

Methanol Worked Examples for the TEA and LCA Guidelines for CO₂ Utilization



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Foreword from CO₂ Sciences

Climate change is one of the largest challenges of our time. One of the major causes of anthropogenic climate change, carbon dioxide, also leads to ocean acidification. Left unaddressed, these two challenges will alter ecosystems and fundamentally change life, as we know it. Under the auspices of the UN Framework Convention on Climate Change and through the Paris Agreement, there is a commitment to keep global temperature increase to well below two degrees Celsius. This will require a variety of strategies including increased renewable power generation and broad scale electrification, increased energy efficiency, and carbon-negative technologies.

We believe that Life Cycle Assessment (LCA) is necessary to prove that a technology could contribute to the mitigation of environmental impacts and that Techno-Economic Assessment (TEA) will show how the technology could be competitively delivered in the market. Together they are a valuable toolkit for promoting carbon capture and utilization (CCU) technology development.

The work presented here was made possible through the vision of the Chairman of CO₂ Sciences Inc., Bernard David, and the expertise of the CEO of CO₂ Sciences Inc., Issam Dairanieh.

The Global CO₂ Initiative was launched during the 2016 meeting of the World Economic Forum with the goal of catalyzing innovative research in CO₂ utilization. Starting July of 2018, the Initiative will continue its work as *The Global CO₂ Initiative at the University of Michigan*.

Development of standardized CO₂ Life Cycle and Techno-economic Assessment Guidelines was commissioned by CO₂ Sciences, Inc., with the support of 3M, EIT Climate-KIC, CO₂ Value Europe, Emissions Reduction Alberta, Grantham Foundation for the Protection of the Environment, R. K. Mellon Foundation, Cynthia and George Mitchell Foundation, National Institute of Clean and Low Carbon Energy, Praxair, Inc., XPrize and generous individuals who are committed to action to address climate change.

Global CO₂ Initiative@UM, August 2018

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List of abbreviations

ANOVA	Analysis of variance
BFD	Block flow diagram
CAPEX	Capital Cost
CAPM	Capital asset pricing model
CCS	Carbon capture and storage
CCU	Carbon capture and utilization
CEPCI	Chemical Engineering Plant Cost Index
CO₂	Carbon dioxide
COGM	Cost of goods manufactured
COGS	Cost of goods sold
ETS	Emission trading system
EU	European Union
FCI	Fixed capital investment
FOAK	First of a kind
GWP	Global warming potential
H₂	Hydrogen
IRR	Internal rate of return
ISBL	Inside battery limits
ISO	International standardization organization
LCA	Life cycle assessment
LCC	Life cycle costing
LCI	Life cycle inventory
LCOE	Levelized cost of electricity
LHV	Lower heating value
MADM	Multiple attribute decision making
MCDA	Multicriteria decision analysis
MODM	Multiple objective decision making
NGO	Non-Governmental Organisation
NOAK	Nth of a kind
NOX	Nitrous Oxides
NPV	Net present value
OPEX	Operational Cost
OSBL	Outside/off-site battery limits
P&ID	Piping and instrumentation diagram
PEM	Proton exchange membrane
PFD	Process flow diagram
R&D	Research and Development
ROI	Return on investment
SA	Sensitivity analysis
SI-UNITS	International System of Units
SMR	Steam methane reforming
TCI	Total cost indicator
TEA	Techno-economic assessment
TRL	Technology readiness level
UA	Uncertainty analysis
US DOE	United States Department of Energy
USD	United States Dollars
WACC	Weighted average cost of capital

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PART A

Introduction

A.1 Introduction

This document contains worked examples of how to apply the accompanying “Guideline for Techno-Economic Assessment of CO₂ Utilization” and “Guideline for Life Cycle Assessment of CO₂ Utilization”, hereafter referred to as "the TEA and LCA guidelines".

These worked examples are **not intended** to be a definitive TEA or LCA report on the process described, but are provided as supporting material to show how the TEA and LCA methodologies described in the guidelines can be specifically applied to tackle the issues surrounding CO₂ utilization.

The modelled examples were constructed using data collected from the literature and from the ecoinvent database v3.4. The aim of these worked example studies is not to prove whether the selected process is economically or environmentally viable or to make process alterations to make it so, but to clearly demonstrate how the proposed guidelines can be used to conduct a transparent TEA and LCA which can then be followed by others.

