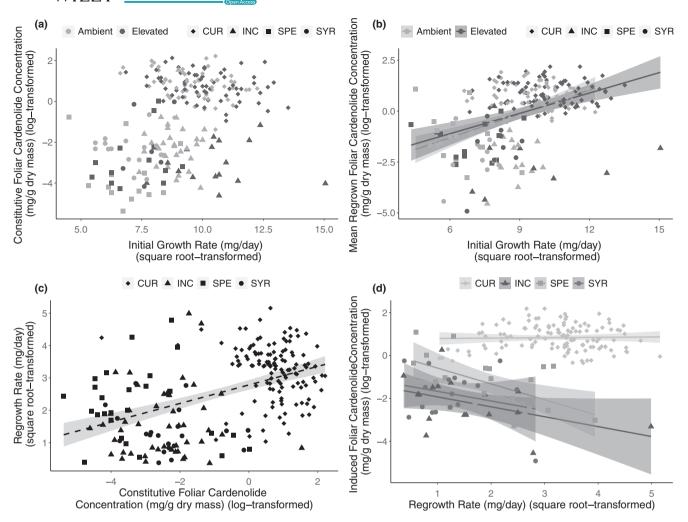


FIGURE 1 Support for only one of the four hypothetical trade-offs relating the growth rate of milkweeds before and after damage to their constitutive and mechanically induced foliar cardenolide concentrations before and after damage. Milkweeds were grown under either elevated (760 ppm) or ambient (400ppm) atmospheric concentrations of CO₂. (a) Nonsignificant effects of CO₂ treatment and predamage growth rate on milkweed constitutive cardenolide concentrations (mg/g dry mass) before damage. (b) Positive effects of CO₂ treatment and predamage growth rate on mechanically induced cardenolide concentrations in regrown leaves (mg/g dry mass) after damage. (c) Slight effects of milkweed constitutive cardenolide concentrations before damage on regrowth rate after damage contingent upon milkweed species. (d) Significant trade-off between mechanically induced cardenolide concentrations (mg/g dry mass) in regrowth leaves and the regrowth rate (mg/day) of milkweeds. Regressions are represented with 95% confidence intervals and milkweed species codes are as follows: CUR = A. curassavica (diamond), INC = A. incarnata (square), SPE = A. speciosa (triangle), and SYR = A. syriaca (circle). In figures (a-c), light gray shapes represent plants grown under ambient CO₂ and dark gray shapes are those grown under elevated CO₂. In figure (d) shading corresponds to milkweed species

illustrating the classic effect of CO_2 fertilization on plant growth (Kimball, 1983;Leadley, Niklaus, Stocker, & Körner, 1999). Initial growth rates of milkweed increased most strongly in A. syriaca (43%) followed by A. incarnata (31%), A. curassavica (12%), and A. speciosa (7%) (species* CO_2 : $F_{3,409} = 3.24$, p = .022, Figure 2a). Surprisingly, previous CO_2 exposure had no effect on regrowth tolerance across milkweed species (CO_2 : $F_{1,61} = 0.09$, p = .77, Figure 2b; Table 2) nor was there an interaction between species and CO_2 treatment on milkweed regrowth tolerance (species* CO_2 : $F_{3,207} = 0.83$, p = .477, Table 2). This result contradicted our original prediction that increased carbon availability and reduced water loss under elevated CO_2 would favor faster rates of regrowth in damaged plants. Milkweed regrowth rate following damage was highest in A. curassavica (10.05 ± 0.45 mg/day)

and lowest in A. syriaca (2.12 \pm 0.34 mg/day) (species: $F_{3,208}$ = 24.27, p < .0001, Figure 2b; Table 2).

Intriguingly, elevated CO_2 weakened the positive relationship between initial plant growth rate and regrowth rate following damage (Regrowth rate* CO_2 : $\mathrm{F}_{1,\,263}$ = 5.99, p = .015, Figure 3; Table 3). In other words, future atmospheric concentrations of CO_2 uncoupled the relationship between regrowth tolerance following damage and initial growth rate before damage. Following mechanical damage, only 278 of the 442 plants (63%) regrew aboveground tissue. Despite previous carbon supplementation, elevated CO_2 did not affect the probability of regrowth (χ^2 = 0.16, p = .6875, Figure 4) nor was there an interaction between milkweed species and CO_2 treatment on regrowth probability (χ^2 = 1.47, p = .689, Figure 4).

Ecology and Evolution

TABLE 2 ANOVA tables of linear mixed effects models used to investigate the four putative trade-offs proposed in this study

							21	atm*sqrt (growth		Random
	species	CO ₂		sqrt (g	rowth rate)	species*CO ₂		ate)		ffect ± SD
F	F _{3,207} = 189.32	F _{1,193} =	0.29	F _{1,195} =	= 2.72	$F_{3,207} = 3.84$	F.	_{1,195} = 0.46	С	hamber
p	< 0.0001	0.5934	6	0.1004	44	0.01047	0	.49931	0	0.06885 ± 0.2624
Trade-c	Trade-off 2: induced resistance \sim sqrt (growth rate) + species + CO_2 + CO_2 *sqrt (growth rate) + random = chamber									
	sqrt (growt	:h rate)	species		CO ₂	species*sqrt (growth rate)	CO ₂ *s	qrt (growth	Rando	m Effect ± SD
F	F _{1,214} = 0.5	8	F _{3,215} = 8.59	9	F _{1,213} = 4.90	F _{3,215} = 1.73	F _{1,215}	= 5.33	chamb	er
р	0.44782		< 0.0001		0.028	0.16224	0.021	88	0.0003	3061 ± 0.0175
	Trade-off 3: sqrt (regrowth rate) \sim log (constitutive) + species + CO_2 + species *CO_2 + random = chamber									
Trade-c	off 3: sqrt (regrowth rate)	~ log (co	nstitutive) +	species ·	+ CO ₂ + specie	es*CO ₂ + random = c	hamber			
Trade-c		~ log (co		species ·		es*CO ₂ + random = c		cies* CO ₂		Random Effect ± SD
Trade-o		nstitutive	e) spe		-	-	spe	cies* CO ₂		
	log (co	nstitutive	e) spe	ecies	27	CO ₂	spe F _{3,2}	-	ĺ	Effect ± SD
F p	log (co F _{1,208} =	nstitutive = 3.66	e) spe F _{3.} < 0	ecies ₂₀₈ = 24.2	27	CO ₂ F _{1,61} = 0.09 0.77115	spe F _{3,2} 0.47	₀₇ = 0.83 7673	(Effect ± SD chamber
F p	log (co F _{1,208} = 0.0571 off 4: induced resistance	nstitutive = 3.66	spe F _{3.} < C growth rate)	ecies ₂₀₈ = 24.2	27 es + CO ₂ + spec	CO ₂ F _{1,61} = 0.09 0.77115	sper F _{3,2} 0.47 rate) + ran	₀₇ = 0.83 7673 ndom = chambe	er	Effect ± SD chamber
F p	log (co F _{1,208} = 0.0571 off 4: induced resistance	nstitutive 3.66 6 ~ sqrt (reg	F _{3.} < C growth rate)	ecies ₂₀₈ = 24.2 0.0001 + specie	27 es + CO ₂ + spec	CO ₂ F _{1,61} = 0.09 0.77115 cies*sqrt (regrowth i	sper F _{3,2} ; 0.47 rate) + ran specie (regro	ndom = chambers*sqrt wth rate)	er	Effect ± SD chamber 0.02114 ± 0.1454 n Effect ± SD

