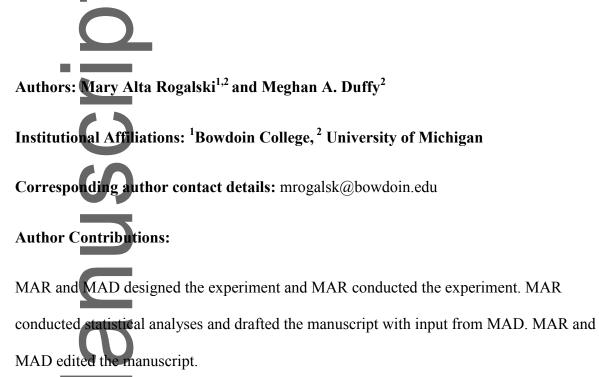
**Title:** Local adaptation of a parasite to solar radiation impacts disease transmission potential, spore yield, and host fecundity

Running fitle: Consequences of light adaptation in a parasite





## Acknowledgements:

We thank Becca Bilich, Haley Essington, Camden Gowler, Morgan Rondinelli, and especially Katie Hunsberger and Clara Shaw for much-appreciated assistance in setting up and maintaining the experiment. We thank Erin Overholt and Craig Williamson for providing light availability data on the study lakes, and we thank Spencer Hall for sharing infection prevalence data on Indiana lakes. Aaron Gilbreath provided valuable statistical consultation.

This is the author manuscript accepted for publication and has undergone full peer review but has not been the only the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the <u>Version of Record</u>. Please cite this article as <u>doi:</u> <u>10.1111/evo.13940</u>.

We thank two anonymous reviewers for providing valuable feedback on this manuscript. Funding was provided by NSF DEB-1305836 to MAD.

0

**Data accessibility**: Data are available on Dryad, doi:10.5061/dryad.2jm63xskd (Rogalski and Duffy 2020).

**Title:** Local adaptation of a parasite to solar radiation impacts disease transmission potential, spore yield, and host fecundity

# Abstract:

Environmentally ransmitted parasites spend time in the abiotic environment, where they are subjected to a variety of stressors. Learning how they face this challenge is essential if we are to understand how host-parasite interactions may vary across environmental gradients. We used a zooplankton-bacteria host-parasite system where availability of sunlight (solar radiation) influences disease dynamics to look for evidence of parasite local adaptation to sunlight exposure. We also examined how variation in sunlight tolerance among parasite strains impacted host reproduction. Parasite strains collected from clearer lakes (with greater sunlight penetration) were most tolerant of the negative impacts of sunlight exposure, suggesting local adaptation to sunlight conditions. This adaptation came with both a cost and a benefit for parasites: parasite strains from clearer lakes produced relatively fewer transmission stages (spores) but these strains were more infective. After experimental sunlight exposure, the most sunlight-tolerant parasite strains reduced host fecundity just as

2

much as spores that were never exposed to sunlight. Sunlight availability varies greatly among lakes around the world. Our results suggest that the selective pressure sunlight exposure exerts on parasites may impact both parasite and host fitness, potentially driving variation in disease epidemics and host population dynamics across sunlight availability gradients.

Key words: Solar radiation, disease, local adaptation, fitness trade-off, *Daphnia dentifera*, *Pasteuria ramosa* 

## Introduction

The abiotic environment can strongly impact host-parasite interactions and resulting disease outcomes (Wolinska and King 2009; Mostowy and Engelstädter 2011; Rogalski et al. 2017). For environmentally transmitted parasites, environmental stress may play a particularly important role: to infect a new host, transmission stages must spend time in the environment, where they are subject to abiotic stressors. Moreover, anthropogenic environmental change is altering the nature and intensity of stressors faced by environmentally transmitted parasites, in some cases quite rapidly (Tucker and Williamson 2011; Lal et al. 2013). Environmental stress can grive adaptation in populations, influencing both their physiology and their interactions with other organisms (Bijlsma and Loeschcke 2005), and the expectation is that parasites should be able to adapt rapidly to environmental change (Cable et al. 2017). However, free-living stages of parasites can be particularly vulnerable to environmental stressors and may be limited in their ability to adapt to changing conditions (Phillips et al. 2017). Furthermore, in the face of significant selection pressure exerted by hosts it may be difficult for parasites to adapt to the additional stress exerted by the environment.

Investigating the degree to which environmentally transmitted parasites can adapt to environmental stressors and how that influences their impacts on hosts may allow us to better predict how host-parasite interactions will change in a changing world.

One environmental stressor that can strongly influence parasite fitness is exposure to sunlight, particularly ultraviolet radiation (UVR) (in this manuscript sunlight is synonymous with solar radiation, inclusive of infrared, visible and ultraviolet radiation). Exposure to the shortest wavelengths of UVR that reach the Earth's surface, UV-B, promotes molecular and cellular damage (Rastogi et al. 2010) and can even be lethal to organisms (Vega and Pizarro 2000; Onzo et al. 2010). Navigating this stress likely shaped the ecology and evolution of life throughout its history (Hessen 2008). Parasites can be particularly sensitive to UVR – often more so than their hosts – such that sunlight exposure can dampen disease. For example, UVR can prevent coral bleaching by killing the UVR-sensitive *Vibrio shiloi* pathogen of corals in shallow, light exposed reefs (Fine et al. 2007). In addition, municipal water managers take advantage of sunlight sensitivity in microorganisms by treating drinking water with UVR to eliminate viral, bacterial, and protistan human pathogens (King et al. 2008).

