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Influence of bioenergy crop production and climate change on ecosystem services

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Abstract: Land use change can significantly affect the provision of ecosystem services and the effects could be exacerbated by projected climate change. We quantify ecosystem services of bioenergy-based land use change and estimate the potential changes of ecosystem services due to climate change projections. We considered 17 bioenergy based scenarios with Miscanthus, switchgrass, and corn stover as candidate bioenergy feedstock. Soil and Water Assessment Tool simulations of biomass/grain yield, hydrology, and water quality were used to quantify ecosystem services fresh water provision (FWPI), food (FPI) and fuel provision, erosion regulation (ERI), and flood regulation (FRI). Nine climate projections from Coupled Model Intercomparison Project phase-3 were used to quantify the potential climate change variability. Overall, ecosystem services of heavily row cropped Wildcat creek watershed were lower than St. Joseph River watershed which had more forested and perennial pasture lands. The provision of

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30 ecosystem services for both study watersheds were improved with bioenergy production
31 scenarios. Miscanthus in marginal lands of Wildcat creek (9% of total area) increased FWPI by
32 27% and ERI by 14% and decreased FPI by 12% from the baseline. For St. Joseph watershed,
33 Miscanthus in marginal lands (18% of total area) improved FWPI by 87% and ERI by 23% while
34 decreasing FPI by 46%. The relative impacts of land use change were considerably larger than
35 climate change impacts in this study. *Editor's note: This paper is part of the featured series on*
36 *SWAT Applications for Emerging Hydrologic and Water Quality Challenges. See the February*
37 *2017 issue for the introduction and background to the series.*

38
39 **(Key Terms:** ecosystem services, bioenergy impacts, Miscanthus, switchgrass, corn stover, fresh
40 water provision, food and fuel provision, SWAT.)

41 **Introduction**

42 Climate variability and change, land use, and land management change can put increasing
43 pressure on our natural resources, especially water, land, and food resources. The increasing
44 global emphasis on bioenergy production can introduce fast growing high biomass yielding
45 perennial grasses and trees to commercial agriculture under favorable economic conditions.
46 Anticipated climate change and variability can exacerbate the potential impacts of land use
47 changes. In general, the inclusion of perennial grasses instead of conventionally managed row
48 cropping is expected to have environmental benefits. Numerous studies have highlighted water
49 quality benefits of perennial grass production (Self-Davis *et al.*, 2003; Parrish and Fike, 2005;
50 McIsaac *et al.*, 2010; Diaz-Chavez *et al.*, 2011; Feng *et al.*, 2015; Cibin *et al.*, 2016) while a crop
51 residue based bioenergy production is predicted to increase soil erosion and sediment loadings to
52 the receiving streams (Delgado, 2010; van Donk *et al.*, 2010; Johnson *et al.*, 2010; Cibin *et al.*,
53 2012). Conventional agriculture bioenergy crop production has also raised concern over food and
54 fuel competition (Pimentel *et al.*, 2009) and carbon sequestration (Franko *et al.*, 2015). A careful
55 environmental impact assessment from different stakeholder perspectives and optimum planning
56 of different land use and management practices are required for inclusion of energy crops into
57 commercial agriculture.

58 Environmental impacts of bioenergy production can be assessed in different ways, such
59 as (1) sustainability indicator analysis (McBride *et al.*, 2011), (2) risk-vulnerability-reliability

60 assessment (Hoque *et al.*, 2013), and (3) absolute and percentage change impact assessment with
61 baseline reference (Cibin *et al.*, 2016; Feng *et al.*, 2015), etc. Ecosystem services evaluation for
62 bioenergy production is another metric which can compare different aspects of ecosystem
63 benefits to people from bioenergy production. Ecosystem services can be classified under
64 supporting (nutrient cycling, soil formation, primary production, etc.), provisioning (food, fresh
65 water, wood and fiber, fuel, etc.), regulating (climate regulation, flood regulation, erosion
66 regulation, disease regulation, etc.) and cultural (aesthetic, spiritual, educational, recreational,
67 etc.) services (MEA, 2005). In an ideal sustainably managed ecosystem there should be a good
68 balance between all ecosystem services (Foley *et al.*, 2005). However, in current intensive
69 agricultural ecosystems, the primary focus is on maximizing food provisioning while other
70 ecosystem services are often not prioritized (Foley *et al.*, 2005).

71 Quantifying ecosystem services of a system is challenging due to lack of quantifying
72 methods and data availability (Seppelt *et al.*, 2011; Logsdon and Chaubey, 2012, Volk, 2013). In
73 addition, not all the valuation metrics perform equally well for the different ecosystem services
74 (Farber *et al.*, 2006), and there is a lack of standardization in how they are valued (Boithias *et al.*,
75 2016; Polasky *et al.*, 2015). To the best of our knowledge, there are no reported efforts on
76 quantification of ecosystem services for bioenergy production systems. In the case of bioenergy
77 systems, measured data at large scale are not available for the United States since large
78 bioenergy production is not yet established. Application of mathematical simulation models can
79 give realistic scenario realizations if bioenergy production scenarios are adequately represented
80 in the models. Logsdon and Chaubey (2013) proposed a methodology to quantify fresh water
81 provisioning, food provisioning, fuel provisioning, flood regulation and erosion regulation using
82 Soil and Water Assessment Tool (SWAT) ecohydrological model (Arnold *et al.*, 1998)
83 simulations, which is part of a broader recent thrust identified in the literature to fully utilize the
84 capabilities of SWAT to model ecosystem services (Francesconi *et al.*, 2016).

85 Two key sets of improved bioenergy cropping system related algorithms have recently
86 been incorporated in SWAT: (1) Cibin (2013) introduced corrections to SWAT that enable more
87 accurate representation of corn stover removal, and (2) Trybula *et al.* (2015) parameterized and
88 improved SWAT to better physiologically represent perennial bioenergy crops such as
89 *Miscanthus* and switchgrass in the Midwest US. These improved SWAT bioenergy algorithms
90 were ported to SWAT version 2012 (SWAT2012), Revision 611 and have only been applied in

91 limited recent SWAT applications including Chen *et al.* (2015), Cibin *et al.* (2016), Gassman *et*
92 *al.* (2016) and Panagopoulos *et al.* (2016). These SWAT improvements, coupled with the
93 methodology advanced by Logsdon and Chaubey (2013), support the overall goal of this study to
94 quantify ecosystem services of futuristic bioenergy production scenarios.

95 Climate change and variability along with land use changes can affect ecosystem services
96 in a watershed. Quantifying of ecosystem service of futuristic climate and land use change can
97 help in developing better adaptation strategies to overcome potential negative impacts and
98 maximize benefits. This research is part of a four-study series discussing policy implications
99 (Kling *et al.*, 2016) and environmental impacts at different spatial scales (Gassman *et al.*, 2016;
100 Panagopoulos *et al.*, 2016), within the broader context of research being conducted within
101 CenUSA Bioenergy (Moore *et al.*, 2014), and quantifies the potential impacts of futuristic
102 bioenergy production scenarios on ecosystem services under current and projected future climate
103 scenarios in the U.S. Corn Belt region. The specific objectives of the study are to: (1) quantify
104 the ecosystem services of plausible bioenergy production scenarios for two watersheds in the
105 Midwest US, and (2) estimate the variability in the provision of ecosystem services due to
106 climate change.

107 **Methodology**

108 Seventeen plausible futuristic bioenergy-based scenarios were developed with
109 *Miscanthus* (*Miscanthus* × *giganteus*) and Shawnee, an upland switchgrass (*Panicum virgatum*
110 *L.*) variety as perennial dedicated bioenergy crops and corn (*Zea mays L.*) stover as crop residue
111 for biofuel production. The scenarios were developed with bioenergy crop production from
112 marginal lands, current pasture lands and prime agricultural lands using SWAT version 2012
113 (SWAT2012), Revision 615, which is the same version used in the other CenUSA Bioenergy
114 project applications (Gassman *et al.*, 2016; Panagopoulos *et al.*, 2016). SWAT was used to
115 represent the scenarios in the study watersheds. The model was parameterized and improved for
116 better physical representation of perennial grasses, specifically upland switchgrass and
117 *Miscanthus* using field measurements in the region by Trybula *et al.* (2015), and for
118 representation of corn stover removal using the methods described by Cibin (2013). The
119 improved model was used in this study to estimate streamflow, erosion, nutrient loading, and
120 crop yield for different bioenergy scenarios to quantify ecosystem services from these scenarios.