Note: Model selection was performed using maximum likelihood. Tables were produced with the R package LmerTest, using type III sums of squares with Satterthwaite approximation for degrees of freedom, random effects estimates ± 1 standard deviation, and fixed effects parameter estimates ± 1 standard deviation.

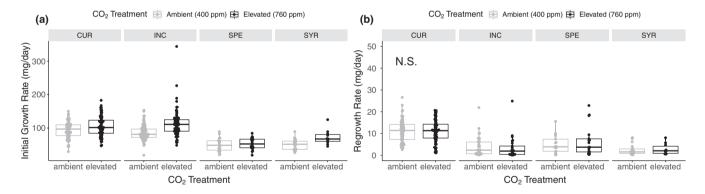
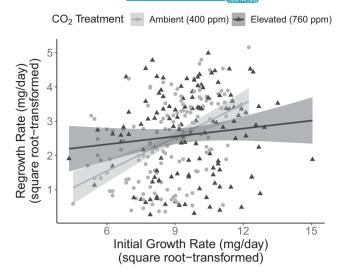


FIGURE 2 Elevated CO₂ increased initial milkweed growth rate but had no lasting effects on regrowth rate following damage. The effects of CO₂ treatment and milkweed species on (a) initial growth rate prior to damage (mg dry mass of above-ground tissue/64 days) and (b) nonsignificant effects of elevated CO2 and milkweed species on regrowth rate following damage (mg dry mass of above-ground tissue/21 days). In boxplots, dark lines represent the median, box boundaries represent first and third quartiles, and whiskers extend to the most extreme data point less than 1.5 times the interquartile range from the box. Milkweed species codes are the same as above. Data are grouped by species and CO2 treatment for ease of interpretation; however, the interaction term was not retained in our models of regrown plants

3.3 | Elevated CO₂ altered the magnitude and diversity of chemical resistance

Elevated CO2 reduced constitutive resistance in A. incarnata by 37%, in A. syriaca by 10%, slightly in A. curassavica by 5%, and increased constitutive resistance in A. speciosa by 22% (species*CO₂: $F_{3,207}$ = 3.84, p = .010, Figure 5a; Table 2). Milkweed species was by far the most important determinant of constitutive cardenolide concentration (species: $F_{3,207}$ = 189.32, p < .0001, Figure 5a; Table 2). In those plants that did regrow following damage, mechanically induced resistance varied substantially by milkweed species (species: $F_{3,215}$ = 8.59, p < .0001, Figure 5b; Table 2). A. curassavica again produced the highest concentrations of foliar cardenolides, followed by A. speciosa, A. syriaca, and A. incarnata. Across all four species,



<code>FIGURE 3</code> Elevated CO_2 uncoupled the positive relationship between initial plant growth rate and regrowth rate following damage. Light gray circles and lines represent plants grown under ambient CO_2 and dark gray triangles and lines are those grown under elevated CO_2 . Regressions are represented with 95% confidence intervals

mechanically induced resistance increased by 28% in those plants previously exposed to elevated CO_2 (CO_2 : $F_{1,213}$ = 4.90 p = .028, Figure 5b; Table 2).

The diversity of cardenolides produced constitutively among milkweed species increased by 24% under elevated CO_2 (CO_2 : $F_{1,68}$ = 4.08, p = .047, Figure 5c; Table 4). Despite a species-specific effect of elevated CO_2 on the total concentration of constitutive resistance, there was no such effect on the diversity of cardenolides produced constitutively (species* CO_2 : $F_{3,206}$ = 2.04, p = .109, Figure 5c; Table 4). Conversely, the diversity of cardenolides produced in the mechanically induced resistance profiles of both A. *incarnata*, and A. *speciosa* declined by 70% and 11% after previous exposure to elevated CO_2 (species* CO_2 : $F_{3,20}$ = 2.67, p = .048, Figure 5d; Table 4).

When comparing the composition of cardenolide communities among individuals before and after damage, the difference between constitutive and mechanically induced foliar tissue was

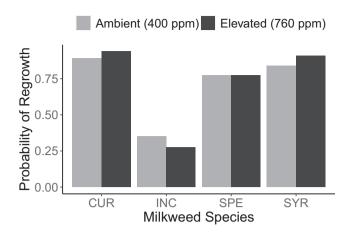


FIGURE 4 Variation among milkweed species in the probability of regrowth after mechanical damage. Light gray bars represent plants grown under ambient CO_2 and dark gray bars are those grown under elevated CO_2 . Data are grouped by species and CO_2 treatment for ease of interpretation; however, the interaction term was not significant in the models. Milkweed species codes are the same as above

the strongest driver of community dissimilarity as determined by PerMANOVA (resistance type: $F_{1, 410} = 55.38$, p = .001, $R^2 = 0.15$, Figure 6; Table 5). There were slight differences between these two resistance profiles among milkweed species driven by elevated CO_2 (resistance type*species* CO_2 : $F_{2, 410} = 2.39$, p = .001, $R^2 = 0.013$, Figure 6; Table 5), and these slight differences likely represent the changes in cardenolide diversity detected above.