Some organisms have responded to UVR stress by avoiding or tolerating exposure (Hansson and Hylander 2009; Rastogi et al. 2010; Rautio and Tartarotti 2010) or by repairing the damage caused by UVR after it occurs (Rastogi et al. 2010; Miner et al. 2015). However, these adaptations can come at a cost. For example, while pigmentation can protect organisms from UVR damage (Hessen 1996), it also makes them more susceptible to visual predators (Hansson 2000). In addition, production of photoprotective pigments is associated with metabolic and fitness costs (Hessen 1996). Such trade-offs may explain why UVR avoidance

and tolerance traits are more often found in populations experiencing greater exposure to sunlight (Stutzman 1999; Tartarotti et al. 2001; Miner et al. 2015).

Exposure to sunlight varies across space and over time. On larger spatial scales, high elevation as well as lower latitude locations experience relatively higher solar radiation intensity. Moreover, characteristics of the atmosphere, such as dust, smog, and cloud cover can affect the amount of sunlight that reaches the Earth's surface (Madronich et al. 1998). In aquatic systems, solar radiation intensity attenuates with depth, and the degree to which sunlight penetrates the water column depends in large part on differences in concentrations of dissolved organic carbon (DOC), which absorbs solar radiation (Morris et al. 1995; Rae et al. 2001). Land development and more frequent storms associated with climate change can drive lake browning, where increased levels of DOC can absorb a greater fraction of UVR (Williamson et al. 2016; Kritzberg 2017). UVR exposure also varies seasonally, with incident sunlight strength at its maximum in the summer. In addition, atmospheric ozone depletion drove greater UVR availability, though the success of the Montreal Protocol prevented largescale depletion of stratospheric ozone and associated dramatic increases in UVR (Williamson et al. 2014). Thus, both natural variation and human activities strongly influence light regimes, with some habitats receiving more sunlight, and some receiving less (Williamson et al. 2014).

Spatial and temporal variation in lake sunlight availability can affect the timing and size of disease epidemics of light-sensitive parasites (Overholt et al. 2012; Shaw 2019). Thus, we might expect the relative strength of the impacts of solar radiation on parasites and their interactions with hosts to vary among habitats. Here we use a zooplankton-bacterial hostparasite system to examine how variation in lake sunlight environments affects the fitness of

the light sensitive parasite as well as how the parasite affects its host. Our focal host is *Daphnia dentifera* Forbes (Hebert 1995), and our focal parasite is the environmentally transmitted, sterilizing, obligate killer, *Pasteuria ramosa* (Ebert et al. 1996). While *Daphnia* are affected by UVR exposure (Rautio and Tartarotti 2010), they appear to be modestly impacted at levels of exposure that can be lethal to their parasites (Overholt et al. 2012). Previous work found that *Pasteuria* spores are sensitive to ambient solar radiation, including PAR and UVR (Shaw 2019; Overholt et al. 2020) but did not consider variation among populations in sunlight sensitivity.

*Pasteuria* epidemics occur in lakes that span a gradient of sunlight availability, though these epidemics tend to be smaller in clearer lakes (Shaw 2019). We took advantage of this natural variation by conducting a common garden experiment to examine whether *Pasteuria* populations are locally adapted to lake sunlight regimes. After finding evidence that *Pasteuria* populations from clearer lakes are less affected by sunlight exposure than strains from darker takes, we examined whether this adaptation came at a cost to the parasites, in either their ability to infect *Daphnia* hosts or their reproduction within hosts post-infection. In addition, we examined how variation in parasite fitness in sunlight-exposed and unexposed conditions affected host fecundity.

### **Materials and Methods**

# Host and parasite

The host species in our study, *Daphnia dentifera* Forbes, is a dominant freshwater microcrustacean of lakes in the Midwestern United States (Hebert 1995). *Daphnia* are cyclically parthenogenic, which allows the maintenance of clonal lineages in the laboratory.

Daphnia play an important role in lake food webs as keystone herbivores of phytoplankton (Miner et al. 2012). Daphnia dentifera commonly suffer infection by our focal parasite, Pasteuria ramosa, a bacterial endoparasite in the Bacillus clade (Ebert et al. 1996). Daphnia encounter Pasteuria spores during filter feeding. Once consumed, Pasteuria spores bind to and then cross the host's gut epithelium and then replicate in the haemocoel (Ebert et al. 1996). Pasteuria infection results in host castration; while infected hosts may produce offspring early in the infection process and may also experience "castration relief" during late infection stages, many infected hosts never reproduce (Auld et al. 2012; Vale and Little 2012; Ben-Ami and Routtu 2013). In contrast, healthy mature *Daphnia* typically produce a clutch of offspring every three to four days when cultured at 20°C. Pasteuria spores become mature inside the host, beginning around 15--20 days and are released upon host death (Ebert et al. 2016). Mature spores are also infectious if the host suffers early mortality, for example via sloppy predation by Chaoborus (Auld et al. 2014a). Significant genetic diversity has been observed within *Pasteuria* populations, likely driven by substantial rates of recombination (Andras and Ebert 2013), and we have found significant phenotypic diversity of *Pasteuria* within and among lakes (Shaw 2019). Rapid change in key traits related to infectivity, spore production, and host genotype-specificity has been observed in *Pasteuria* populations both in the field and laboratory, over timespans as short as a few years to even a few months (Decaestecker et al. 2007; Auld et al. 2014a,b). In addition, Pasteuria spores can remain viable for years in the lake environment before being consumed by a new host (Decaestecker et al. 2004, 2007). These conditions set the stage for the possibility that *Pasteuria* may adapt in response to environmental stress.

This article is protected by copyright. All rights reserved.