THE LCA WORKED EXAMPLE USES A MODEL PRODUCED SOLELY TO ASSIST IN THE USE AND INTERPRETATION OF THE ACCOMPANYING LCA GUIDELINES. THE FOLLOWING LCA WORKED EXAMPLE HAS THEREFORE NOT UNDERGONE AN EXTERNAL REVIEW IN ACCORDANCE WITH ISO 14040/14044 AND CONSEQUENTLY IT SHOULD NOT BE USED IN COMPARISONS OF OTHER LIFE CYCLE ASSESSMENTS OF CCU AND/OR METHANOL TECHNOLOGIES.

The TEA example is provided in Part A, followed by the LCA example in Part B. In Part C, the reader will find a short example of how a TEA and LCA can be integrated. This integration is not an exhaustive example. As many aspects can be analyzed to produce combined indicators and many approaches to multi-criteria decision making applied. However, it is included here to provide a starting point and initial example of how integration can be carried out.

A.1.1 About the chosen CCU process

These examples focus on the production of methanol from CO₂ for the purpose of providing a carbon feedstock for the chemicals industry. Methanol production was chosen as it is a familiar CO₂ utilization (CCU) route, which has been much studied, therefore it is hoped that the reader will be able to focus on the described methodology for conducting the assessment rather than understanding the specifics of the process route. For this reason, much of the Inventory data for the modelled process used for the LCA worked example has been placed into an appendix and only the critical data used for modelling the sections which are specific to CCU are provided within the main body of the report. This is done to help the reader focus upon applying the TEA and LCA guidelines and not the specifics of the data used. Having said that, the limitations of the data used is discussed in the examples, as this is likely to be an issue for many TEA and LCA reports of CCU technologies.

As methanol production via hydrogenation and PEM electrolysis of water to produce hydrogen are both at high technology readiness levels (TRL7+); a CO₂ capture technology currently at a lower TRL (membrane separation at TRL3 or 4) was selected to demonstrate the differences that can be observed in the interpretation phase when working on TEA and LCA studies of processes with lower TRLs. It is acknowledged that there are many unknown variables with membrane capture, and it is not within the remit of this work to draw conclusions on their application. However, it is known that organizations wish to conduct TEA and LCA studies across a range of TRLs and therefore we hope to demonstrate here how this could affect the results.

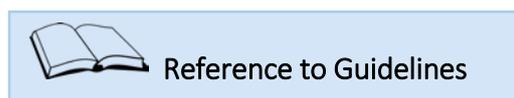
The intended application of these studies is as references on how to apply the guidelines for CO₂ utilization to a comparative assessment between a CCU technology and a conventional (reference or benchmark) technology. These examples are for public use and are targeted at the TEA and LCA practitioner who wishes to assess a CCU process. Methanol is chosen as the process to assess due to its production demand and its feasibility as a CO₂-based chemical.

A.1.2 How to read the worked examples

The subsequent TEA and LCA are written as worked examples, not in the format of a formal report for either academic, corporate or policy audiences. The examples are structured in this way to enable the reader to clearly understand how the guidelines have been applied by the authors, rather than focus on a specific style of reporting.

To enhance understanding the following explanations have been included:

- Light blue-coloured text boxes are used to refer the reader to specific sections of the guidelines:



- Blue-coloured speech bubbles give an explanation of why certain decisions were taken or choices made:

A blue speech bubble with a tail pointing towards the left, containing the text "This is here because...." in white.

This is here because....

The studies contain example Executive and Technical Summaries for the reader's benefit. All '**shall**' aspects of the guidelines are covered and '**should**' and '**may**' aspects and included as appropriate to each study.

The guidelines for conducting TEA and LCA of CO₂ utilization processes, together with the accompanying worked examples, were commissioned by The Global CO₂ Initiative/CO₂ Sciences.

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PART B TEA Worked Example

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B.1. Executive Summary

Methanol has a huge global market as a commodity chemical acting as a feedstock for many chemical processes. It can be synthesized from CO₂ instead of from natural gas, which means it becomes possible to indirectly substitute the fossil carbon currently used in a multitude of plastics, polymers and carbon containing chemicals with carbon from CO₂. Doing so would release the chemicals industry from its current dependency upon fossil oil feedstock.