4 | DISCUSSION

Our study reveals the limitations of a trade-off framework at the physiological level when considering how complex defense phenotypes respond to environmental change. Of the four hypothesized trade-offs among aspects of plant growth and resistance framing the study, we found support for only one between regrowth tolerance and mechanically induced chemical resistance (foliar cardenolide concentration following mechanical damage). The strength of this trade-off was unaffected by previous exposure to elevated CO₂ but

 $\textbf{TABLE 3} \quad \text{ANOVA table of a linear mixed effects model describing the effects of elevated CO}_2 \text{ on the relationship between initial plant growth rate and regrowth rate following damage}$

Mode	Model: sqrt (regrowth rate) ~ sqrt (growth rate) + CO ₂ + species + species* sqrt (growth rate) + CO ₂ * sqrt (growth rate) + random = chamber								
	sqrt (growth rate)	species	CO ₂	species*sqrt (growth rate)	species* CO ₂	CO ₂ *sqrt (growth rate)	Random Effect ± SD		
F	F _{1,261} = 0.01	F _{3,257} = 2.88	F _{1,260} = 5.95	F _{3,257} = 2.14	F3 _{,260} = 1.17	F _{1,263} = 5.99	chamber		
р	0.90362	0.03633	0.01543	0.09542	0.32088	0.01505	0.07448 ± 0.2729		

Note: As above, model selection was performed using maximum likelihood. Tables were produced with the R package LmerTest, using type III sums of squares with Satterthwaite approximation for degrees of freedom, random effects estimates \pm 1 standard deviation, and fixed effects parameter estimates \pm 1 standard deviation.

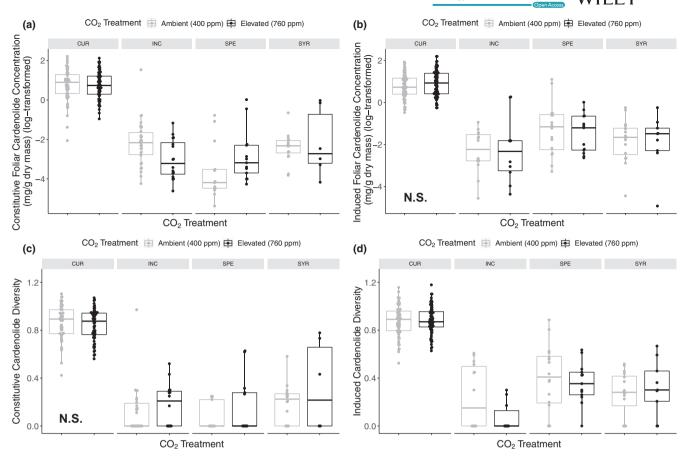


FIGURE 5 Elevated CO_2 altered the total concentration of milkweed constitutive defense and the diversity of mechanically induced defense following damage. (a) The effects of elevated atmospheric concentrations of CO_2 on constitutive cardenolide concentrations of milkweed (mg/g dry mass), (b) the mechanically induced cardenolide concentrations of milkweeds (mg/g dry mass), (c) the diversity of cardenolides produced constitutively, and (d) the diversity of cardenolides produced in the mechanically induced resistance response following damage. Data are grouped by species and CO_2 treatment for ease of interpretation; however, the interaction term was not retained in our models for B and C. Dark gray points represent plants grown under elevated CO_2 and light gray points and lines are those grown under ambient CO_2 . Milkweed species codes are the same as above

varied substantially among milkweed species, presumably reflecting species-specific allocation patterns to defense following damage. In contrast to expected trade-offs, we found positive relationships among some growth and resistance traits. However, the positive relationship between growth rate prior to damage and mechanically induced chemical resistance was weaker under previous exposure to elevated CO_2 . Our data add to a growing body of work that demonstrates the complex nature of plant growth and resistance relationships and highlights the need to test allocation strategies of plants in the context of rapidly changing environmental resources on ecological time scales as well as across evolutionary contexts.

Multiple mechanisms may govern the direction and magnitude of growth and resistance relationships in plants. These mechanisms include nutrient limitation, allocation costs, genetic linkage of defense traits, and ecological costs (Boege, Dirzo, Siemens, & Brown, 2007;Fine, Miller, & Mesones, 2006;Simms & Rausher, 1987;Strauss, Siemens, Decher, & Mitchell-Olds, 1999;Tao et al., 2016;Tucker & Avila-Sakar, 2010;Wise & Abrahamson, 2007;Züst & Agrawal, 2017). Among plants that

regrew following damage, we found evidence of a trade-off between mechanically induced cardenolide concentrations and regrowth tolerance in three of four milkweed species (Figure 1d). This finding supports previous studies that have reported negative relationships between milkweed growth and cardenolide production (Hochwender et al., 2000; Tao et al., 2016; Züst, Rasmann, & Agrawal, 2015). However, ours is the first study within the milkweed system to show interspecific differences in regrowth tolerance and mechanically induced resistance relationships following damage. Interestingly, previous exposure to elevated CO2 had no effect on the strength of this trade-off, indicating that the legacy of carbon supplementation in isolation may not be a critical driver of plant induced defense syndromes. Only the tropical A. curassavica, native to central America, failed to display a trade-off between mechanically induced resistance and regrowth tolerance. Higher herbivore pressure at southern latitudes may select for higher levels of both defense traits in this species as compared to the other three perennials native to N. America (Rasmann & Agrawal, 2011). The positive relationship between innate plant growth and mechanically induced

TABLE 4 ANOVA tables of linear mixed effects models describing the relationships between the diversity of constitutive and induced cardenolides and growth rates dependent on milkweed species and elevated CO₂

Model: constitutive diversity \sim sqrt (growth rate) + species + CO_2 + species $*CO_2$ + random = chamber									
	sqrt (growth rate)	species		CO ₂	species* CO ₂	Random Effect ± SD			
F	F _{1,201} = 0.76	F _{3,207} = 260	0.56	F _{1,68} = 4.077	F _{3,206} = 2.04	chamber			
р	0.38452	<0.0001		0.04741	0.10937	0.003 ± 0.054			
Model: in	duced diversity ~ sqrt (regrow	rth rate) + species +	CO ₂ + sqrt (regrow	vth rate)*species + spe	ecies*CO ₂ + random =	chamber			
	sgrt (regrowth			sgrt (regrowth		Random			
	sqrt (regrowth rate)	species	CO ₂	sqrt (regrowth rate)*species	species* CO ₂				
F		species F _{3,205} = 29.67	CO ₂ F _{1,72} = 1.95	• • •	species* CO ₂ F _{3,203} = 2.67	Random			

Note:: As above, model selection was performed using maximum likelihood. Tables were produced with the R package LmerTest, using type III sums of squares with Satterthwaite approximation for degrees of freedom, random effects estimates \pm 1 standard deviation, and fixed effects parameter estimates \pm 1 standard deviation.