7

#### Common garden experiment

#### Parasite spore collection and experimental sunlight exposure

We collected *Daphnia dentifera* individuals that were heavily infected with *Pasteuria ramosa* from six lakes in southeastern Michigan and one lake in Indiana between mid-August and mid-September 2017. The study lakes span a gradient of sunlight availability (Table S1) and represent a subset of a larger set of lakes that are regularly monitored for disease dynamics in the summer and fall (Hall et al. 2011; Duffy et al. 2015). *Daphnia* samples were collected by pooling three vertical tows of the entire water column in the deep basin of the lake, using a 153-µm-pore-size Wisconsin net. We isolated *Daphnia dentifera* individuals infected with *Pasteuria ramosa* by examining collected zooplankton under a dissecting microscope at 50 × magnification

We pooled groups of 20 *Pasteuria*-infected hosts from a given lake (hereafter a "strain") and stored them in 250  $\mu$ L filtered lake water in small plastic tubes in the dark at 4°C. We collected 1--5 *Pasteuria* strains (i.e. 20--100 *Daphnia* infected with *Pasteuria*, separated into 1--5 separate tubes) from each lake (Table S2). On September 21, 2017 we homogenized each strain by putverizing the infected hosts with a motorized pestle for two minutes. We then split the spore slurry for each individual strain evenly into two quartz vials (0.8 ml capacity). One vial was clear, allowing sunlight to penetrate ("sunlight-exposure treatment"). The second vial was covered in dark plastic to prevent any sunlight exposure ("control treatment"). The next morning, we suspended the clear and darkened vials of *Pasteuria* spores at a depth of 0.5 m below the surface of North Lake (the clearest study lake), leaving the vials suspended in the lake for four days.

8

Upon retrieving the vials, we transferred the *Pasteuria* spores from each vial into a small plastic tube, estimated the spore density for each strain x treatment using a Neubauer chamber cell (4 counts per replicate), and stored the spores in the dark at -20°C for six months prior to the experiment. *Pasteuria* spores retain infectivity for at least a year when stored at -20°C (Duffy and Hunsberger 2019).

# Laboratory infection assays

We conducted a common garden infection assay to 1) test our hypothesis that *Pasteuria* from clearer lakes would be relatively less impacted by exposure to sunlight, 2) determine whether *Pasteuria* populations from lakes spanning a sunlight availability gradient varied in key traits, and 3) examine how variation in *Pasteuria* transmission potential impacted *Daphnia* host reproduction.

Before beginning our common garden experiment, we maintained a single clonal genotype of *Daphnia dentifera* under standardized conditions for three generations to minimize variation in host condition (Text S1). This clone originated from one of our study lakes (Midland Lake) and was chosen based on pilot studies showing it to be a good genetic match with *Pasteuria* from most of the study lakes. We placed 48--72 h-old *Daphnia* individually into a well with 2 mL filtered lake water and 10,000 spores of a *Pasteuria* strain from either the sublight-exposure or control treatment. We exposed 10--12 host replicates per parasite strain x treatment (based on availability of spores), for a total of 332 *Daphnia* across 14 strains from 7 lakes.

We incubated the *Daphnia* for 48 hours and fed them  $2 \times 10^4$  cells of *Ankistrodesmus* per day. This is half the typical food density, which promotes uptake of parasite spores by the

host. The two-day incubations took place inside a darkened container, to prevent light exposure from impacting the *Pasteuria* spores before infection (Text S2). We then transferred each *Daphnia* to individual beakers filled with 35 mL of spore free filtered lake water. Hereafter, the *Daphnia* had no further opportunities for infection, as individuals were isolated in separate beakers, in spore-free water. We fed each *Daphnia*  $1 \times 10^6$  cells of *Ankistrodesmus* four times a week and changed the water twice weekly. At each water change we counted and removed any *Daphnia* offspring produced.

We ended the experiment 24 days after infection. This timing allowed the *Pasteuria* infections to progress to maturity, without killing many of the *Daphnia* hosts, ensuring the hosts' carapaces were intact and allowing for an accurate count of *Pasteuria* spore production. This approach mimics an environment in which a predator kills an infected host, which is likely given mortality rates in lakes and increased predation on infected hosts (Duffy and Hall 2008; Auld et al. 2014a). We assessed infection status by examining each *Daphnia* under a dissecting microscope at 50 × magnification. In cases where infection status was unclear, we examined *Daphnia* under a compound microscope at 400 × magnification. To quantify shore yields (number of spores produced per host), we placed each infected *Daphnia* in 100  $\mu$ L intered lake water in a plastic microcentrifuge tube, homogenized the host using a motorized pestie, and counted 4 replicate subsamples with a Neubauer counting chamber at 400 × magnification.

#### Measures of parasite transmission potential and host reproduction

Transmission potential is a component of parasite fitness that incorporates the likelihood of infection and the spore yield given infection. We calculated infectivity for each parasite strain from the sunlight-exposure and control treatments as the proportion of exposed *Daphnia* 

individuals that were infected by the end of the trial. Out of 332 replicates, 18 *Daphnia* that died when it was too early to determine infection status and five that were males were excluded from analysis. We calculated transmission potential by multiplying infectivity (number of infected hosts/ number of exposed hosts) by mean spore yield (number of spores/ infected host) for each parasite strain x treatment combination (Auld et al. 2014a). Thus, transmission potential gives the number of spores produced per exposed host individual. We estimated host fecundity by summing the total number of *Daphnia* offspring produced per individual by the end of the trial (through 24 days post infection).

There was a poor genetic match between *Pasteuria* strains from Walsh Lake and North Lake and the *Daphnia* clone used in the infection assays. No *Daphnia* were infected by *Pasteuria* from Walsh Lake and only a single *Daphnia* individual was infected from North Lake *Pasteuria*. These two lakes were excluded from further analysis. For comparison, infectivity in the control treatment for the other five lakes ranged from 50% to 92% (with 0--58% infectivity in the sunlight-exposure treatment).