121 Five ecosystem services were evaluated including fresh water provision (FWPI), food provision
122 (FPI), fuel provision (FuPI), erosion regulation (ERI), and flood regulation (FRI). Daily and
123 annual time step SWAT simulations were used to calculate the biophysical value of ecosystem
124 services. Simulations were conducted in a parallel computing framework on Linux computer
125 clusters maintained by Research Computing (RCAC), a high-performance computing facility
126 located at Purdue University (Purdue, 2016).

127 Study Area

128 The study was conducted in two agriculturally dominant watersheds located primarily in
129 Indiana in the eastern Corn Belt (Figure 1): (1) Wildcat Creek, which drains 2045 km² of
130 predominantly agricultural land characterized by 70% corn/soybean production, 9% forest, and
131 5% pasture, and (2) St. Joseph River, with a 2800 km² drainage area that is also predominantly
132 agricultural but with lower corn/soybean production (37%), 25% pasture, 12% forest and 8%
133 forested wetlands. The Wildcat Creek watershed is located in central Indiana with flat terrain,
134 highly productive corn/soybean areas and drains to the Wabash River and eventually to the Gulf
135 of Mexico. The St. Joseph River watershed is located in northern Indiana with drainage areas in
136 Indiana, Michigan and Ohio. The watershed has a hilly terrain with marginal agricultural lands
137 and pasture areas. The St. Joseph River drains to Lake Erie which has experienced excess algal
138 blooms and eutrophication in recent years (Obenour *et al.*, 2014; Scavia *et al.*, 2016). Soils of
139 both watersheds were formed from compacted glacial till and are classified as good agricultural
140 soils. The dominant soil textures are silt loam, silty clay loam, and clay loam. Annual
141 precipitation of both watersheds were near 1000 mm with Wildcat Creek watershed receiving
142 slightly more precipitation than St Joseph River watershed. The two watersheds represent
143 distinctive terrain, land use and management characteristics of the Midwest US watersheds.
144 SWAT was parameterized for the two watersheds with detailed spatial representation and was
145 calibrated and validated for crop yield, streamflow and water quality (Cibin *et al.*, 2016). Daily
146 stream flow Nash-Sutcliffe efficiency of calibrated model was above 0.65 for all stream gauge
147 stations in both watersheds under calibration and validation periods. We encourage readers to
148 refer to the supplementary information of Cibin *et al.* (2016) to find detailed discussion on model
149 development, calibration and validation. Hydrology and water quality impacts of growing
150 bioenergy crops in these two watersheds were studied using SWAT and results indicated
151 improved water health with the introduction of perennial bioenergy crops in the watershed (Cibin

152 *et al.*, 2016). Two sets of climate data were used in this study: (1) measured precipitation and
153 temperature data from the National Climatic Data Center (<http://www.ncdc.noaa.gov/>) for 14
154 years (1996-2009), and (2) climate projection data representing future climate change scenarios
155 discussed below. Solar radiation, wind speed and relative humidity were generated by SWAT
156 using long term mean monthly data for the region.

157 **Climate projection data**

158 The climate projection based analyses were done using precipitation and temperature data
159 from nine General Circulation Climate (GCM) model projections obtained from the World
160 Climate Research Programme's (WCRP's) Coupled Model Intercomparison Project phase 3
161 (CMIP3) (WRCP, 2016). These nine projections consisted of: 3 GCM models (GFDL CM2.0.1,
162 UKMO HadCM3 3.1 and NCAR PCM 1.3) in combination with each of the three future
163 emission scenarios (A1B, A2, B1). Further description of these GCMs and projections are
164 provided by WRCP (2016). The three models used in this study were identified to be
165 representative of previous impact assessments in the region (e.g., Kling *et al.* 2003). Precipitation
166 and temperature data were bias corrected and statistically downscaled with a resolution of 1/8°.
167 Bias correction was done using empirical statistical technique mapping the probability density
168 functions for monthly precipitation and temperature from climate model projections and
169 measured data. Bias corrected monthly precipitation and temperature was disintegrated to daily
170 by selecting daily time series from the monthly historic climatology. A detailed discussion of
171 downscaling and bias correction methodology can be obtained from Cherkauer and Sinha (2010).
172 Climate data from 1950-2050 was used for the analysis considering the first 10 years for model
173 warm up, and three thirty-year periods representing past (1960-1989), present (1990-2019) and
174 future (2020-2049) climate conditions. The gridded precipitation and temperature data were
175 input to SWAT. In general, average monthly temperature and precipitation increased with all
176 three emission scenarios for the future time period (2020-2049) compared to past (1960-1989)
177 (Table S1).

178 **Bioenergy scenarios**

179 Plausible futuristic bioenergy scenarios for the region were carefully defined with
180 marginal lands as potential areas for growing bioenergy crops as well as the prime agricultural
181 land conversion as somewhat extreme case scenarios when bioenergy production becomes more

182 economical. The stover residue from corn after grain harvest is identified as immediate
183 biofeedstock for biofuel production in the Corn Belt region due to wide availability (Lal, 2004;
184 Wilhelm *et al.*, 2004). The only stover removal scenario considered included 50% stover
185 removal from low slope (<2% slope) areas. Table 1 provides details of the seventeen scenarios
186 considered in this study and area distribution in the two watersheds under each scenario.
187 Marginal lands are proposed as viable first choice areas for growing bioenergy crops due to less
188 competition between food and fuel production and potential environmental benefits (Robertson
189 *et al.*, 2008; Gopalakrishnan *et al.*, 2009; Cai *et al.*, 2010). Agricultural marginal lands in the
190 study watersheds were identified with three conditions: (1) highly erodible areas; or (2)
191 agricultural low productive marginal lands; or (3) land capability based marginal lands. These
192 criteria are generally consistent with the definition of marginal lands used within CenUSA
193 Bioenergy (Moore *et al.*, 2014). Corn/soybean areas with $\geq 2\%$ slope were considered as
194 potential highly erodible marginal lands in the study. Agriculturally lower productive areas were
195 identified as areas with less than 5th-percentile SWAT simulated corn yield in the watershed
196 (Cibin *et al.*, 2016). Soils with land capability class (LCC) greater than 2 were identified as land
197 capability-based marginal lands (Feng *et al.*, 2015). Hypothetical extreme case prime agricultural
198 land conversion to bioenergy crops were considered in this study to understand the potential
199 impacts of extreme case scenarios. Six prime agricultural land conversion scenarios were
200 considered with 50% and 100% corn/soybean area conversion to perennial grasses. In 50%
201 agricultural area conversion scenarios the areas were selected using random selection and
202 strategic selection methods. Strategic selection was based on slope criteria, with the top 50%
203 highest slope corn/soybean areas selected for land conversion to bioenergy crops. The
204 comparison of random and strategic selection should provide insight towards the opportunities in
205 design of optimum cropping patterns. The improved SWAT model (Trybula *et al.*, 2015) was
206 used in this study to represent the perennial bioenergy crops. A detailed discussion of stover
207 removal representation and bioenergy crop representation in the study watersheds can be
208 obtained from Cibin *et al.* (2012 and 2016).