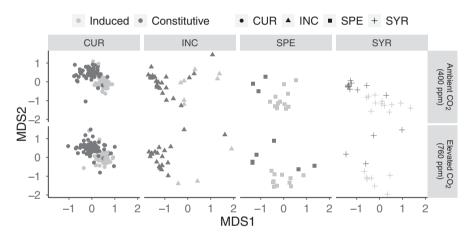


FIGURE 6 Previous exposure to elevated CO_2 caused slight changes in both the constitutive and mechanically induced cardenolide communities of milkweed. Dark gray points represent the constitutive cardenolide communities produced by plants before damage, and light gray points are the cardenolide communities detected in the mechanically induced response of milkweed following damage. Those plants grown under ambient CO_2 are in the upper panel and those grown under elevated CO_2 are in the lower panel. Milkweed species codes are the same as above

TABLE 5 PerMANOVA describing the effects of elevated CO₂ on the composition of constitutive and induced cardenolide communities

PerMANOVA						
	F	R ²	Р			
species	F _{3,410} = 12.12	0.10	.001			
CO ₂	$F_{1,410} = 1.06$	0.003	.404			
resistance type	$F_{1,410} = 55.38$	0.15	.001			
resistance type*species	$F_{3,410} = 4.74$	0.04	.001			
resistance type*CO ₂	$F_{1,410} = 1.04$	0.003	.391			
species*CO ₂	$F_{2,410} = 2.26$	0.011	.003			
resistance type*species*CO ₂	$F_{2,410} = 2.39$	0.013	.001			

resistance could also reflect selection for vigorous plants capable of mounting a strong response to herbivory (Hahn et al., 2019; Figure 1c). Interestingly, with faster predamage growth rates under elevated CO₂, plants produced lower levels of induced resistance likely as a result of suppressed phytohormonal signaling pathways (Guo et al., 2012;Ode et al., 2014).

Despite finding no influence of elevated CO_2 on three of the four relationships between growth and resistance in our study, elevated CO_2 altered aspects of both milkweed growth and resistance independently. Notably, elevated CO_2 uncoupled the positive relationship between initial plant growth rate and regrowth tolerance following damage (Figure 3). Often plants with high innate growth rates can regrow faster following damage (Rosenthal & Kotanen, 1994). However, in our study, those plants that were fast growing under elevated CO_2 did not maintain a proportionately high level of regrowth

under ambient CO₂ following damage. Because the regrowth period took place in a greenhouse under ambient CO2 with homogenous soil nutrients and water availability, these data potentially indicate the legacy of elevated CO₂ in altering phytohormonal signaling pathways responsible for regrowth tolerance (Guo et al., 2012). The constitutive resistance of both A. incarnata and A. syriaca declined under elevated CO2 and increased in A. speciosa. Despite these effects of elevated CO2 on constitutive defense, no legacy of this treatment was detected in the mechanically induced resistance response of the milkweed species. Such conserved induction responses despite previous exposure to elevated CO2 suggests that changes in chemical resistance due to elevated CO2 detected by this and other studies (Ode et al., 2014; Zavala, Gog, & Giacometti, 2017; Zavala et al., 2013) rely on continuous carbon supplementation and simultaneous manipulation of phytohormonal signaling pathways rather than previous allocation decisions made by the plant before damage.

Monarch caterpillars are iconic milkweed herbivores undergoing significant declines, due, in part, to changing environmental conditions in both overwintering and summer breeding grounds (Stephen Malcolm, 2017; Stenoien et al., 2016). Roadside milkweed patches are important habitat for monarchs and regularly experience mowing events (Kasten, Stenoien, Caldwell, & Oberhauser, 2016; Mueller & Baum, 2014). Appropriately timed mowing treatments can increase monarch fecundity within milkweed patches by increasing the availability of high-quality foliage and releasing monarchs from the presence of enemies (Borkin, 1982; Fischer, Williams, Brower, & Palmiotto, 2015; Haan & Landis, 2019; Knight, Norris, Derbyshire, & Flockhart, 2019). Our study reveals that elevated CO₂ changes the composition and reduces the diversity of cardenolides produced after simulated mowing in both A. incarnata, and A. speciosa, two milkweed species commonly found in the N. American summer breeding grounds (Woodson, 1954). Critically, the composition of cardenolide communities produced by milkweed can alter monarch interactions with natural enemies, such as a prevalent protozoan pathogen (Decker, Roode, & Hunter, 2018; Decker, Soule, Roode, & Hunter, 2019; Sternberg et al., 2012). Given the conservation importance of roadside milkweed patches that are regularly mowed throughout N. America, changes in regrowth tissue chemical quality could have implications for monarch populations. Yet, attempts to predict how migratory monarchs that depend on roadside milkweed corridors will perform under global environmental change remain challenging (Zipkin, Ries, Reeves, Regetz, & Oberhauser, 2012).

Our study, though comprehensive in its investigation of growth and chemical resistance before and after damage, does not incorporate the entire suite of defenses expressed by milkweeds. Additional direct and indirect defenses include trichomes, latex, leaf toughness, and volatile emissions that attract natural enemies (Agrawal & Fishbein, 2006;Agrawal & Konno, 2009;Hochwender et al., 2000;Meier & Hunter, 2019;Zalucki et al., 2001). This suite of defense strategies may also generate resource-based trade-offs and alter plant-herbivore interactions (Züst & Agrawal, 2017;Züst et al., 2015). Thus, further studies exploring the fitness costs of

regrowth tolerance and multiple defenses under future environmental conditions, and the responses of herbivore populations to these changes, are greatly needed.

On an evolutionary timescale, the influence of resource clines has illustrated the existence of trade-offs between growth and resistance, lending broad support to the RAH (Coley & Chapin, 1985; Endara & Coley, 2011; Strauss & Agrawal, 1999). Currently, no well-established theory makes predictions about how trade-offs among defense traits will respond to rapid environmental change within one generation. In our study, the identity of the milkweed species determined our ability to detect a trade-off between regrowth tolerance and resistance following mechanical damage, and previous exposure to elevated CO2 weakened a positive relationship between innate growth rate and constitutive defense. Given the rapid rate of environmental change predicted globally (Stocker et al., 2013), studies measuring the rate of plant resistance and growth evolution as well as which environmental change drivers are crucial determinants of plant fitness will be vital to predicting plant-insect interactions. This knowledge can be used to inform policy decisions which reduce the use of pesticides (Strauss & Murch, 2004) and improve weed control programs (Williams, Walsh, & Boydston, 2004).