The goal of this study is to determine the economic performance and technical viability of Methanol (MeOH) production via CO₂ hydrogenation within a renewable power to liquid context. The study conducts a TEA from an R&D perspective; the goal is to estimate total production costs and identify key cost drivers as well as gauge the market value of MeOH produced from CO₂ hydrogenation. As such indicators such as CAPEX, OPEX, electricity consumption have been calculated. The proposed CCU project can contribute to the 'methanol economy' concept; MeOH is one of the most flexible chemicals and an important energy carrier. Within this study methanol is considered in its application as a chemical feedstock. This report is of public use and is targeted for the TEA practitioner that wishes to learn about assessment of CCU processes.

The study is based on a 1,000 metric tonnes per day (t/d) plant situated in Germany. CO₂ is captured from flue gas by a membrane capture technology followed by a cryogenic unit (both at TRL 4), hydrogen is produced by PEM electrolyser using renewable energy (TRL 9) and then methanol is synthesised via CO₂ hydrogenation (TRL 7). The studied system is compared with a conventional methanol plant using fossil feedstocks and producing methanol by steam reforming of methane. Data for the CCU plant design has been taken from literature and then modelled using simulations.

This study concluded that the estimated price of methanol produced from CO₂ on a 1,000 t/d plant in Germany using membrane carbon capture and renewable hydrogen from water electrolysis was 3.5-fold greater than the current market price of conventional methanol. The factors that mostly affect the methanol price are associated with the electrolysis unit, i.e. electricity price and electrolyser efficiency. Changes in the renewable energy price cause wide fluctuations in the CCU methanol market price of -67% to +20%.

There are several technology and market developments that can significantly improve the economics of the modelled processes. If energy independence and security are prioritised and societies begin to place a meaningful monetary burden on carbon-intensive technologies, these developments could accelerate the adoption of sustainable products such as the investigated MeOH.



This section has been provided as an example of the level of detail that should be included in an Executive Summary.