209 **Ecosystem services quantification**

210 The ecosystem services related to fresh water provision (FWPI), food provision (FPI),
211 fuel provision (FuPI), erosion regulation (ERI), and flood regulation (FRI) were quantified for
212 the baseline and bioenergy scenarios with current and future climate scenarios. Ecosystem

213 services were quantified using the methodology developed by Logsdon and Chaubey (2013).
214 Ecosystem services are estimated by comparing the current state (or simulated scenario results)
215 with standard or targeted values in each ecosystem service attribute. For example, FWPI is
216 estimated by comparing simulated streamflow and water quality with minimum and seasonal
217 flow requirement, sediment, and nutrient concentration standards. An ecosystem service index
218 ≥ 1 indicates that the watershed is meeting the service target. An ecosystem service index less
219 than 1, represents that the system is not meeting the service target. Logsdon and Chaubey (2013)
220 have provided suggestions on standards/ targets and have evaluated ecosystem services for a few
221 land use scenarios in the Wildcat Creek watershed. This study used similar standards
222 (parameters, Table 2) for quantifying ecosystem services for the baseline and bioenergy
223 scenarios. Fuel provisioning was estimated in this study by comparing whether the watershed can
224 support a medium sized (30 million gallon) biofuel refinery. Only cellulosic biofeedstock based
225 fuel production was considered in the fuel provisioning estimation. The ecosystem services were
226 quantified annually, and average annual values for each scenario were analyzed in detail at the
227 watershed outlet.

228 **Results and Discussion**

229 **Ecosystem services of current land use (Baseline scenario):**

230 The ecosystem services of the baseline scenario indicated good (close to one) fresh water
231 provisioning for the St. Joseph River watershed while for the Wildcat creek watershed, fresh
232 water provisioning was low (Figure 2). A detailed analysis of the different components of fresh
233 water provisioning indicated that streamflow for both watersheds was generally above the
234 environmental flow requirements (Figures S2 and S3). The predicted sediment concentrations for
235 the Wildcat Creek watershed was much higher than that of the St. Joseph River watershed. A
236 higher sediment concentration from the Wildcat Creek watershed could be attributed to
237 intensively managed cropping lands in the watershed which accounted for about 70% of the total
238 area and a lower streamflow. Even though the St. Joseph River watershed has a hilly terrain, the
239 percentages of forested and grassland (pasture) areas were much higher than the Wildcat Creek
240 watershed which tend to reduce soil erosion. The simulated nutrient loading from Wildcat Creek
241 was also higher than St. Joseph River watershed (Figures S2 and S3). The total phosphorus
242 concentration was higher than the water quality standard of 0.3 mg/l for both watersheds and the

243 loading trend was similar to sediment loading. The predicted nitrate concentration in Wildcat
244 Creek was estimated (2.9 mg/l) to be higher than the St. Joseph River (1.5 mg/l) and was above
245 the drinking water standard (10 mg/l) for many days (Figures S2 and S3). Almost 90% of the
246 cropped land in the Wildcat Creek watershed has tile drainage, a pathway of nitrate nitrogen
247 export (Kladivko *et al.*, 2004). The total amount of nitrogen and phosphorus fertilizer applied in
248 the Wildcat Creek watershed is also more than in the St. Joseph River watershed due to larger
249 cropped area. The flood regulation service for both watersheds was low with values of 0.3 and
250 0.4 for Wildcat Creek and St. Joseph River, respectively, and the Erosion Regulation Index
251 greater than 1.8 for both watersheds, indicating that the erosion from the watersheds never
252 exceeded the tolerable soil loss. Food provisioning for the baseline scenarios was estimated as
253 close to one and fuel provisioning was zero for both watersheds. The minimum grain yield
254 parameter for food provisioning was assumed as the 10 year mean grain yield from the
255 corresponding watershed and the model was calibrated during the same period based on USDA-
256 NASS (2016) data. Food provisioning for the St. Joseph River watershed was slightly more than
257 one (1.25) since the model over estimated grain yield from the watershed. This signifies the
258 importance of model calibration in the application of simulation models in quantifying
259 ecosystem services. Some of the corn grain produced in the watershed could be used in the grain
260 based ethanol production. However, the current study considered only second generation biofuel
261 production from bioenergy crops as the source of fuel production and thus the baseline fuel
262 provisioning is estimated as zero for both watersheds.

263 **Ecosystem service of bioenergy scenarios:**

264 The provision of ecosystem services generally improved in both watersheds with the
265 introduction of perennial energy crops into agricultural production (Figure 3). Introducing
266 *Miscanthus* on high slope CS areas (Scenario 1) of the Wildcat Creek watershed (Table 1)
267 increased fuel provisioning to 0.7 which is equivalent to 21 million gallons of biofuel production,
268 with an associated 8% reduction in food provisioning, improving fresh water provisioning by
269 22% and erosion regulation by 12% in comparison to the baseline scenario (Figure 2). Adoption
270 of *Miscanthus* on marginal lands in the St. Joseph River watershed, which has more area with
271 $\geq 2\%$ slope (Table 1), improved fresh water provisioning to 1.7 and could potentially produce
272 about 56 million gallons of biofuel (Figure 3). Food provisioning from the watershed is reduced
273 to 0.8 or about a 32% reduction from the baseline. Inclusion of bioenergy crops in agriculture

274 didn't have a significant impact for flood regulation. The flood regulation ranged between 0.29
275 and 0.34 for Wildcat creek and between 0.4 and 0.43 for the St Joseph River (Table S1).
276 Previous simulation studies also reported that the impacts on streamflow with inclusion of
277 bioenergy crops to agriculture was minimal compared to water quality impacts (Cibin *et al.*,
278 2016).

279 Maximum improvement in freshwater provisioning, flood regulation and erosion
280 regulation for both watersheds was estimated with conversion of all CS area to switchgrass
281 (Scenario 13, Table S1). The maximum fuel provisioning was with *Miscanthus* grown in all CS
282 areas (Scenario 12) in the Wildcat creek watershed, with a potential 277 million gallons of
283 biofuel production. The maximum biofuel production potential from the St. Joseph River
284 watershed was also with *Miscanthus* growing in all CS areas (174 million gallons) followed by
285 *Miscanthus* in all pasture areas (120 million gallons). The corn stover removal scenario resulted
286 in maximum fuel provisioning and minimal impacts on food provisioning and other ecosystem
287 services. The magnitude of changes in ecosystem services were heavily associated with the
288 magnitude of land use change. A normalized comparison of change in ecosystem service with
289 change in land use area (Figure S4) indicate that bioenergy placement with high slope areas
290 resulted in the maximum environmental benefits.

291 Ecosystem services from *Miscanthus* and switchgrass scenarios were very similar except
292 for fuel provisioning (Figure 3, Table S1). The predicted *Miscanthus* yields were almost twice as
293 much as the Shawnee switchgrass yields with a similar difference in fuel provisioning services.
294 Water quality benefits of both grasses were reported to be in a similar range (Trybula *et al.*,
295 2015), thus fresh water provisioning and erosion regulation remained similar for both grasses.
296 Both perennial grasses were simulated with the same fertilization rates as described in Trybula *et*
297 *al.* (2015). Also the SWAT parameters related to soil erosion simulation are considered to be the
298 same as those suggested in Trybula *et al.* (2015). For example, SWAT uses the Modified
299 Universal Soil Loss Equation (MUSLE; Williams and Berndt, 1976) to estimate soil erosion and
300 the USLE minimum crop factor (C factor) for both perennial grasses was set at 0.003. These
301 factors also contributed towards similar values for fresh water provisioning and erosion
302 regulation for both grasses. The changes in food provisioning are driven by the amount of CS
303 land use change and thus remained the same for both *Miscanthus* and switchgrass scenarios.

304 Marginal land area conversion to bioenergy crops are considered as the first choice for
305 bioenergy production in the case of agricultural area conversion due to potential environmental
306 benefits and associated minimal impacts on food production. This study considered three types
307 of marginal land definitions and the results indicate that slope based marginal land conversions
308 have maximum benefits in freshwater provisioning and erosion regulation. Crop productivity
309 based marginal lands (Scenario 3 and 4) were found to have less impact on food provisioning
310 (Figure S4). The St. Joseph River watershed had more marginal land area available and could
311 potentially produce 81 million gallons of biofuel when *Miscanthus* was grown in all marginal
312 lands (Scenario 7). The Wildcat Creek watershed can support one 30 million gallon refinery if all
313 marginal land areas were converted to *Miscanthus* production with an associated 12% reduction
314 in food provisioning.