ACKNOWLEDGMENTS

Many thanks to H., Streit, R. Ricart, C. Chappell, R. Peterson, J. Kristofik, B. Shilling, K. Moriarty, A. Meier, K. Crocker, K. Sanchez, J. Den Uyl, A. Potts. This work was supported by National Science Foundation grants DEB-1256115 and DEB-1257160.

CONFLICT OF INTEREST

Authors have no sources of conflict of interest.

AUTHOR CONTRIBUTION

Leslie E Decker: Conceptualization (equal); Data curation (lead); Formal analysis (lead); Funding acquisition (supporting); Investigation (lead); Methodology (equal); Project administration (lead); Resources (equal); Visualization (lead); Writing-original draft (lead); Writing-review & editing (lead). Mark D. Hunter: Conceptualization (equal); Data curation (supporting); Formal analysis (supporting); Funding acquisition (lead); Investigation (supporting); Methodology (equal); Project administration (supporting); Resources (lead); Supervision (equal); Visualization (supporting); Writing-original draft (supporting); Writing-review & editing (supporting).

AUTHOR CONTRIBUTIONS

LED and MDH designed the experiment, collected and analyzed the data. LED wrote the manuscript, the MDH contributed significantly to drafts and approved the final version.

DATA AVAILABILITY STATEMENT

Data are available in the Dryad Digital Repository: https://doi.org/10.5061/dryad.v6wwpzgs3

ORCID

Leslie E. Decker https://orcid.org/0000-0003-1325-052X

Mark D. Hunter https://orcid.org/0000-0003-3761-8237

REFERENCES

- Agrawal, A. A. (2011). Current trends in the evolutionary ecology of plant defence. *Functional Ecology*, 25, 420-432. https://doi.org/10.1111/j.1365-2435.2010.01796.x
- Agrawal, A. A., & Fishbein, M. (2006). Plant defense syndromes. *Ecology*, 87. S132–S149.
- Agrawal, A. A., & Fishbein, M. (2008). Phylogenetic escalation and decline of plant defense strategies. Proceedings of the National Academy of Sciences, 105, 10057–10060. https://doi.org/10.1073/pnas.08023 68105
- Agrawal, A. A., & Karban, R. (1999). Why induced defenses may be favored over constitutive strategies in plants. *The Ecology and Evolution of Inducible Defenses*, https://doi.org/10.1159/000331745
- Agrawal, A. A., & Konno, K. (2009). Latex: A Model for Understanding Mechanisms, Ecology, and Evolution of Plant Defense Against Herbivory. *Annual Review of Ecology Evolution and Systematics*, 40, 311–331. https://doi.org/10.1146/annurev.ecolsys.110308.120307
- Agrawal, A. A., Petschenka, G., Bingham, R. A., Weber, M. G., & Rasmann, S. (2012). Toxic cardenolides: Chemical ecology and coevolution of specialized plant-herbivore interactions. *New Phytologist*, 194, 28– 45. https://doi.org/10.1111/j.1469-8137.2011.04049.x
- Ainsworth, E. A., & Long, S. P. (2005). What have we learned from 15 years of free air CO enrichment (FACE)? A meta-analytic review of the response of photosynthesis, canopy properties and plant production to rising CO. *The New Phytologist*, 179, 5.
- Anderson, M. J. (2001). A new method for non parametric multivariate analysis of variance. *Austral Ecology*, *26*, 32–46. https://doi.org/10.11 11/j.1442-9993.2001.01070.pp.x
- Bain, J. C., & Day, F. P. (2019). Legacy effects of long-term CO2 enrichment on plant biomass recovery from fire seven years after return to ambient CO2 levels1. The Journal of the Torrey Botanical Society, 146, 1. https://doi.org/10.3159/torrey-d-17-00044.1
- Baucom, R. S., & De Roode, J. C. (2011). Ecological immunology and tolerance in plants and animals. *Functional Ecology*, 25, 18–28. https://doi.org/10.1111/j.1365-2435.2010.01742.x
- Bazzaz, F., Ackerly, D., Woodward, F., & Rochefort, L. (2013). CO enrichment and dependence of reproduction on density in an annual plant and a simulation of its population dynamics. *Journal of Ecology*, 80(4), 643–651.
- Bekaert, M., Edger, P. P., Hudson, C. M., Pires, J. C., & Conant, G. C. (2012). Metabolic and evolutionary costs of herbivory defense: Systems biology of glucosinolate synthesis. *New Phytologist*, 196(2), 596–605. https://doi.org/10.1111/j.1469-8137.2012.04302.x
- Bidart-Bouzat, M. G., Mithen, R., & Berenbaum, M. R. (2005). Elevated ${\rm CO_2}$ influences herbivory-induced defense responses of Arabidopsis thaliana. Oecologia, 145, 415–424. https://doi.org/10.1007/s0044 2-005-0158-5
- Boege, K., Dirzo, R., Siemens, D., & Brown, P. (2007). Ontogenetic switches from plant resistance to tolerance: Minimizing costs with age? *Ecology Letters*, 10, 177–187. https://doi.org/10.1111/j.1461-0248.2006.01012.x
- Borkin, S. S. (1982). The Great Lakes Entomologist Notes on Shifting Distribution Patterns and Survival of Immature Danaus Plexippus (Lepidoptera: Danaidae) on the Food Plant Asclepias syriaca. *The Great Lakes Entomologist*, 15.
- Burnham, K. P., & Anderson, D. R. (2002). Model Selection and Multimodel Inference: A Practical Information-Theoretic Approach, (2nd ed.). New York, NY: Springer Science & Business Media.
- Coley, D. C., & Chapin, F. S. (1985). Resource Availability and Plant Antiherbivore Defense. *Science* (80-), 230, 895–899. https://doi.org/10.1126/science.171.3973.757