### Statistical analyses

# Evidence of local adaptation to light in Pasteuria

All statistical analyses were conducted using the statistical program R v 3.5.2 (R Core Team 2016). We used general linear mixed modeling (GLMM) to evaluate the extent to which sunlight exposure in the common garden experiment impacted *Pasteuria* transmission potential and to see if this impact varied according to the UVR environment (absorption coefficient  $a_{d320}$ , an index of UVR availability) in the study lake from which the *Pasteuria* strains were collected (i.e. whether there was a genotype by environment interaction). We

included transmission potential as the response variable and  $a_{d320}$ , treatment (sunlightexposed or control), and their interaction as fixed effects. We included the random effect of parasite strain to account for lack of independence in the data for host replicates exposed to parasite strains from the same lake. We used Akaike's information criterion (AIC) to compare the saturated model with nested simpler models containing fewer fixed effects to evaluate the significance of the treatment x  $a_{d320}$  interaction (i.e. G x E) as well as the sunlight-exposure treatment (Table S3). We used the lme4 package to conduct GLMM analyses (Bates et al. 2015).

#### Parasite infectivity-spore yield trade-off

Parasite transmission potential depends on both infectivity (likelihood of infecting a host) and spore yield treplication within the host after infection). We used a mixed effects hurdle model to test whether either of these aspects of parasite fitness differed among *Pasteuria* strains collected from our study lakes that span a UVR availability gradient. The first part of the hurdle model included infection status of *Daphnia* individuals exposed to a given strain as the binomial response variable, and the second part evaluated spore yield of individual infected hosts as the response variable (with a zero-truncated negative binomial distribution). For both stages of the hurdle model we included  $a_{d320}$ , treatment, and their interaction as fixed effects. All models included parasite strain as a random effect. We evaluated whether infectivity or spore yield was explained by UVR penetration ( $a_{d320}$ ) in the home lake environment, the sunlight-exposure treatment, or their interaction by comparing AIC of the saturated model with nested simpler models containing fewer fixed effects (Table S4). We used the glmmTMB package (Brooks et al. 2017) to conduct the hurdle model analysis.

#### Impact of parasite transmission potential on host reproduction

Using a GLMM with the lme4 package, we evaluated how variation in parasite transmission potential influenced mean *Daphnia* host fecundity, which considers reproduction at the population level (mean total number of offspring/ replicate host for a given parasite strain x treatment). Mean host fecundity was the response variable, and transmission potential (mean spore yield weighted by infectivity of spores), treatment (sunlight-exposed or control), and their interaction were fixed effects. The models included parasite strain nested within lake as a random effect. We compared AIC values of the saturated model with simpler models including tewer predictors to evaluate the significance of each fixed effect (Table S5).

# Results

# Parasite local adaptation

Sunlight exposure strongly influenced *Pasteuria* transmission potential ( $\chi^2 = 17.56$ , p<0.001), causing on average a 70% decrease in spore yield per exposed host (an average decrease of 88,043 spores/exposed host ± 16,986 SE). Most notably, the impact of sunlight exposure on transmission potential differed among *Pasteuria* strains according to the UVR environment of the lake of origin (significant  $a_{d320}$  x treatment interaction:  $\chi^2 = 5.30$ , p = 0.021). Sunlight exposure most strongly decreased transmission potential of *Pasteuria* strains from darker

lakes (Figure 1).

*Pasteuria* strains from darker lakes (higher  $a_{d320}$ ) were less likely to infect *Daphnia* than strains from clearer lakes (Figure 2A;  $\chi^2 = 11.41$ , p<0.001). Infection was much less likely with *Pasteuria* from the sunlight-exposed treatment (Figure 2B,  $\chi^2 = 74.02$ , p<0.001), but there was not a significant interaction between  $a_{d320}$  and treatment ( $\chi^2 = 1.24$ , p = 0.27). Thus, the pattern of lower infectivity of *Pasteuria* strains from darker lakes was consistent across the sunlight-exposure and control treatments.

For those *Daphnia* that were successfully infected, spore yield was higher for *Pasteuria* strains from darker lakes (significant effect of  $a_{d320}$ ; Figure 2C-D;  $\chi^2 = 8.10$ , p = 0.004). There was no difference in spore yield according to sunlight-exposure treatment ( $\chi^2 = 0.69$ , p = 0.41) and no significant interaction between  $a_{d320}$  and treatment ( $\chi^2 = 0.32$ , p = 0.58).

# Effects of parasite transmission potential on mean host fecundity

The average fecundity of *Daphnia* replicates for a given strain x treatment combination declined with increasing *Pasteuria* transmission potential (significant effect of transmission potential on mean host offspring production:  $\chi^2 = 16.28$ , p<0.001). *Daphnia* replicates that were exposed to *Pasteuria* strains that had been exposed to sunlight produced more offspring, on average (significant effect of treatment:  $\chi^2 = 13.70$ , p<0.001; Figure 3), a result of a combination of lower infectivity (Figure 2) and resulting castration (Figure S1) of the sunlight-exposure treatment. In addition, there was a significant interaction between experimental treatment and parasite transmission potential (Figure 3;  $\chi^2 = 10.41$ , p = 0.001). We observed a negative relationship between *Pasteuria* transmission potential and mean

*Daphnia* fecundity for the sunlight-exposure treatment (Figure 3, separate GLMM for sunlight-exposure treatment data, with lake as a random effect:  $\chi^2 = 13.10$ , p<0.001), but there was no relationship between *Pasteuria* transmission potential and *Daphnia* fecundity for the control treatment data (separate GLMM for control treatment data, with lake as a random effect:  $\chi^2 = 0.49$ , p = 0.482).