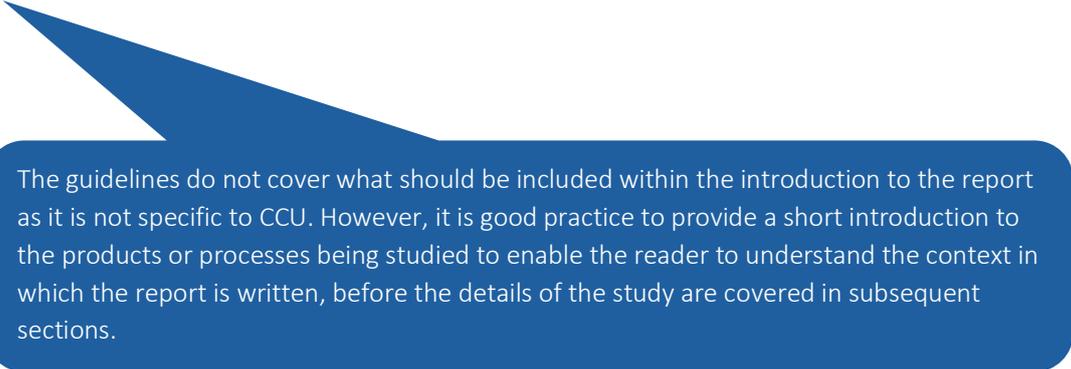
B.2. Technical Summary

GOAL	CCU product	Methanol as a chemical feedstock	
	Intended application and reasons for study	What is the economic performance of Methanol production via CO ₂ hydrogenation within a renewable power to liquid context?	
	Brief description	CO ₂ is captured via membrane capture from a cement plant, H ₂ is produced via PEM water electrolysis, subsequently methanol is produced via thermochemical synthesis	
	Intended audience	TEA practitioners	
	Commissioners and assessors	.	
	Limitations of study	<ul style="list-style-type: none"> Based on literature data, deviations in real world process will occur No infrastructure for H₂ production included Low TRL process for Carbon Capture 	
SCOPE	System boundary (e.g. cradle to gate)	Cradle to gate	
	Benchmark system	Steam reforming of methane to produce methanol	
	Plant size	1,000 metric tonnes per day	
	Functional Unit	The production of 1 tonne of MeOH for use in chemical industry	
	System elements and technology maturity	Carbon capture via membrane: TRL 4 PEM water electrolysis: TRL 9 Methanol Synthesis: TRL 7	
	Assessment indicators	Minimum MeOH Selling Price (MMSP), CapEx, OpEx, CO ₂ conversion, energy consumption	
INVENTORY	Data Source	<input type="checkbox"/> Primary sources <input checked="" type="checkbox"/> Secondary sources <input type="checkbox"/> Stoichiometric data	<input checked="" type="checkbox"/> Process modelling based data <input type="checkbox"/> Mixed sources <input type="checkbox"/> Other (please specify)
	Energy sources (select all that apply)	<input checked="" type="checkbox"/> Grid mix <input checked="" type="checkbox"/> Power station with Carbon Capture <input checked="" type="checkbox"/> Wind <input type="checkbox"/> Solar	<input type="checkbox"/> Nuclear <input type="checkbox"/> Hydro <input type="checkbox"/> Future (see timeframes) <input type="checkbox"/> Other (please specify)
	Base year	2016	
	Currency	Euro	
	Location	Germany	
	Plant life time	20 years	
	CO₂ sources and price (if applicable)	Not applicable – capture included in boundary	
	H₂ sources and prices (if applicable)	Not applicable – H ₂ production included in boundary	
	Energy consumption per functional unit	10.84 MWh/t _{MeOH}	
CAPEX per functional unit	336 €/t _{MeOH}		
OPEX per functional unit	1101 €/t _{MeOH}		
Price per functional unit	1402 €/t _{MeOH}		
INTERPRETATION	Sensitivity analysis main factors	Electricity price, electrolyser cost	
	Uncertainty manipulated variables	Electricity price, electrolyser efficiency, electrolyser cost, tax rate and CapEx of the CO ₂ capture unit.	
	Main Conclusions	<ul style="list-style-type: none"> The deterministic MMSP is 1402 €/t and there is a confidence interval of 95% that the MMSP is in the range of 1,238€/t to 1,448 €/t. These figures are 4.9-5.8 fold greater than the fossil MeOH price. Sensitivity analysis revealed that the process is OpEx intensive with the electricity price to pose as the major cost component. In fact, the factors that mostly affect the MMSP are associated with the electrolysis unit, i.e. electricity price and electrolyser efficiency rather than with variables related to the CO₂ conversion plant. 	

B.3. Introduction

Methanol can be used as a liquid fuel, either directly or after dehydration to dimethyl ether, but currently its main use is as a commodity chemical where it acts as a feedstock for many chemical processes. The market for methanol is large, with global demand reaching 70 Mt in 2015, supplied by over 90 methanol plants worldwide [1]. Methanol can be synthesized from CO₂ and is itself used as a chemical building block in the synthesis of many different compounds. By producing methanol using carbon capture and utilization (CCU) technologies from waste CO₂, rather than from steam methane reformation as per the conventional route, it becomes possible to indirectly substitute fossil carbon with carbon derived from CO₂ into a multitude of plastics, polymers and carbon containing chemicals. In this way, the dependency of the chemicals industry upon fossil oil feedstock is significantly reduced.

This report assesses the economic performance and technical viability of a CCU methanol plant capable of producing 1,000 MTPD.



The guidelines do not cover what should be included within the introduction to the report as it is not specific to CCU. However, it is good practice to provide a short introduction to the products or processes being studied to enable the reader to understand the context in which the report is written, before the details of the study are covered in subsequent sections.

B.4. Goal



Reference to TEA Guidelines

A checklist of items to be included in each section of the report is included in **Chapter B.8 Reporting** of the guidelines, which for the goal is as follows:

Goal of the study

- State goal – the intended application of the study and the reasons for the study
- State the target audience for the study
- State commissioner and authors of the study
- State limitations in the usability from assumptions or methods
- State the base case with current conditions

B.4.1 The Goal

The goal of this study is to assess the economic performance and technical viability of methanol production via CO₂ hydrogenation. The modelled 1000 tonnes per day plant incorporates carbon capture via membranes and cryogenics, PEM hydrogen production and methanol synthesis using renewable power. The plant model is located in Germany and based on 2016 costs.

The current study conducts a TEA from an R&D perspective; the goal is to estimate total production costs and identify key cost drivers as well as assess the profitability of MeOH production from CO₂ hydrogenation. The proposed CCU project can contribute to the ‘methanol economy’ concept; MeOH is one of the most flexible chemicals and an important energy carrier. Therefore, MeOH can play a significant role in the future at the crossroads of sustainable chemical production, energy generation and security [1,2].