- Crawley, M. J. (2012). Statistical Modelling. Chichester, UK: John Wiley & Sons Ltd.
- Decker, L. E., de Roode, J. C., & Hunter, M. D. (2018). Elevated atmospheric concentrations of carbon dioxide reduce monarch tolerance and increase parasite virulence by altering the medicinal properties of milkweeds. *Ecology Letters*, 21, 1353–1363. https://doi.org/10.1111/ele.13101
- Decker, L. E., Soule, A. J., de Roode, J. C., & Hunter, M. D. (2019). Phytochemical changes in milkweed induced by elevated CO 2 alter wing morphology but not toxin sequestration in monarch butterflies. Functional Ecology, https://doi.org/10.1111/1365-2435.13270
- Drake, B. G., Gonzalez-Meler, M. A., & Long, S. P. (1997). More efficient plants: A consequence of rising atmospheric CO? Annual Review of Plant Physiology and Plant Molecular Biology, 48, 609–639. https://doi. org/10.1146/annurev.arplant.48.1.609
- Drake, B. G., Leadley, P. W., Arp, W. J., Nassiry, D., & Curtis, P. S. (1989). An open top chamber for field studies of elevated atmospheric CO_2 concentration on saltmarsh vegetation. Functional Ecology, 3, 363. https://doi.org/10.2307/2389377
- Endara, M. J., & Coley, P. D. (2011). The resource availability hypothesis revisited: A meta-analysis. *Functional Ecology*, *25*(2), 389–398. https://doi.org/10.1111/j.1365-2435.2010.01803.x
- Fine, P. V. A., Miller, Z. J., Mesones, I. et al (2006). The growth-defense trade-off and habitat specialization by plants in Amazonian forests. *Ecology*, 87, https://doi.org/10.1890/0012-9658(2006)87[150:TG-TAHS]2.0.CO:2
- Fineblum, W. L., Rausher, M. D. (1995). Tradeoff between resistance and tolerance to herbivore damage in a morning glory. *Nature*, 377, 517–520. https://doi.org/10.1038/377517a0
- Fischer, S. J., Williams, E. H., Brower, L. P., & Palmiotto, P. A. (2015). Enhancing monarch butterfly reproduction by mowing fields of common milkweed. *American Midland Naturalist*, 173, 229–240. https://doi.org/10.1674/amid-173-02-229-240.1
- Fornoni, J. (2011). Ecological and evolutionary implications of plant tolerance to herbivory. *Functional Ecology*, 25, 399–407. https://doi.org/10.1111/j.1365-2435.2010.01805.x
- Fornoni, J., Nú Ñ Ez-Farfán J., Valverde P. L. (2003). Evolutionary ecology of tolerance to herbivory: Advances and perspectives. *Comments on Theoretical Biology*, *8*, 643–663. https://doi.org/10.1080/08948 550390234703
- Gershenzon, J. (1994). Metabolic costs of terpenoid accumulation in higher plants. *Journal of Chemical Ecology*, 20(6), 1281–1328. https://doi.org/10.1007/BF02059810
- Guo, H., Sun, Y., Ren, Q., Zhu-Salzman, K., Kang, L. E., Wang, C., ... Ge, F. (2012). Elevated CO₂ Reduces the Resistance and Tolerance of Tomato Plants to *Helicoverpa armigera* by Suppressing the JA Signaling Pathway. PLoS ONE, 7, e41426. https://doi.org/10.1371/journal.pone.0041426
- Haan, N. L., & Landis, D. A. (2019). Grassland disturbance increases monarch butterfly oviposition and decreases arthropod predator abundance. *Biological Conservation*, 233, 185–192. https://doi. org/10.1016/j.biocon.2019.03.007
- Hahn, P. G., Agrawal, A. A., Sussman, K. I., & Maron, J. L. (2019). Population variation, environmental gradients, and the evolutionary ecology of plant defense against herbivory. *American Naturalist*, 193(1), 20–34. https://doi.org/10.1086/700838
- Hahn, P. G., & Maron, J. L. (2016). A framework for predicting intraspecific variation in plant defense. *Trends in Ecology & Evolution*, 31(8), 646–656. https://doi.org/10.1016/j.tree.2016.05.007
- Herms, D. A., & Mattson, W. J. (1992). The dilemma of plants: To grow or defend. The Quarterly Review of Biology, 67, 283–335. https://doi. org/10.1086/417659
- Hochwender, C. G., Marquis, R. J., & Stowe, K. A. (2000). The potential for and constraints on the evolution of compensatory ability in *Asclepias syriaca*. *Oecologia*, 122, 361–370. https://doi.org/10.1007/s004420050042