315 Corn stover removal in low slope CS areas (Scenario 9) increased fuel provisioning in
316 both watersheds (Figure 3, Table S1). Fresh water provisioning and erosion regulation were
317 slightly reduced (<5%) for the corn stover removal scenario compared to the baseline. Corn
318 stover is generally left in the field after grain harvest and is expected to improve soil cover, soil
319 moisture retention, and reduce sediment and nutrient losses. Previous studies have reported
320 reduced streamflow and nitrate loading, and increased sediment and phosphorus loading with
321 stover removal (Cibin *et al.*, 2012). The combined effect of these factors may have cancelled out
322 the negative impacts of stover removal with less impact on fresh water provisioning. Simulations
323 showed that stover removal from the low slope areas had relatively minor environmental
324 impacts, and therefore, could be considered as a future best management practice, if stover is
325 utilized as a biofeedstock for biofuel production. Improved fuel provisioning from stover
326 removal with minimal impacts on food provisioning and other ecosystem services also suggests
327 stover as a suitable biofeedstock from the region. Pasture area conversion for bioenergy
328 production also increased fuel provisioning with minimal impacts on other ecosystem services.
329 Introducing *Miscanthus* in pasture areas yields fuel provisioning of 0.6 and 0.4 for the Wildcat
330 Creek and St. Joseph River watersheds, respectively. Only corn/soybean yield was considered in
331 the food provisioning, thus pasture area conversion had no impact. In pasture area conversion to
332 biofuel production, one perennial grass (*Miscanthus* or switchgrass) replaces another perennial
333 grass (Tall fescue) and thus simulation results showed minimal impacts on ecosystem services.

334 Prime agricultural land conversion to bioenergy crops significantly improved fresh water,

335 fuel provisioning and erosion regulation while significantly reducing food provisioning (Figure
336 3). Comparison of random placement and strategic placement of bioenergy crops within the
337 watersheds yields higher freshwater provisioning and erosion regulation with strategic placement
338 of energy crops. Food provisioning from both random and strategic placement is very similar
339 while the fuel provisioning was estimated slightly lower for the strategic selection scenario
340 compared to random placement. The random selection scenario improved freshwater
341 provisioning with *Miscanthus* by 50% from the baseline while the strategic placement scenario
342 improved fresh water provisioning by 145% from the baseline in the Wildcat creek watershed.
343 For the St. Joseph River watershed, randomly selected 50% CS area conversion had lower fresh
344 water provisioning (1.5) than that of high slope marginal lands (1.73) which accounted for 33%
345 of the CS area. Comparison of random and strategic selection provides prospects on improving
346 ecosystem service benefits from bioenergy production with careful selection of areas where
347 bioenergy crops could be grown.

348 **Ecosystem service of bioenergy scenarios under climate change scenarios:**

349 Changes in ecosystem services of both the baseline and bioenergy scenarios, in response
350 to climate change and variability, was quantified using the previously described nine projections
351 of future climate (three GCMs each executed with three emission scenarios). Each future climate
352 scenario was used as input to SWAT in the two watersheds for 100 years (10-year warm up, 3
353 thirty year periods of past, present and future).

354 The climate projection simulation results show that current (1990-2019) and future
355 climate periods (2020-2049) have lower ecosystem services compared to the past time period
356 (1960-1989) for baseline scenario. The reduction in fresh water provisioning, flood regulation
357 and erosion regulation are less than 6% in both watersheds (Table S2 and S3) and the changes in
358 ecosystem services were within the uncertainty band of the nine future climate projections
359 (Figures 4 and 5). Among all ecosystem services considered, climate change had the greatest
360 effect on food provisioning. Baseline scenario food provisioning decreased by 9% due to the
361 future climate compared to past climate in both watersheds. Air temperature increased in all
362 future climate scenarios compared to the past climate (Figure S1). Precipitation trends varied
363 across different seasons with increased precipitation in spring months (March-May) and reduced
364 summer precipitation (June – August) (Figure S1). The changes in temperature and growing

365 season precipitation had more impact on crop growth in the future climate period compared to
366 other ecosystem services. Flood regulation for the baseline scenario in the St. Joseph watershed
367 was reduced by 6% in 1990-2019 compared to 1960-1989 and increased by 3% in 2020-2049
368 compared to 1990-2019. A similar trend was also seen in the Wildcat Creek watershed with a
369 reduction in flood regulation between current and past climates, and no change in flood
370 regulation under future climate compared to current conditions. The changes in flood regulation
371 were very small and this fell within the climate change prediction uncertainty between the 9
372 projections (Figure 4 and 5). There was little difference in ecosystem services between the three
373 emission scenarios (A1B, A2 and B1) for the future climate period (2020-2049) (Figure S5). The
374 largest observed difference was with the A1B emission scenario, which had an improved flood
375 regulation index (0.57) compared to the other two emission scenarios (0.5) in the St. Joseph
376 River watershed. Flood regulation of 0.57 with A1B emission scenario represents 2.53 flood
377 events with 84.4 m³/sec magnitude and 4.37 days duration per year, while 0.5 flood regulation
378 with A2 and B1 scenario represents 2.96 flood events with 93.2 m³/sec magnitude and 6 days
379 duration per year.

380 The changes in ecosystem services in response to future climate change for different
381 bioenergy scenarios followed similar trends to those of the baseline scenarios (Figures 4 and 5).
382 Similar to food provisioning in the baseline, the food and fuel provisioning for biofuel scenarios
383 also indicated higher variations with climate change compared to other ecosystem services.
384 Changes in food provisioning was consistently around 9% for all scenarios in both watersheds
385 for future climate. Impacts of climate change on fuel provisioning was slightly more for
386 switchgrass based scenarios (9%, average change of all switchgrass scenarios comparing future
387 and past climate) compared to *Miscanthus* (7%, average change of all switchgrass scenarios
388 comparing future and past climate) for all land use change scenarios. Prime agricultural area
389 conversion to perennials (Scenarios 12-17) had maximum impacts with climate change on flood
390 regulation (10% in Wildcat creek and 6% in St. Joseph watershed, comparing future and past
391 climates). Climate change had minimal impacts on fresh water provisioning and erosion
392 regulation in both watersheds with all scenarios, with changes less than 5%. In general, land use
393 change is a more significant driver affecting ecosystem services than climate change for the
394 conditions evaluated in this study. For example in scenario 7 of St. Joseph watershed, land use
395 change induces 82% change in freshwater provisioning while climate change induces only 2%

396 change in future climate compared to past climate (Figure 5. A). This could be due to the scale of
397 land use change scenarios discussed in this study and considering the future climate change only
398 until 2050.

399 **Conclusions**

400 Five provisioning and regulating ecosystem services were evaluated for seventeen
401 futuristic bioenergy based land use change scenarios for two watersheds in the U.S. Corn Belt
402 region. Uncertainty in ecosystem services from climate change and variability was assessed
403 using climate projection data. SWAT was used to represent the bioenergy production scenarios.
404 In general, water quality is improved with perennial grasses in agricultural areas. All ecosystem
405 services except food provisioning were improved with bioenergy production compared to the
406 baseline. Introduction of bioenergy production provided a balanced ecosystem with regard to all
407 five ecosystem services considered in this study. The major findings of the study include:

- 408 ▪ The comparison of the baseline scenarios for the two study watersheds shows that an
409 increase in agricultural area reduces fresh water provisioning of the watershed.
- 410 ▪ Perennial bioenergy crops in agricultural areas improve fresh water provisioning and
411 erosion regulation in both watersheds.
- 412 ▪ Corn stover for bioenergy production increases fuel provisioning from watersheds
413 with minimal impact on other ecosystem services if stover removal occurs primarily
414 on low-sloping lands.
- 415 ▪ Flood regulation is least affected among the five ecosystem services with bioenergy
416 based land use changes.
- 417 ▪ Perennial bioenergy crop production in high slope areas substantially increases
418 ecosystem service benefits.
- 419 ▪ Impacts of land use change on ecosystem services is expected to be greater than the
420 climate change and variability impacts.
- 421 ▪ Climate change had more impact on food and fuel provisioning compared to other
422 ecosystem services considered in this study.