The report is of public use and by accompanying the guideline document serves as an example for the TEA practitioner that wishes to assess a CCU process or product. Cost and prices are reported in EUR 2016 and the plant is assumed to be located in Germany. The process design and economic parameter value choices, underlying this analysis, are based on public domain literature. For these reasons, the results are not indicative of potential performance, but are meant to represent the most likely performance given the assumptions (time and location) and the current state of public knowledge. This study was commissioned by The Global CO₂ Initiative as part of the Guidelines project and was conducted by the University of Sheffield.



Reference to TEA Guidelines

This goal definition is described in **Guideline B.1**. The present report conducts a direct TEA from an R&D perspective. A feasibility assessment was carried that incorporates quantification of costs, profits and value for a power to MeOH plant *via* captured CO₂ hydrogenation. The report serves to provide information and showcase good practices to a potential CCU practitioner. In order to raise awareness, the report is publicly available. As stated in **Guideline B.1**, data derives from up to date conditions, and assumptions and results are location and time dependent.

B.4.2 Assessment scenarios

Scenario analysis is different from the high, medium, and low cases exploration. High, medium, and low scenarios look at different rates of progress along a path that is based on a single set of projections. Scenario analysis focuses on the areas of greatest uncertainty for a country or an operation, systematically develops several plausible alternative future environments in which the operation might be implemented, and determines how they would affect its success. Scenario analysis involves constructing or developing scenarios and integrating the content of scenarios into the decision making process. In the present study, after sensitivity and uncertainty analyses, hot spots will be identified. Scenarios for renewable electricity supply via onshore and offshore wind production will be assessed.



Reference to TEA Guidelines

Guideline B.2 presents possible ways to conduct scenario analysis. It is expected that, electricity price will be a significant costs component and therefore different scenarios of generating electricity will be investigated in the present study.

B.5. Scope of Study



Reference to TEA Guidelines

A checklist of items to be included in each section of the report is included in **Chapter B.8 Reporting of the guidelines**, which for the goal is as follows:

Scope of the study

- State product application(s) and functional unit, including consistency with goal and scope and reference flow
- State elements and boundaries of product system in a graphical scheme
- State the benchmark process
- State the selected indicator and methods, and including consistency with study goal and data availability associated with technology maturity

B.5.1 Product application and functional unit

MeOH can be used in the chemical industry as a solvent and as a C_1 building block for producing intermediates and synthetic hydrocarbons, including polymers and single-cell proteins. Furthermore, MeOH can be mixed with conventional gasoline without requiring any technical modification in the vehicle fleet and it can be used as a convenient energy carrier for hydrogen storage and transportation, as an easily transportable fuel (see Figure 1). It may be produced both from (1) fossil fuels (natural gas, shale gas and coal) and (2) non-fossil fuel sources (residue/biomass/renewables+CO₂). Consequently, MeOH is one of the most important and versatile platform chemicals for the chemical industry [3].

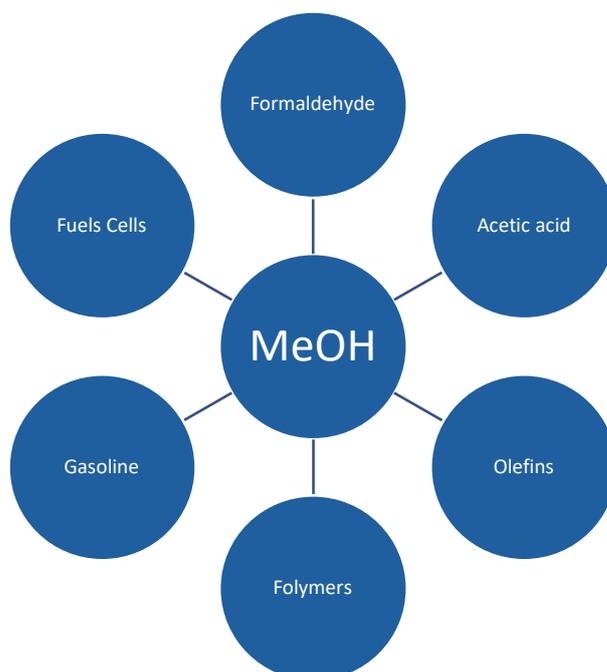


Figure 1. Different MeOH applications

In the present study MeOH is considered as a building block chemical, *i.e.* as intermediate for the production of other chemicals (*e.g.* formaldehyde, acetic acid). The use of MeOH as feedstock, driven largely from the MeOH to olefins process, is projected to increase and consequently the role of emerging fossil sources (such as shale gas) will be developed. However, strong evidence already exists that the future development of such fossil sources has been overestimated [4]. For this reason, alternative sources such as CO₂ can positively contribute to meet the increasing demand of MeOH in a sustainable manner. The functional unit of this study is 1 tonne of methanol considered as a chemical feedstock. The production output of the CCU methanol plant is 1000 tonne per day.