- Hungate, B. A., Johnson, D. W., Dijkstra, P. et al (2006). Nitrogen cycling during seven years of atmospheric CO2 enrichment in a scrub oak woodland. *Ecology*, https://doi.org/10.1890/04-1732
- Hunter, M. D. (2001). Effects of elevated atmospheric carbon dioxide on insect-plant interactions. Agricultural and Forest Entomology, 3, 153– 159. https://doi.org/10.1046/j.1461-9555.2001.00108.x
- Jia, X., Zhao, Y., Liu, T., & Huang, S. (2016). Elevated CO₂ affects secondary metabolites in Robinia pseudoacacia L. seedlings in Cd- and Pb- contaminated soils. *Chemosphere*, 160, 199–207. https://doi.org/10.1016/J.CHEMOSPHERE.2016.06.089
- Karban, R., & Baldwin, I. T. (2013). Induced Responses to Herbivory. Chicago, IL: University of Chicago Press.
- Kasten, K., Stenoien, C., Caldwell, W., & Oberhauser, K. S. (2016). Can roadside habitat lead monarchs on a route to recovery? *Journal of Insect Conservation*, 20, 1047–1057. https://doi.org/10.1007/s1084 1-016-9938-y
- Kimball, B. A. (1983). Carbon dioxide and agricultural yield: An assemblage and analysis of 430 prior observations. *Agronomy Journal*, 75, 779. https://doi.org/10.2134/agronj1983.000219620075000 50014x
- Klaiber, J., Dorn, S., & Najar-Rodriguez, A. J. (2013). Acclimation to elevated CO₂ increases constitutive glucosinolate levels of brassica plants and affects the performance of specialized herbivores from contrasting feeding guilds. *Journal of Chemical Ecology*, 39, 653–665. https://doi.org/10.1007/s10886-013-0282-3
- Knight, S. M., Norris, D. R., Derbyshire, R., & Flockhart, D. T. T. (2019). Strategic mowing of roadside milkweeds increases monarch butterfly oviposition. *Global Ecology and Conservation*, 19, e00678. https://doi.org/10.1016/j.gecco.2019.e00678
- Lau, J. A., & Tiffin, P. (2009). Elevated carbon dioxide concentrations indirectly affect plant fitness by altering plant tolerance to herbivory. *Oecologia*, 161, 401–410. https://doi.org/10.1007/s0044 2-009-1384-z
- Leadley, P. W., Niklaus, P. A., Stocker, R., & Körner, C. (1999). A field study of the effects of elevated CO₂ on plant biomass and community structure in a calcareous grassland. *Oecologia*, 118, 39–49. https:// doi.org/10.1007/s004420050701
- Malcolm, S. B. (2017). Anthropogenic impacts on mortality and population viability of the monarch butterfly. *Annual Review of Entomology*, 63, 277–302. https://doi.org/10.1146/annurev-ento-02011 7-043241
- Malcolm, S. B., & Zalucki, M. P. (1996). Milkweed latex and cardenolide induction may resolve the lethal plant defence paradox. *Entomologia Experimentalis Et Applicata*, 80, 193–196. https://doi.org/10.1111/j.1570-7458.1996.tb00916.x
- Marshall, C. B., Avila-Sakar, G., & Reekie, E. G. (2008). Effects of nutrient and CO₂ availability on tolerance to herbivory in Brassica rapa. *Plant Ecology*, 196, 1-13. https://doi.org/10.1007/s1125 8-007-9331-8
- Matiella, T. J. (2012) The effects of carbon dioxide on three Species of milkweed (*Asclepiadaceae*) and Monarch butterfly (*Danaus plexippus*) larva feeding preference. Retrieved from ProQuest Diss. Theses.
- McCloud, E. S., & Baldwin, I. T. (1997). Herbivory and caterpillar regurgitants amplify the wound-induced increases in jasmonic acid but not nicotine in Nicotiana sylvestris. *Planta*, 203(4), 430–435. https://doi.org/10.1007/s004250050210
- Meier, A. R., & Hunter, M. D. (2019). Mycorrhizae alter constitutive and herbivore-induced volatile emissions by milkweeds. *Journal of Chemical Ecology*, 45, 610–625. https://doi.org/10.1007/s1088 6-019-01080-6
- Mueller, E. K., & Baum, K. A. (2014). Monarch-parasite interactions in managed and roadside prairies. *Journal of Insect Conservation*, 18, 847–853. https://doi.org/10.1007/s10841-014-9691-z
- Núñez-Farfán, J., Fornoni, J., & Valverde, P. L. (2007). The evolution of resistance and tolerance to herbivores. *Annual Review of Ecology*

- Evolution and Systematics, 38, 541–566. https://doi.org/10.1146/annurev.ecolsys.38.091206.095822
- Ode, P. J., Johnson, S. N., & Moore, B. D. (2014). Atmospheric change and induced plant secondary metabolites are we reshaping the building blocks of multi-trophic interactions? *Curr Opin Insect Sci*, *5*, 57–65. https://doi.org/10.1016/j.cois.2014.09.006
- Oksanen, B. F., Friendly, M. et al (2017). Vegan: An introduction to ordination. Vegan Community Ecol. Packag. R Packag. version 2.4-3.
- R Core Team (2019). R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing: Vienna, Austria.
- Rasmann, S., & Agrawal, A. A. (2011). Latitudinal patterns in plant defense: Evolution of cardenolides, their toxicity and induction following herbivory. *Ecology Letters*, 14, 476–483. https://doi.org/10.1111/j.1461-0248.2011.01609.x
- Rhoades, D. F. (1985). Offensive-defensive interactions between herbivores and plants: Their relevance in herbivore population dynamics and ecological theory. American Naturalist, 125, 205–238. https://doi.org/10.1086/284338
- Robinson, E. A., Ryan, G. D., & Newman, J. A. (2012). A meta-analytical review of the effects of elevated CO₂ on plant-arthropod interactions highlights the importance of interacting environmental and biological variables. *New Phytologist*, 194, 321–336. https://doi.org/10.1111/j.1469-8137.2012.04074.x
- Rosenthal, J. P., & Kotanen, P. M. (1994). Terrestrial plant tolerance to herbivory. *Trends in Ecology & Evolution*, 9, 145–148. https://doi.org/10.1016/0169-5347(94)90180-5
- Ryan, G. D., Rasmussen, S., & Newman, J. A. (2010). Global atmospheric change and trophic interactions: Are there any general responses? *Plant Commun from an Ecoogical Perspect*, 179–214, https://doi.org/10.1007/978-3-642-12162-3_11
- Simms, E. L., & Rausher, M. D. (1987). Costs and benefits of plant resistance to herbivory. *American Naturalist*, 130, 570–581. https://doi.org/10.1086/284731
- Stamp, N. (2003). Out of the quagmire of plant defense hypotheses. *The Quarterly Review of Biology*, 78, 23-55. https://doi.org/10.1086/367580
- Stamp, N. (2004). Can the growth-differentiation balance hypothesis be tested rigorously? *Oikos*, 107, 439–448. https://doi.org/10.1111/j.0030-1299.2004.12039.x
- Stenoien, C., Nail, K. R., Zalucki, J. M., Parry, H., Oberhauser, K. S., & Zalucki, M. P. (2016). Monarchs in decline: A collateral land-scape-level effect of modern agriculture. *Insect Science*, https://doi.org/10.1111/1744-7917.12404
- Sternberg, E. D., Lefevre, T., Li, J., De Castillejo, C. L. F., Li, H., Hunter, M. D., & De Roode, J. C. (2012). Food plant-derived disease tolerance and resistance in a natural butterfly- plant-parasite interactions. *Evolution (N Y)*, 66, 3367–3377. https://doi.org/10.5061/dryad.82j66
- Stiling, P., Moon, D., Rossi, A. et al (2013). Direct and legacy effects of long-term elevated CO $_2$ on fine root growth and plant-insect interactions. *New Phytologist*, 200, 788–795. https://doi.org/10.1111/nph.12295
- Stocker, T. F., Qin, D., Plattner, G. K., Tignor, M., Allen, S. K., Boschung, J., ... Midgley, P. M. (Eds.) (2013). Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. https://doi.org/10.1017/CBO9781107415324
- Strauss, S. Y., & Agrawal, A. A. (1999). The ecology and evolution of plant tolerance to herbivory. *Trends in Ecology & Evolution*, 14, 179–185. https://doi.org/10.1016/S0169-5347(98)01576-6
- Strauss, S. Y., & Murch, P. (2004). Towards an understanding of the mechanisms of tolerance: Compensating for herbivore damage by enhancing a mutualism. *Ecological Entomology*, 29, 234–239. https://doi.org/10.1111/j.0307-6946.2004.00587.x
- Strauss, S. Y., Rudgers, J. A., Lau, J. A., & Irwin, R. E. (2002). Direct and ecological costs of resistance to herbivory. Trends in Ecology & Evolution, 17(6), 278–285. https://doi.org/10.1016/S0169-5347(02)02483-7