423 In this study, ecosystem services were evaluated using SWAT simulations. There is a

424 need to quantify uncertainty in the model simulations. Additionally, only five ecosystem
425 services were considered in this study, and future work should include the quantification of more
426 ecosystem services. Real implications of bioenergy production on certain ecosystem services are
427 not limited to the watershed scale evaluated in this study. For example, changes in corn or
428 soybean production in a watershed may not affect the food requirement of people in the
429 watershed since a majority of the food production is exported and consumed outside the
430 watershed. However, evaluation of watershed-scale ecosystem services for bioenergy production
431 helps in estimating relative change in ecosystem services for different ‘what-if scenarios’ and
432 can guide decision-making related to meeting bioenergy production goals while enhancing
433 ecosystem services.

434

435 **Supporting Information**

436 Additional supporting information may be found online under the Supporting Information
437 tab for this article: Figures include evaluation of climate projections, comparison of SWAT
438 simulation with ecosystem service parameters, comparison of percent change in ecosystem
439 services and land use change, and ecosystem service comparison for emission scenarios. Tables
440 include ecosystem service evaluated with measured and projected weather data.

441 **Acknowledgements**

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443 Number DE-EE0004396 and USDA-NIFA under Award Number 2009-51130-06029.

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580 **Table 1.** Description of biofuel scenarios. Baseline scenario describes current land use in the watersheds. Abbreviations: corn/soybean
 581 (CS), Land Capability Class (LCC), harvest 50% of stover available during harvest operation (Stover 50). *Miscanthus* was modeled as
 582 *Miscanthus × giganteus*, switchgrass was modeled as *Panicum virgatum*. The total watershed area for Wildcat Creek is 2045 km² and
 583 St. Joseph watershed is 2756 km².

| Scenarios | Bioenergy crops | Land use converted | Scenario Name | Area converted | | | |
|-----------|-------------------|--|---------------|-------------------------|---------------------|-------------------------|---------------------|
| | | | | Wildcat Creek | | St. Joseph | |
| | | | | Area (km ²) | % of watershed area | Area (km ²) | % of watershed area |
| 1 | <i>Miscanthus</i> | high slope marginal land (slope \geq 2%) | SlopeMarg-M | 119.5 | 6% | 347 | 13% |
| 2 | Switchgrass | high slope marginal land (slope \geq 2%) | SlopeMarg-S | 119.5 | 6% | 347 | 13% |
| 3 | <i>Miscanthus</i> | agricultural marginal land | AgMarg-M | 59.6 | 3% | 119.1 | 4% |
| 4 | Switchgrass | agricultural marginal land | AgMarg-S | 59.6 | 3% | 119.1 | 4% |
| 5 | <i>Miscanthus</i> | land capability marginal land (LCC>2) | LccMarg-M | 22.8 | 1% | 186.4 | 7% |
| 6 | Switchgrass | land capability marginal land (LCC>2) | LccMarg-S | 22.8 | 1% | 186.4 | 7% |
| 7 | <i>Miscanthus</i> | combined marginal land | AllMarg-M | 177.9 | 9% | 496.8 | 18% |
| 8 | Switchgrass | combined marginal land | AllMarg-S | 177.9 | 9% | 496.8 | 18% |
| 9 | Stover 50 | low slope corn/soybean | Stover50 | 1329.6 | 65% | 702.3 | 25% |
| 10 | <i>Miscanthus</i> | pasture area conversion | Past-M | 102.5 | 5% | 710.4 | 26% |
| 11 | Switchgrass | pasture area conversion | Past-S | 102.5 | 5% | 710.4 | 26% |
| 12 | <i>Miscanthus</i> | 100% CS conversion | 100CS-M | 1449.1 | 71% | 1049.2 | 38% |
| 13 | switchgrass | 100% CS conversion | 100CS-S | 1449.1 | 71% | 1049.2 | 38% |

| | | | | | | | |
|----|-------------------|---------------------------------------|--------------|-------|-----|-------|-----|
| 14 | <i>Miscanthus</i> | 50% CS conversion- random selection | 50CS-M-Rand | 725.6 | 35% | 524 | 19% |
| 15 | switchgrass | 50% CS conversion- random selection | 50CS-S-Rand | 725.6 | 35% | 524 | 19% |
| 16 | <i>Miscanthus</i> | 50% CS conversion-strategic selection | 50CS-M-Strat | 723.6 | 35% | 524.7 | 19% |
| 17 | switchgrass | 50% CS conversion-strategic selection | 50CS-S-Strat | 723.6 | 35% | 524.7 | 19% |

585 **Table 2.** Ecosystem service quantification parameters used for the two study watersheds, based
 586 on the methods of Logsdon and Chaubey (2013).

| Ecosystem service | Parameters | Estimated as | Units | Wildcat Creek | St. Joseph |
|--------------------------|--|--------------------------------------|------------------------------|---------------|------------|
| Fresh water provisioning | Minimum required flow | 30% of long term mean USGS flow | m ³ /sec | 6.7 | 9.0 |
| | Seasonal environmental flow requirement | 10% seasonal mean flow | m ³ /sec (winter) | 2.4 | 3.6 |
| | | | m ³ /sec (summer) | 2.1 | 2.4 |
| | TSS standard | IDEM, (2011) | mg/L | 46.3 | 46.3 |
| | Nitrate | IDEM, (2011) | mg/L | 10 | 10 |
| Total Phosphorus | IDEM, (2011) | mg/L | 0.03 | 0.03 | |
| Food Provisioning | Minimum grain yield | 10 year average corn yield (NASS) | Mg/year | 6.5E+05 | 2.9E+05 |
| | | 10 year average soybean yield (NASS) | Mg/year | 2.3E+05 | 1.6E+05 |
| Fuel Provisioning | Feedstock to support 30 million gallon ethanol plant | | Mg/year | 3.5E+05 | 3.5E+05 |
| Erosion regulation | Max allowable erosion rate | USDA T factor | Mg/ha/year | 4.4 | 4.1 |
| Flood regulation | Flood flow | Q ₁₀ of flow | m ³ /sec | 49 | 77.9 |
| | Long term average flood duration | | days | 4.8 | 5.3 |
| | Long term average flood frequency | | count | 7.8 | 7.3 |
| | Long term average flood magnitude | | m ³ /sec | 84.0 | 127.7 |

587
 588 **List of Figures**
 589 **Figure 1.** Location map of the two study watersheds with locations of observational weather stations and
 590 future climate projection grids identified.

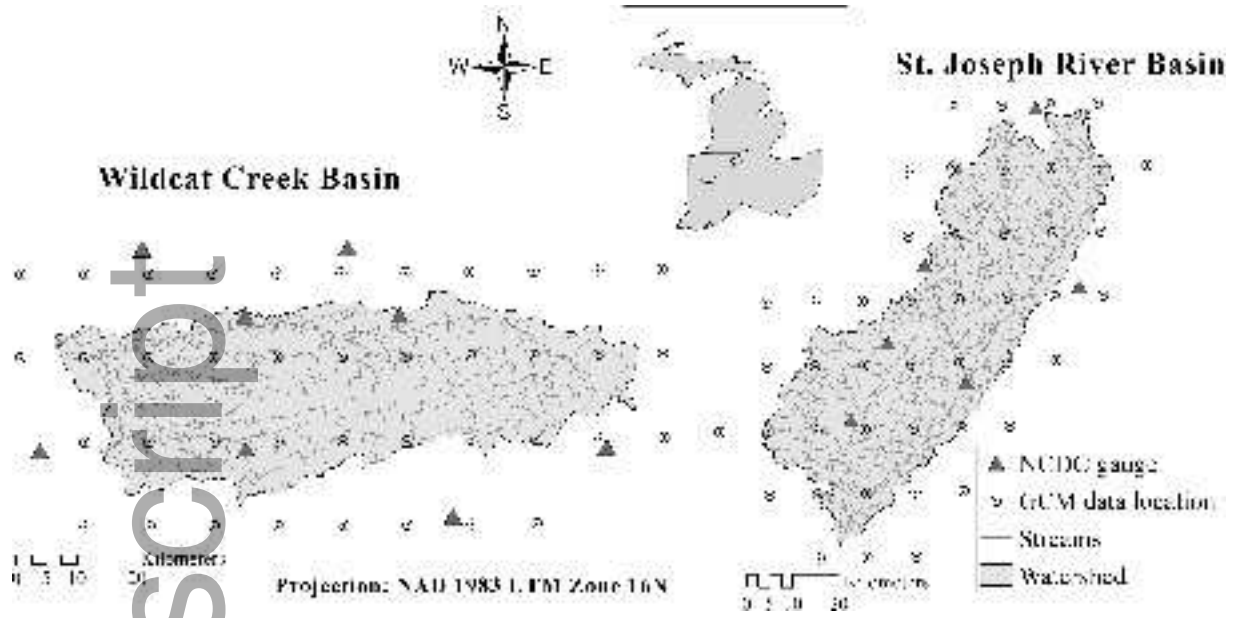
591 **Figure 2.** Quantification of ecosystem services for base line and *Miscanthus* in high slope ($\geq 2\%$ slope)
 592 areas in Wildcat Creek watershed (top) and St Joseph River watershed (bottom). Ecosystem
 593 services considered were freshwater provisioning (FWPI), food provisioning (FPI), fuel
 594 provisioning (FuPI), erosion regulation (ERI), and flood regulation (FRI).