# Discussion

Exposure to ambient solar radiation strongly impacted parasite transmission potential, with *Pasteuria* from darker lakes particularly sensitive to sunlight, suggesting local adaptation (Figure 1). This adaptation appears to be associated with both a cost and a benefit for the parasite. Pasteuria strains from clearer lakes produced fewer spores per infected host (Figure 2 C-D) but were more capable of causing infection, compared to strains from darker lakes (Figure 2A-B). These influences of solar radiation on parasite infectivity, spore yield, and resulting transmission potential were reflected in host reproduction. Daphnia populations that were exposed to parasites from the sunlight-exposure treatment had higher fecundity, on average, than the group of *Daphnia* exposed to parasites from the corresponding control group (Figure 3). However, parasite strains that had relatively high transmission potential following sunlight exposure more strongly reduced fecundity of their Daphnia hosts (Figure 3, sunlight-exposure treatment). Notably, parasite strains whose transmission potential was least impacted by sunlight exposure were as virulent to their hosts as all strains from the control treatment (Figure 3). Thus, adaptation to sunlight exposure in this virulent parasite effectively removed the fecundity benefit that host populations experienced when their sunlight-sensitive parasites were exposed to solar radiation.

In our experiment, exposure to ambient sunlight harmed *Pasteuria* spores, reducing their ability to infect their *Daphnia* host (Figure 2A-B). Our study was not designed to separate the effects of UVR from other wavelengths of light, such as photosynthetically active radiation (PAR). However, we know that *Pasteuria* is sensitive to both visible light and UVR (Overholt et al. 2020). Our incubation depth of 0.5 m in North Lake allowed exposure to about 14% of available UVR and 80% of PAR based on measures of  $a_{d320}$  and  $a_{d440}$ , respectively. Thus, it is possible that exposure to UVR, or PAR, or both drove the patterns observed in our experiment.

We observed evidence of a trade-off between *Pasteuria* infectivity and spore yield. *Daphnia* were more often infected by *Pasteuria* strains from clearer lakes (Figure 2A-B), but these parasite strains from clearer lakes produced fewer spores per infected host (Figure 2C-D). Sunlight exposure greatly diminished infectivity, regardless of the lake of origin of the spores. However, for those hosts that became infected, spore yields were unaffected by the sunlight-exposure treatment. Together these infectivity and spore yield gradients led to big differences in how transmission potential for parasite strains from clearer and darker lakes were impacted by sunlight exposure. High spore yields for parasite strains from darker lakes contributed to high transmission potential in the control treatment; however, sunlight exposure greatly diminished transmission potential for these strains from darker lakes, relative to strains from clearer lakes (Figure 1). This evidence of a trade-off between infectivity and spore yield supports the idea that resistance to sunlight exposure is metabolically costly. Other researchers have seen reduced metabolic function in organisms with increased ability to tolerate or avoid UVR damage (Raven 1991; Hessen 1996). Since spore production is known to be limited by resource availability (Hall et al. 2009; Cressler et

al. 2014), our results suggest that the ability to withstand sunlight exposure may draw resources away from *Pasteuria* reproduction.

It is unclear why parasite strains from clearer lakes would be better able to infect *Daphnia* hosts when the spores were incubated in the dark (control treatment). One possibility is that, despite taking care to avoid sunlight exposure outside of the experimental treatment, Pasteuria spores were exposed to solar radiation during the process of collecting infected hosts and processing the spores. This modest exposure to sunlight may have resulted in some spore mortality during experimental set up, with relatively lower rates of mortality among sunlight-adapted strains from the clearer lakes. However, it is also possible that *Pasteuria* UVR defenses increase the chance of infecting a host. Pasteuria infection of Daphnia is a complex yet well-studied process involving several key stages, including spore activation upon contact with the host, which involves shedding the outer spore shell, as well as attaching to and penetrating the host gut (Ebert et al. 2016). One possibility is that if sunlighttolerant *Pasteuria* protect themselves with an altered spore shell, perhaps this also increases the odds of shedding this outer shell, leading to an increase in activation rate. Another possibility is that UVR chemical or physical protective barriers aid in gut attachment and penetration or increase resistance to immune system attacks. Further research investigating Pasteuria phenotypic responses to sunlight exposure could provide insight into the mechanisms driving both decreased spore yields and increased infectivity in sunlight-adapted parasite strains.

We predicted that sunlight exposure would drive variation in *Pasteuria* transmission potential, which would then influence mean host reproduction. Our results supported this hypothesis, but also suggest that the effect of *Pasteuria* transmission potential on host

<sup>17</sup> 

reproduction saturates. Reduced *Pasteuria* transmission potential due to reduced infectivity in the sunlight-exposure treatment provided relief to the *Daphnia* host, allowing fewer animals to be infected and castrated, yielding an increase in mean host fecundity (Figures S1, 2 and 3). However, *Pasteuria* fitness levels beyond ~75,000 spores/exposed host did not yield additional decreases in host fecundity. Thus, while there was great variation in baseline *Pasteuria* transmission potential among strains that had been maintained in the dark, exposure to these spores had an equally negative impact on host fecundity (Figure 3, control treatment). Notably, some of the most sunlight-tolerant *Pasteuria* strains that were exposed to ambient sunlight reduced host reproduction about as much as strains that were unexposed to solar radiation. Thus, it appears that local adaptation to sunlight exposure to solar radiation.