- Strauss, S. Y., Siemens, D. H., Decher, M. B., & Mitchell-Olds, T. (1999). Ecological costs of plant resistance to herbivores in the currency of pollination. *Evolution (N Y)*, 53, 1105–1113. https://doi.org/10.1111/j.1558-5646.1999.tb04525.x
- Tao, L., Ahmad, A., de Roode, J. C., & Hunter, M. D. (2016). Arbuscular mycorrhizal fungi affect plant tolerance and chemical defences to herbivory through different mechanisms. *Journal of Ecology*, 104, 561–571. https://doi.org/10.1111/1365-2745.12535
- Tao, L., Berns, A. R., & Hunter, M. D. (2014). Why does a good thing become too much? Interactions between foliar nutrients and toxins determine performance of an insect herbivore. Functional Ecology, 28, 190–196. https://doi.org/10.1111/1365-2435.12163
- Tao, L., & Hunter, M. D. (2012). Does anthropogenic nitrogen deposition induce phosphorus limitation in herbivorous insects? Global Change Biology, 18, 1843–1853. https://doi. org/10.1111/j.1365-2486.2012.02645.x
- Tucker, C., & Avila-Sakar, G. (2010). Ontogenetic changes in tolerance to herbivory in Arabidopsis. *Oecologia*, 164, 1005–1015. https://doi. org/10.1007/s00442-010-1738-6
- Tuomi, J., Niemelä, P., Chapin, F. S., Bryant, J. P., & Sirén, S., (1988).
 Defensive responses of trees in relation to their carbon/nutrient balance. In *Mechanisms of woody plant defenses against insects* (pp. 57–72). Springer, New York, NY.
- van der Meijden, E., Wijn, M., & Verkaar, H. J. (1988). Defence and regrowth, alternative plant strategies in the struggle against herbivores. *Oikos*, *51*, 355. https://doi.org/10.2307/3565318
- Vannette, R. L., & Hunter, M. D. (2011). Genetic variation in expression of defense phenotype may mediate evolutionary adaptation of *Asclepias syriaca* to elevated CO₂. *Glob Chang Biol*, 17, 1277–1288. https://doi.org/10.1111/j.1365-2486.2010.02316.x
- Williams, M. M., Walsh, D. B., & Boydston, R. A. (2004). Integrating arthropod herbivory and reduced herbicide use for weed management. Weed Science, 52, 1018–1025. https://doi.org/10.1614/WS-04-034R
- Wilsey, B. J. (2001). Effects of Elevated CO2 on the Response of *Phleum pratense* and *Poa pratensis* to Aboveground Defoliation and Root-Feeding Nematodes. *International Journal of Plant Sciences*, 162, 1275–1282. https://doi.org/10.1086/323702
- Wise, M. J., & Abrahamson, W. G. (2007). Effects of resource availability on tolerance of herbivory: A review and assessment of three opposing models. The American Naturalist, 169, 443–454. https://doi.org/10.1086/512044
- Woodson, R. E. (1954). The North American species of Asclepias L. Annals of the Missouri Botanical Garden, 41, 1–211. https://doi.org/10.2307/2394652
- Zalucki, M. P., Brower, L. P., & Alonso-M, A. (2001). Detrimental effects of latex and cardiac glycosides on survival and growth of first-instar

- monarch butterfly larvae Danaus plexippus feeding on the sandhill milkweed Asclepias humistrata. *Ecological Entomology*, *26*, 212–224. https://doi.org/10.1046/j.1365-2311.2001.00313.x
- Zas, R., Moreira, X., & Sampedro, L. (2011). Tolerance and induced resistance in a native and an exotic pine species: Relevant traits for invasion ecology. *Journal of Ecology*, 99, 1316–1326. https://doi.org/10.1111/j.1365-2745.2011.01872.x
- Zavala, J. A., Gog, L., & Giacometti, R. (2017). Anthropogenic increase in carbon dioxide modifies plant – insect interactions. *The Annals of Applied Biology*, 170, 68–77. https://doi.org/10.1111/aab.12319
- Zavala, J. A., Nabity, P. D., & DeLucia, E. H. (2013). An emerging understanding of mechanisms governing insect herbivory under elevated CO₂. Annual Review of Entomology, 58, 79–97. https://doi.org/10.1146/annurev-ento-120811-153544
- Zehnder, C. B., & Hunter, M. D. (2009). More is not necessarily better: The impact of limiting and excessive nutrients on herbivore population growth rates. *Ecological Entomology*, *34*, 535–543. https://doi.org/10.1111/j.1365-2311.2009.01101.x
- Zipkin, E. F., Ries, L., Reeves, R., Regetz, J., & Oberhauser, K. S. (2012). Tracking climate impacts on the migratory monarch butterfly. *Glob Chang Biol*, 18, 3039–3049. https://doi.org/10.1111/j.1365-2486.2012.02751.x
- Züst, T., & Agrawal, A. A. (2017). Trade-offs between plant growth and defense against insect herbivory: An emerging mechanistic synthesis. *Annual Review of Plant Biology*, *68*, 513–534. https://doi.org/10.1146/annurev-arplant-042916-040856
- Züst, T., Rasmann, S., & Agrawal, A. A. (2015). Growth-defense tradeoffs for two major anti-herbivore traits of the common milkweed *Asclepias syriaca*. *Oikos*, 124, 1404–1415. https://doi.org/10.1111/oik.02075

SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

How to cite this article: Decker LE, Hunter MD. Interspecific variation and elevated CO₂ influence the relationship between plant chemical resistance and regrowth tolerance. *Ecol Evol.* 2020;10:5416–5430. https://doi.org/10.1002/ece3.6284