595 **Figure 3.** Fresh water provision (FWPI), food provision (FPI) and fuel provision (FuPI) for the seventeen
596 bioenergy scenarios compared with baseline scenario.

597 **Figure 4.** Ecosystem service comparison of different scenarios under climate change scenarios for the Wildcat
598 Creek watershed. The error bar in each figure indicates the range of ecosystem services between the 9
599 projections of future climate. (A) compares baseline scenario with all marginal land converted to
600 *Miscanthus* (Scenario 7); (B) compares baseline with pasture area converted to *Miscanthus* (Scenario
601 10); (C) compares baseline with 50% stover removal (Scenario 9); and (D) compares random and
602 strategic conversion of 50% CS area to *Miscanthus* (Scenario 14 and 16).

603 **Figure 5.** Ecosystem service comparison of different scenarios under climate change scenario for St.
604 Joseph River watershed. The error bar in figure indicates the range of ecosystem services
605 between the 9 projections of future climate. (A) compares baseline scenario with all marginal
606 land converted to *Miscanthus* (Scenario 7); (B) compares baseline with pasture area converted
607 to *Miscanthus* (Scenario 10); (C) compares baseline with 50% stover removal (Scenario 9); and
608 (D) compares random and strategic conversion of 50% CS area to *Miscanthus* (Scenario 14 and
609 16).