How might parasite adaptation and the associated costs and benefits, influence disease dynamics in the field? While sunlight exposure and intensity may limit the size, duration, and timing of *Pasteuria* epidemics in the field (as seen in Shaw 2019), our results suggest that adaptation to solar radiation could contribute to higher rates of disease transmission in clearer lakes than we might expect otherwise. While sunlight exposure may benefit *Daphnia* by reducing infection by this castrating obligate killer, increased sunlight tolerance in the parasite may remove some of this benefit.

Sunlight penetration is changing in lakes around the world owing to impacts associated with climate change and land use (Williamson et al. 2014, 2016). We might expect to see earlier onset of disease, with larger, longer lasting epidemics in these darker lakes. Without the need to invest in costly UVR protection mechanisms, *Pasteuria* in these darker lakes may also produce more spores per infected host, further boosting transmission potential.

Alternatively, since *Pasteuria* strains from darker lakes appear to be less infectious, it is possible that lake browning could lead to no net change in transmission. Longer term studies could help elucidate connections between changing sunlight availability and shifts in disease dynamics.

Identifying evolutionary responses to environmental stress and the impacts that local adaptation might have on species interactions are key challenges for evolutionary ecology. Solar radiation is a widespread environmental stressor known to strongly affect many organisms. Our finding that adaptation to solar radiation is linked with changes in key parasite traits as well as host-parasite interactions highlights the importance of considering evolutionary interactions beyond host-parasite coevolution in trying to untangle the drivers of disease. Further research into the physiological and genetic mechanisms associated with solar radiation adaptation will help us better understand why we see impacts on parasite spore yields and infectivity. Field research and experiments can help us uncover the consequences for host populations as well as disease dynamics and any cascading effects on ecosystem function. As human activities alter the availability of solar radiation in lake environments globally, this work will aid in predicting how these changes in sunlight exposure might influence disease

### Literature Cited

<BIBL>

Andras, J. P., and D. Ebert. 2013. A novel approach to parasite population genetics:
Experimental infection reveals geographic differentiation, recombination and hostmediated population structure in *Pasteuria ramosa*, a bacterial parasite of *Daphnia*. Mol.
Ecol. 22:972–986.

- Auld, S. K. J. R., S. R. Hall, and M. A. Duffy. 2012. Epidemiology of a *Daphnia*multiparasite system and its implications for the Red Queen. PLoS One 7:1–6.
- Auld, S. K. J. R., S. R. Hall, J. H. Ochs, M. Sebastian, and M. A. Duffy. 2014a. Predators and patterns of within-host growth can mediate both among-host competition and evolution of transmission potential of parasites. Am. Nat. 184:77–90.
- Auld, S. K.J. R. P. J. Wilson, and T. J. Little. 2014b. Rapid change in parasite infection traits over the course of an epidemic in a wild host-parasite population. Oikos 123:232–238.
- Bates, D., M. Maechler, B. M. Bolker, and S. Walker. 2015. Fitting linear mixed-effects models using {lme4}. J. Stat. Softw. 67:1–48.
- Ben-Ami, F., and J. Routtu. 2013. The expression and evolution of virulence in multiple infections: The role of specificity, relative virulence and relative dose. BMC Evol. Biol. 13:1-11.
- Bijlsma, R., and V. Loeschcke. 2005. Environmental stress, adaptation and evolution: An overview. J. Evol. Biol. 18:744–749.
- Brooks, M. E. J. K. K., K. van Benthem, A. Magnusson, C. W. Berg, A. Nielsen, H. J. Skaug,
  M. Maechler, and B. M. Bolker. 2017. glmmTMB balances speed and flexibility among packages for zero-inflated generalized linear mixed modeling. R J. 9:378–400.
- Cable, J., I. Barber, B. Boag, A. R. Ellison, E. R. Morgan, K. Murray, E. L. Pascoe, S. M.Sait, A. J. Wilson, and M. Booth. 2017. Global change, parasite transmission and disease control: lessons from ecology. Philos. Trans. R. Soc. B Biol. Sci. 372:20160088.

- Cressler, C. E., W. A. Nelson, T. Day, and E. McCauley. 2014. Starvation reveals the cause of infection-induced castration and gigantism. Proc. R. Soc. B Biol. Sci. 281.
- Decaestecker, E., S. Gaba, J. A. M. Raeymaekers, R. Stoks, L. Van Kerckhoven, D. Ebert, and L. De Meester. 2007. Host-parasite "Red Queen" dynamics archived in pond sediment. Nature 450:870–873.
- Decaestecker, E., C. Lefever, L. De Meester, and D. Ebert. 2004. Haunted by the past: Evidence for dormant stage banks of microparasites and epibionts of *Daphnia*. Limnol. Oceanogr. 49:1355–1364.
- Duffy, M. A., and S. R. Hall. 2008. Selective predation and rapid evolution can jointly dampen effects of virulent parasites on *Daphnia* populations. Am. Nat. 171:499–510.
- Duffy, M. A., and K. K. Hunsberger. 2019. Infectivity is influenced by parasite spore age and exposure to freezing: Do shallow waters provide *Daphnia* a refuge from some parasites?J. Plankton Res. 41:12–16.
- Duffy, M. A., T. Y. James, and A. Longworth. 2015. Ecology, virulence, and phylogeny of *Blastulidium paedophthorum*, a widespread brood parasite of *Daphnia* spp. Appl.
   Environ. Microbiol. 81:5486–5496.
- Ebert, D., D. Duneau, M. D. Hall, P. Luijckx, J. P. Andras, L. Du Pasquier, and F. Ben-Ami.
  2016. A population biology perspective on the stepwise infection process of the bacterial pathogen *Pasteuria ramosa* in *Daphnia*. Elsevier Ltd.
- Ebert, D., P. Rainey, T. M. Embley, and D. Scholz. 1996. Development, life cycle, ultrastructure and phylogenetic position of Pasteuria ramosa Metchnikoff 1888:

Rediscovery of an obligate endoparasite of *Daphnia magna* Straus. Philos. Trans. R. Soc. B 351:1689–1701.