The market for MeOH as chemical feedstock is much larger compared to the market for MeOH as fuel. Furthermore, decarbonisation of transportation can be achieved via other technologies such as electric vehicles while options for substituting crude oil as a precursor for a variety of products are limited. Carbon will still be essential in the long run for chemicals and polymers and renewable MeOH can pose as a sustainable carbon source.



Reference to TEA Guidelines

As described in **Guideline B.3**, the product application must be included in the Scope of the study; here MeOH is treated as feedstock to produce chemicals. MeOH demand as chemical feedstock is projected to increase and the proposed CCU scenario can provide sustainable solutions. The functional unit definition is in accordance with **Guideline B.3** for substitutes CO₂-based chemicals and it is expressed on a mass basis; based on the scale of the plant metric tonnes of MeOH was selected as functional unit.

B.5.2 System elements and system boundaries

The examined power to liquid concept consists of an electrolyser, a CO₂ capture unit and MeOH synthesis infrastructure. The inputs to the system are deionised water, green electricity and flue gas (22% CO₂) derived from a typical cement production unit. The system boundaries treat H₂ and CO₂ as intermediate products rather than as inputs to the system. In this way, lack of transparency caused by sourcing CO₂ and H₂ priced from third parties is eliminated and instead costs of onsite production are considered. For the study, it was assumed that both CO₂ capture and hydrogen production occur onsite and hence long distance transportation costs (*e.g. via pipelines*) can be avoided. Boundaries are set as cradle to gate, therefore transportation or other costs after the factory gate are not considered.

The product system consists of four major elements:

- 1) CO₂ membrane based capture from flue gas,
- 2) H₂ production in a proton exchange membrane (PEM) electrolyser,
- 3) MeOH production *via* hydrogenation
- 4) utilities.

These elements constitute the inside battery limits (ISBL) plant; the outside battery limits (OSBL) plant consists of the balance of plant (BOP). The latter involves the required infrastructure for grid and water

supply connections. MeOH production is subdivided to the compression, synthesis and recovery sections. The utilities section includes the CHP (combined heat and power) unit, the wastewater treatment plant and cooling towers. Figure 3 is a simplified visual representation of the system boundaries.