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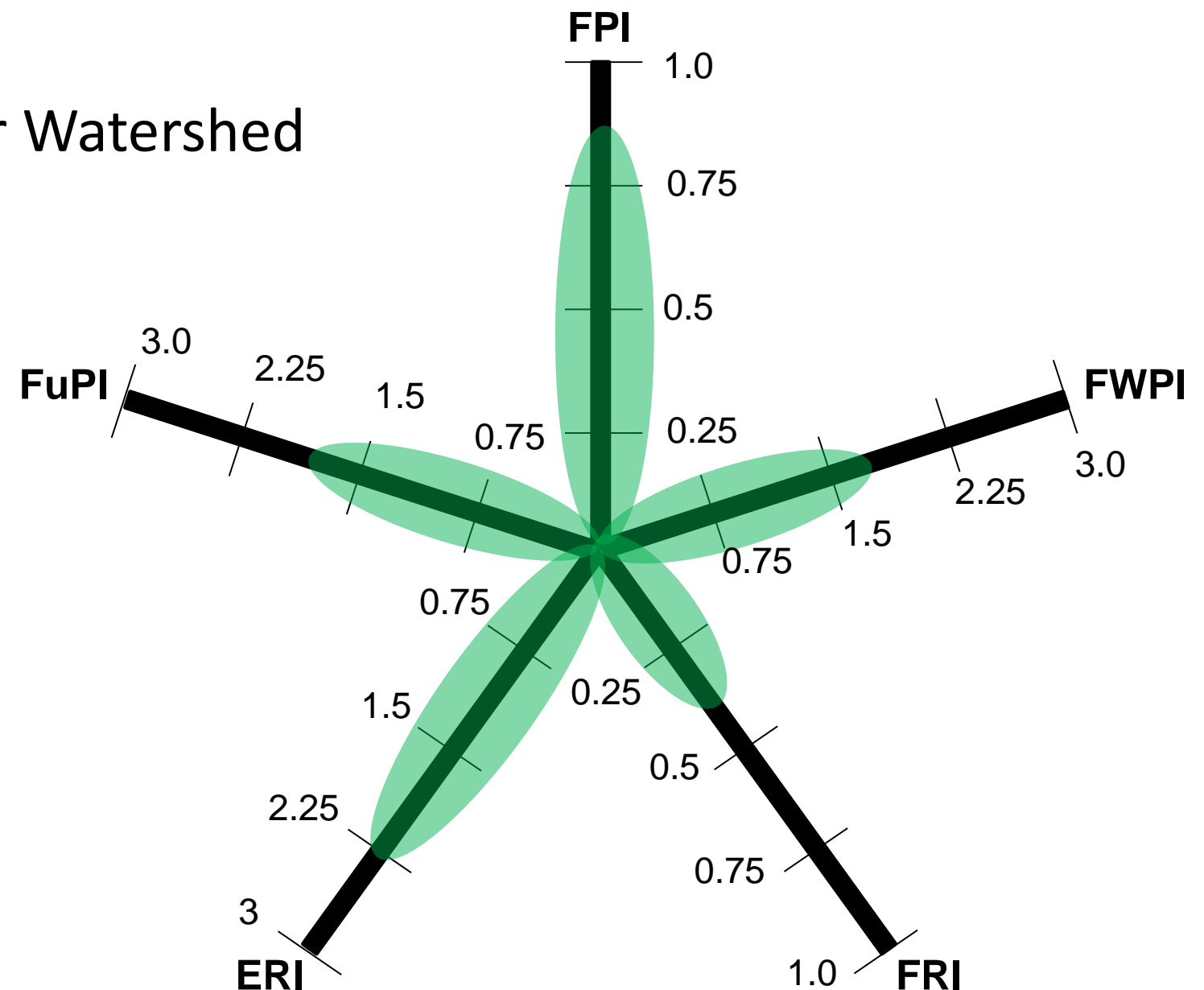
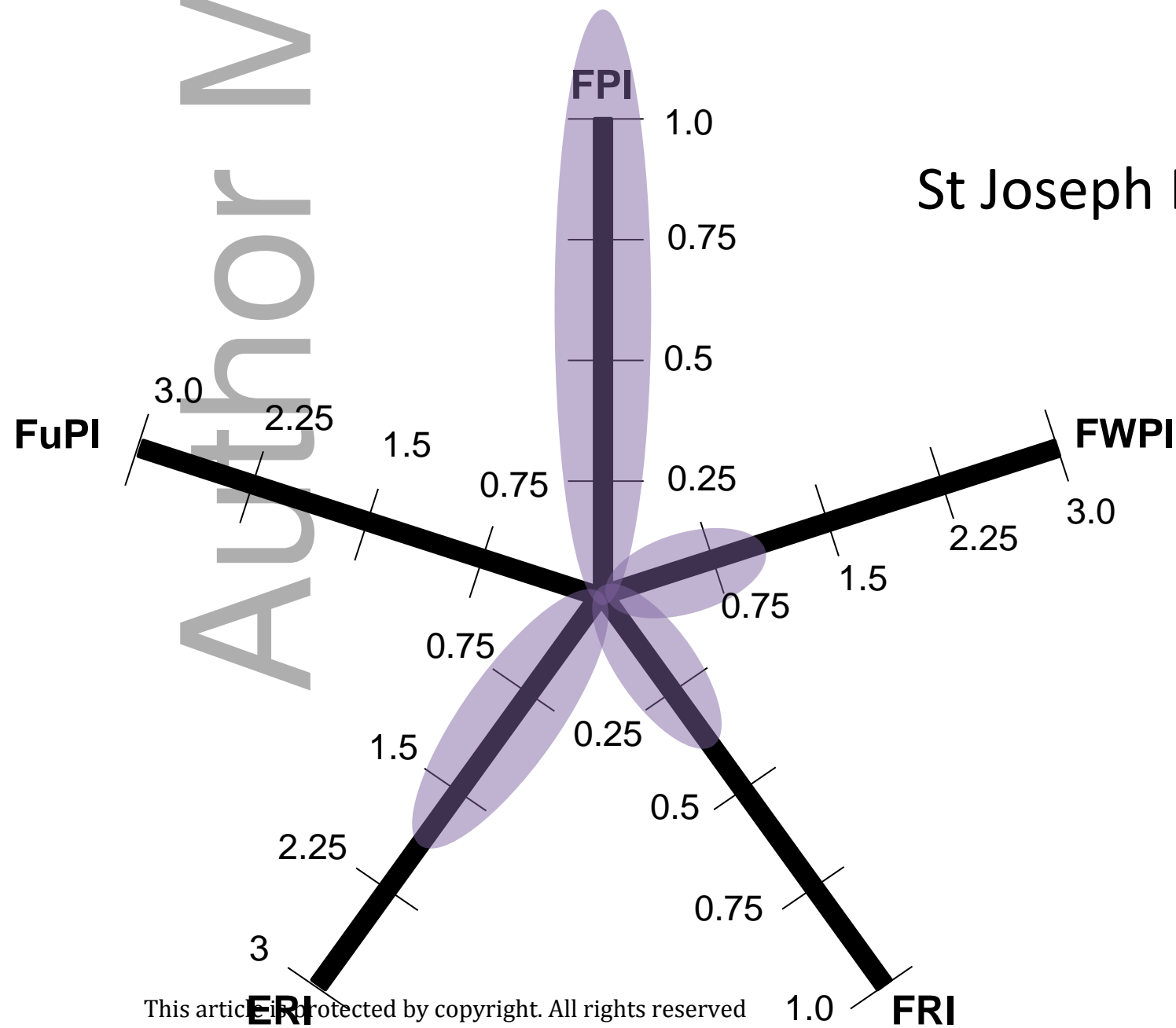
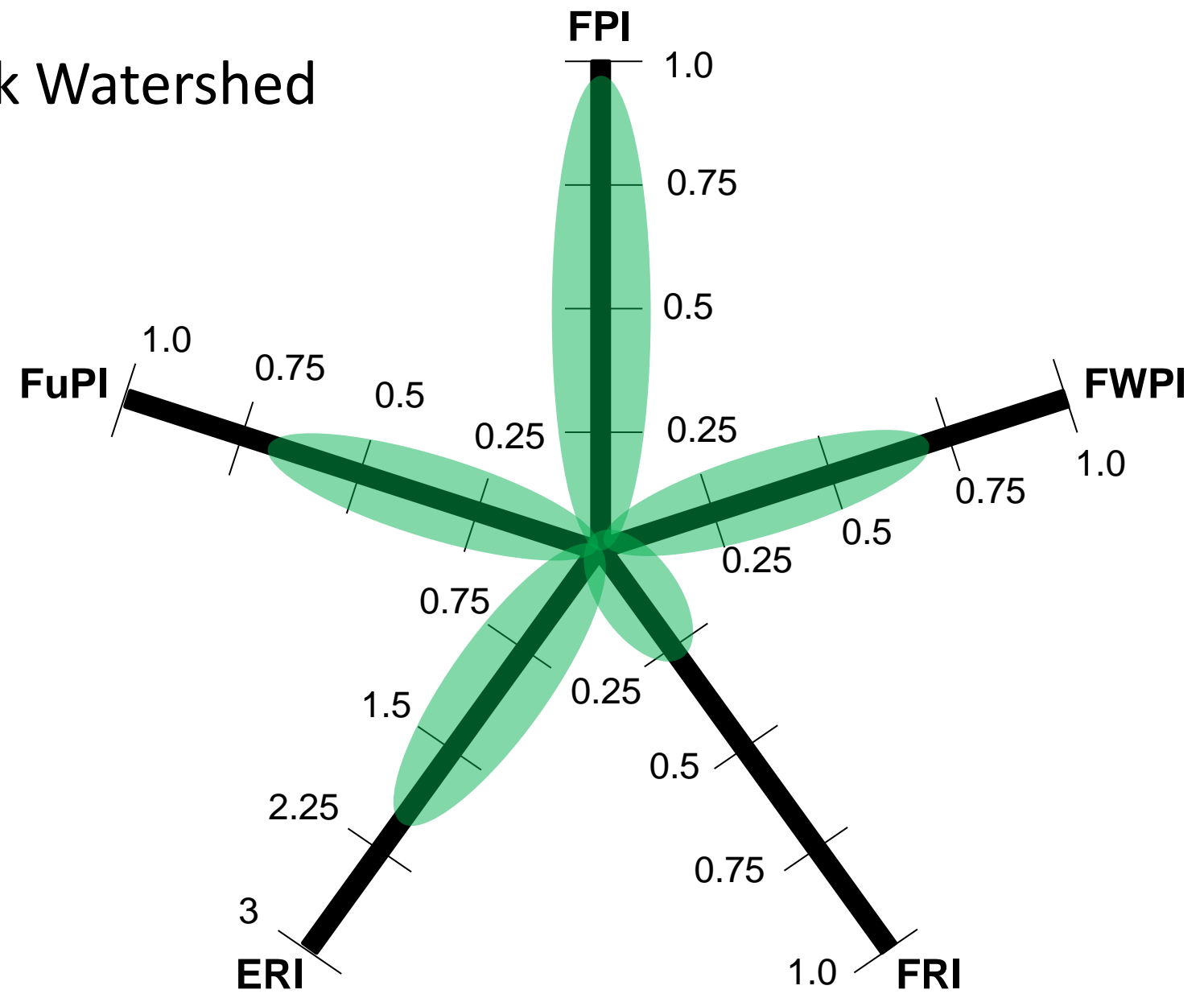
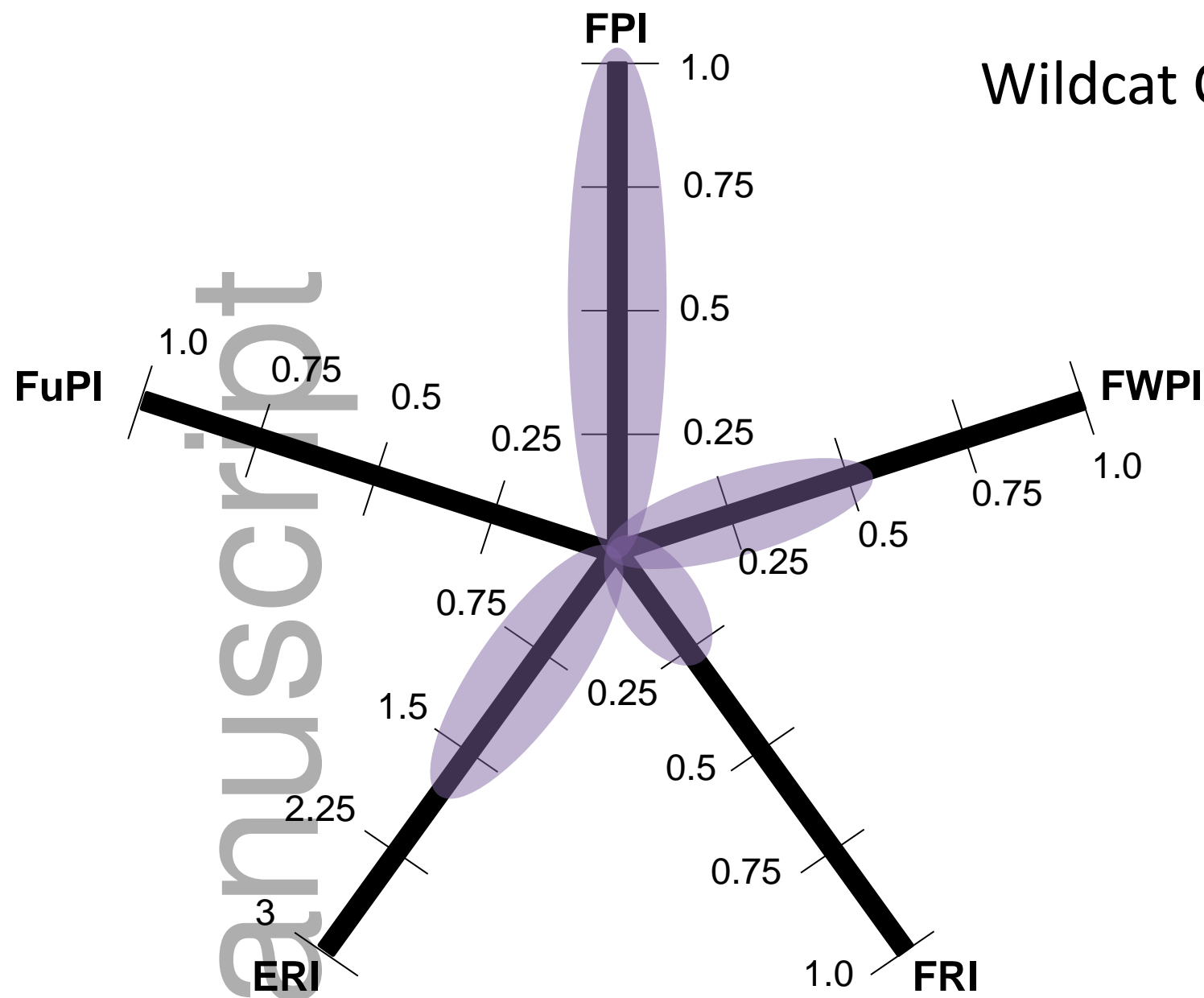