- Fine, M., T. Bann Israely, E. Rosenberg, and Y. Loya. 2007. Ultraviolet radiation prevents bleaching in the Mediterranean coral *Oculina patagonica*. Mar. Ecol. Prog. Ser. 226:249–254.
- Hall, S. R., C. R. Becker, M. A. Duffy, and C. E. Ca. 2011. Epidemic size determines population-level effects of fungal parasites on *Daphnia* hosts. Oecologia 166:833–842.
- Hall, S. R., J. L. Simonis, R. M. Nisbet, A. J. Tessier, and C. E. Cáceres. 2009. Resource ecology of virulence in a planktonic host- parasite system: An explanation using dynamic energy budgets. Am. Nat. 174:149–162.
- Hansson, L.-A. 2000. Induced pigmentation in aooplankton : A trade-off between threats from predation and ultraviolet radiation. Society 267:2327–2331.
- Hansson, L. A., and S. Hylander. 2009. Effects of ultraviolet radiation on pigmentation,photoenzymatic repair, behavior, and community ecology of zooplankton. Photochem.Photobiol. Sci. 8:1266–1275.
- Hebert, P. D. N. 1995. The *Daphnia* of North America: An illustrated fauna. CyberNatural Software, University of Guelph.
- Hessen, D. O. 1996. Competitive trade-off strategies in Arctic *Daphnia* linked to melanism and UV-B stress. Polar Biol. 16:573–579.
- Hessen, D. O. 2008. Solar radiation and the evolution of life. Pp. 123–136 *in* E. Bjertness, ed. Solar Radiation and Human Health. The Norwegian Academy of Science and Leters, 22

Oslo.

- King, B. J., D. Hoefel, D. P. Daminato, S. Fanok, and P. T. Monis. 2008. Solar UV reduces
   *Cryptosporidium parvum* oocyst infectivity in environmental waters. J. Appl. Microbiol. 104:1311–1323.
- Kritzberg, E. S. 2017. Centennial-long trends of lake browning show major effect of afforestation. Limnol. Oceanogr. Lett. 2:105–112.
- Lal, A., M. G. Baker, S. Hales, and N. P. French. 2013. Potential effects of global environmental changes on cryptosporidiosis and giardiasis transmission. Trends Parasitol. 29:83–90. Elsevier Ltd.
- Madronich, S., R. L. Mckenzie, and M. M. Caldwell. 1998. Changes in biologically active ultraviolet radiation reaching the Earth's surface. J. Photochem. Photobiol. B Biol. 46:5–

# 19.

- Miner, B. E., P. M. Kulling, K. D. Beer, and B. Kerr. 2015. Divergence in DNA photorepair efficiency among genotypes from contrasting UV radiation environments in nature. Mol. Ecol. 24:6177–6187.
- Miner, B. E., L. De Meester, M. E. Pfrender, W. Lampert, and N. G. H. Jr. 2012. Linking genes to communities and ecosystems: *Daphnia* as an ecogenomic model. Proc. R. Soc. B Biol. Sci. 279:1873–1882.
- Morris, D. P., H. Zagarese, C. E. Williamson, E. G. Balseiro, B. R. Hargreaves, B.Modenutti, R. Moeller, and C. Queimalinos. 1995. The attenuation of solar UV radiation in lakes and the role of dissolved organic carbon. Limnol. Oceanogr. 40:1381–1391.

- Mostowy, R., and J. Engelstädter. 2011. The impact of environmental change on host-parasite coevolutionary dynamics. Proc. R. Soc. B Biol. Sci. 278:2283–2292.
- Onzo, A., M. W. Sabelis, and R. Hanna. 2010. Effects of ultraviolet radiation on predatory mites and the role of refuges in plant structures. Environ. Entomol. 39:695–701.
- Overholt, E. P., M. A. Duffy, M. P. Meeks, T. H. Leach, and C. E. Williamson. 2020. Light exposure decreases infectivity of the *Daphnia* parasite *Pasteuria ramosa*. J. Plankton Res. 00:1-4.
- Overholt, E. P., S. R. Hall, C. E. Williamson, C. K. Meikle, M. A. Duffy, and C. E. Cáceres. 2012. Solar radiation decreases parasitism in *Daphnia*. Ecol. Lett. 15:47–54.
- Phillips, A. J., C. J. Carlson, E. R. Dougherty, N. C. Harris, K. R. Burgio, C. F. Clements, and C. A. Cizauskas. 2017. Parasite vulnerability to climate change: an evidence-based functional trait approach. R. Soc. Open Sci. 4:160535.
- R Core Team. 2016. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.
- Rae, R., C Howard-Williams, I. Hawes, A. M. Schwarz, and W. F. Vincent. 2001.
  Penetration of solar ultraviolet radiation into New Zealand lakes: Influence of dissolved organic carbon and catchment vegetation. Limnology 2:79–89.
- Rastogi, R. P., Richa, A. Kumar, M. B. Tyagi, and R. P. Sinha. 2010. Molecular mechanisms of ultraviolet radiation-induced DNA damage and repair. J. Nucleic Acids 2010:1–32.
- Rautio, M., and B. Tartarotti. 2010. UV radiation and freshwater zooplankton: Damage, protection and recovery. Freshw. Rev. 3:105–131.

<sup>24</sup> 

- Raven, J. A. 1991. Responses of aquatic photosynthetic organisms to increased solar UVB. J.Photochem. Photobiol. B Biol. 9:239–244.
- Rogalski, M. A., and M. A. Duffy. 2020. Data from "Local adaptation of a parasite to solar radiation impacts disease transmission potential, spore yield, and host fecundity". Dryad, https://doi.org/10.5061/dryad.2jm63xskd.
- Rogalski, M. A., C. D. Gowler, C. L. Shaw, R. A. Hufbauer, and M. A. Duffy. 2017. Human drivers of ecological and evolutionary dynamics in emerging and disappearing infectious disease systems. Philos. Trans. R. Soc. B Biol. Sci. 372:1–9.
- Shaw, C. L. 2019. Drivers of epidemic timing and size in a natural aquatic system. University of Michigan.
- Stutzman, P. 1999. A comparative study of ultraviolet radiation tolerance in different populations of Diaptomus minutus. J. Plankton Res. 21:387–400.
- Tartarotti, B., I. Laurion, and R. Sommaruga. 2001. Large variability in the concentration of mycosporine-like amino acids among zooplankton from lakes located across an altitude gradient. Limnol. Oceanogr. 46:1546–1552.
- Tucker, A. J., and C. E. Williamson. 2011. Lakes in a new light: Indirect effects of ultraviolet radiation. Freshw. Rev. 4:115–134.
- Vale, P. F., and T. J. Little. 2012. Fecundity compensation and tolerance to a sterilizing pathogen in *Daphnia*. J. Evol. Biol. 25:1888–1896.
- Vega, M. P., and R. Pizarro. 2000. Lethal effect induced by ultraviolet-b in a planktonic copepod: Role of the post-irradiation time on mortality measurements. J. Freshw. Ecol. 25

15:1–5.

Williamson, C. E. E. P. Overholt, J. A. Brentrup, R. M. Pilla, T. H. Leach, S. G. Schladow, J. D. Warren, S. S. Urmy, S. Sadro, S. Chandra, and P. J. Neale. 2016. Sentinel responses to droughts, wildfires, and floods: Effects of UV radiation on lakes and their ecosystem services. Front. Ecol. Environ. 14:102–109.

Williamson, C. F., R. G. Zepp, R. M. Lucas, S. Madronich, A. T. Austin, C. L. Ballaré, M. Norval, B. Sulzberger, A. F. Bais, R. L. McKenzie, N. D. Paul, D.-P. H\u00e4der, and J. F. Bornman. 2014. Solar ultraviolet radiation in a changing climate. Nat. Clim. Chang. 4:434–441.

Wolinska, J., and K. C. King. 2009. Environment can alter selection in host-parasite interactions. Trends Parasitol. 25:236–244.



Author

**Figure 1.** Sunlight exposure reduced parasite transmission potential, particularly for strains that originated from darker lakes. Lower  $a_{d320}$  values are relatively clearer lakes, with greater ultraviolet radiation (UVR) penetration. Each point represents the transmission potential for a parasite strain x treatment combination from a given lake (1-5 strain replicates per lake x treatment combination). Black circles represent the control treatment and gray triangles represent the sunlight exposed treatment. Data is slightly jittered along the x axis to show overlapping data points.

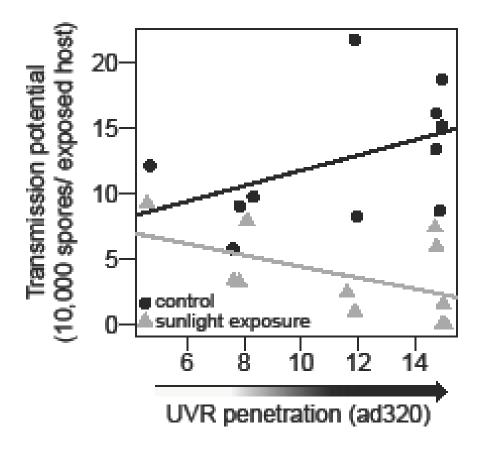
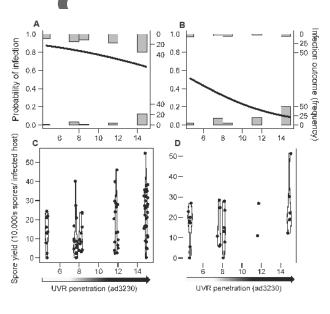


Figure 2. Pasteuria strains from darker lakes were less infectious but yielded more spores. Panels A and C include data from the control treatment; B and D are from the sunlightexposure treatment. Lower  $a_{d320}$  values are relatively clearer lakes with greater ultraviolet radiation (U (R). A-B) The black line (left axis) shows the probability of infection: in the control treatment, approximately 90.1% of individuals became infected when exposed to spores from the clearest lake, versus approximately 62.5% from the darkest lake. In the sunlight-exposure treatment, approximately 58.3% of individuals became infected when exposed to spores from the clearest lake, versus 10.9% from the darkest lake. The bars (right axis) show binned data of frequency of infected (top) and uninfected (bottom) individuals for different  $a_{d_{220}}$  values. Each bar combines all *Daphnia* individuals from the parasite strain replicates for a given lake that were either uninfected (bottom) or infected (top); because there were different numbers of parasite strain replicates for different lakes, the total frequencies differ across lakes, with lakes with more replicates having larger total frequencies. C-D) Spore yield from infected hosts (uninfected hosts not included in analysis/plot).



28

**Figure 3.** Mean host fecundity decreased with increasing parasite transmission potential (spore yield weighted by infectivity) in the sunlight-exposure treatment (gray) but not the control treatment (black). Each data point represents the mean host fecundity and transmission potential for a particular parasite strain x treatment combination.