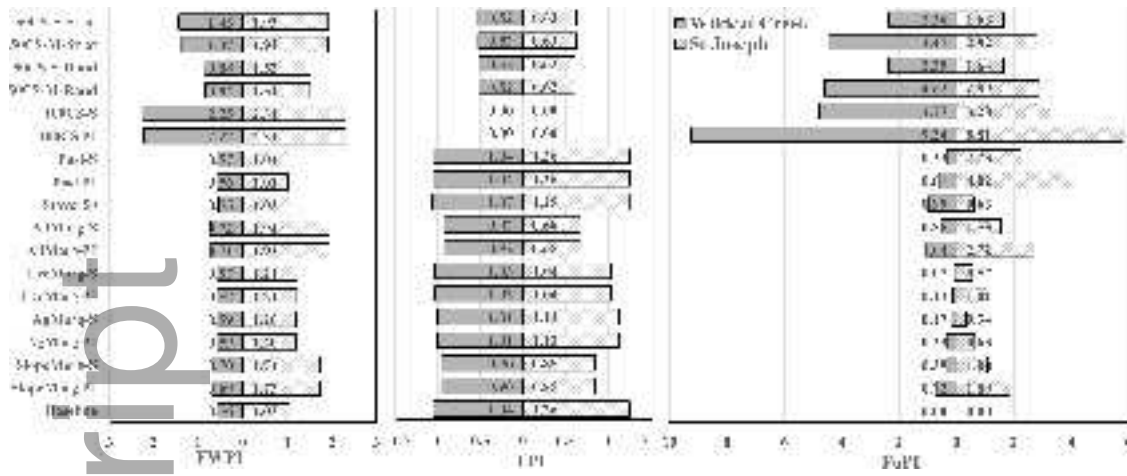
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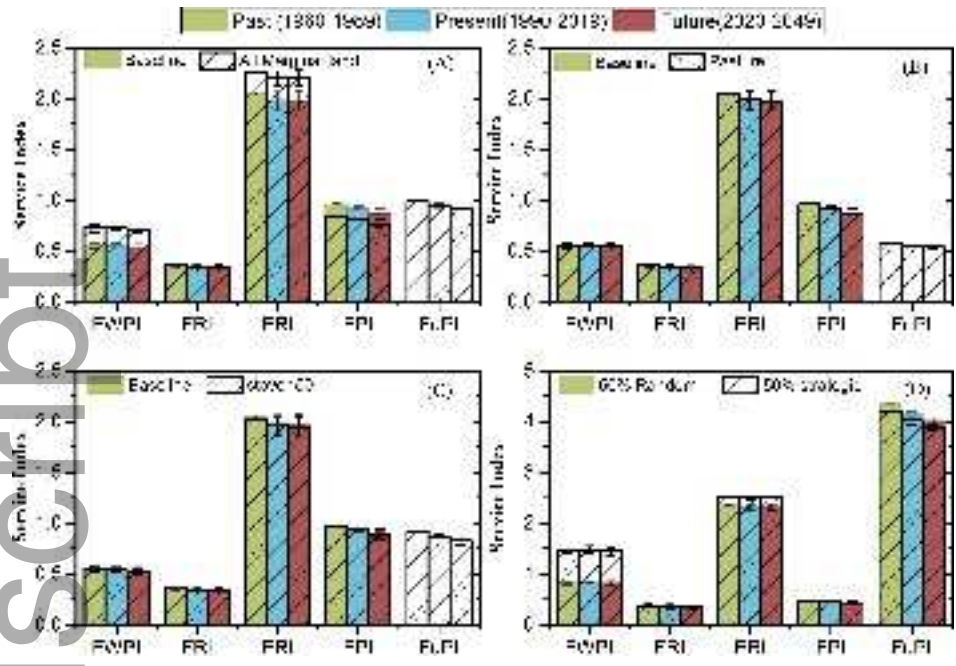
Baseline

Miscanthus in High Slope Areas

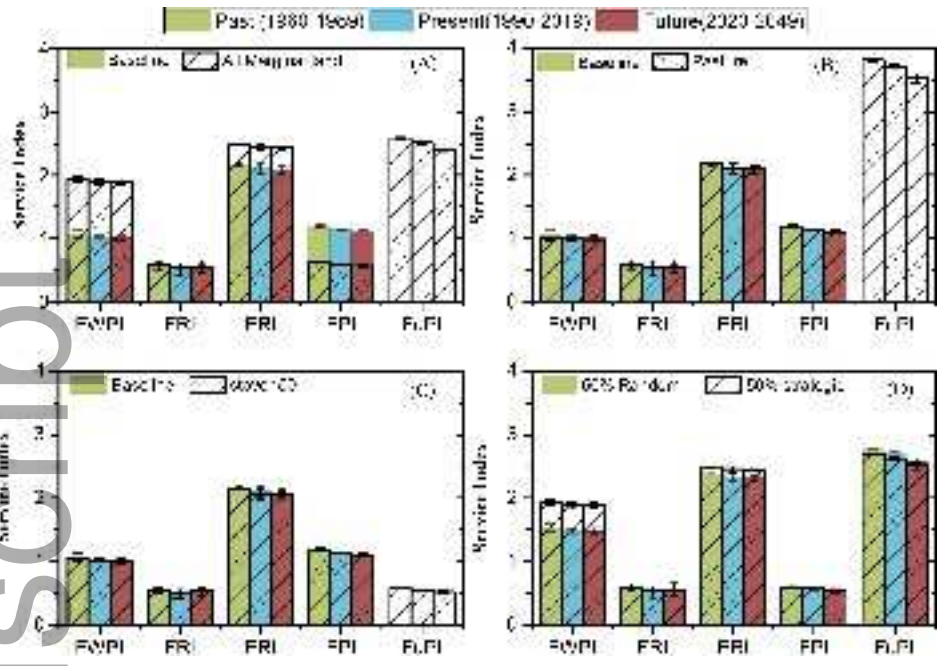




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