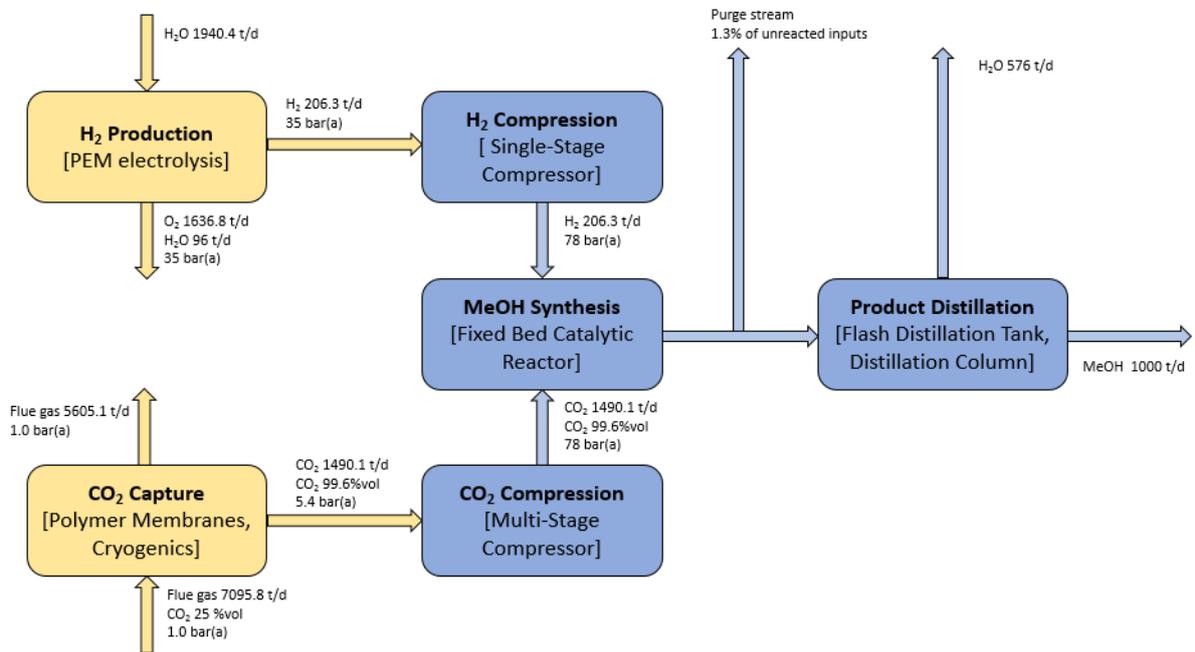
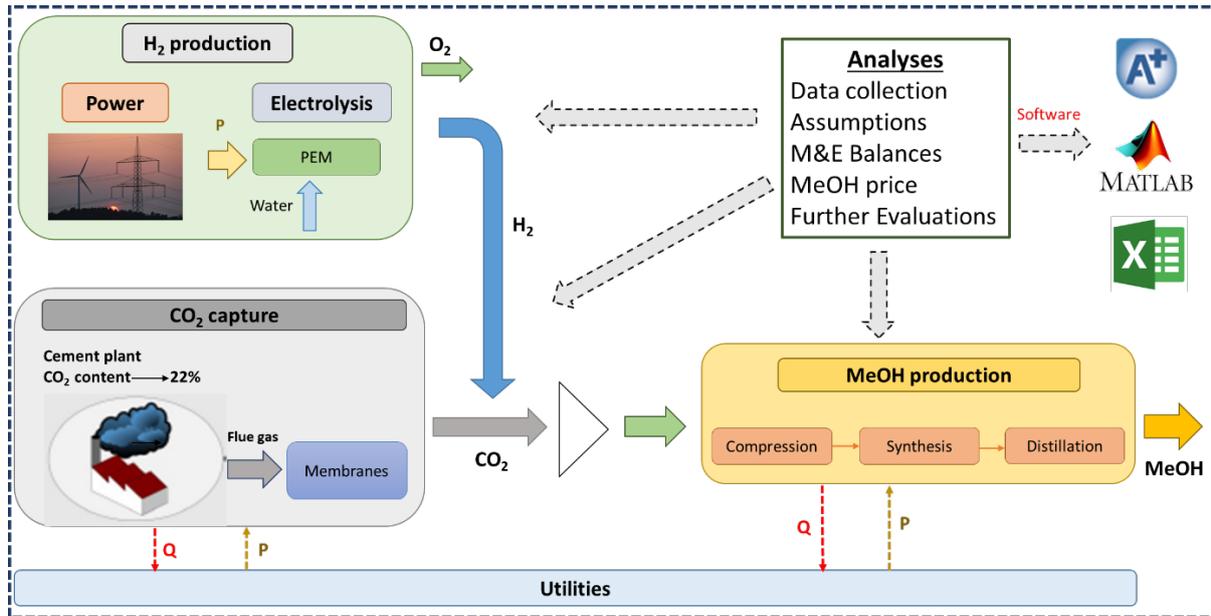


Figure 3. System elements and boundaries of the proposed MeOH plant



Reference to TEA Guidelines

In this section, both unit processes and system elements are presented as recommended in **Guideline B.4**. System boundaries are set to include capture of CO₂ and production of H₂ to ensure transparency in cost calculations. Figure 3 is a visual representation of the incorporated infrastructures and the basic material and energy flows.

B.5.3 Benchmark system

The benchmark system for the proposed power to MeOH configuration is a conventional steam methane reforming MeOH plant utilising as natural gas as feedstock. The scale of the of a typical benchmark plant is 2,500-5,000 t/d. As CCU methanol plants would be likely to follow a distributed model based on feedstock availability, smaller plants are envisaged and therefore the CCU plant is scaled at 2.5-5 times smaller than a benchmark plant (CCU plant scaled at 1000 t/d) [5]. As far as the price of commercial MeOH, Figure 2 reveals an average price of 248 €/t for 2016 [6]. The European contract price for the first quarter of 2018 has been initially settled at 380 €/t [6]. Significant volatilities that can be observed are mainly driven by changes in feedstock prices (natural gas, shale gas, coal) [7]. The Power to MeOH technology can, in theory, eliminate or limit the dependency on these volatilities in sourcing as feedstock prices depend mostly on local, technological and market factors.

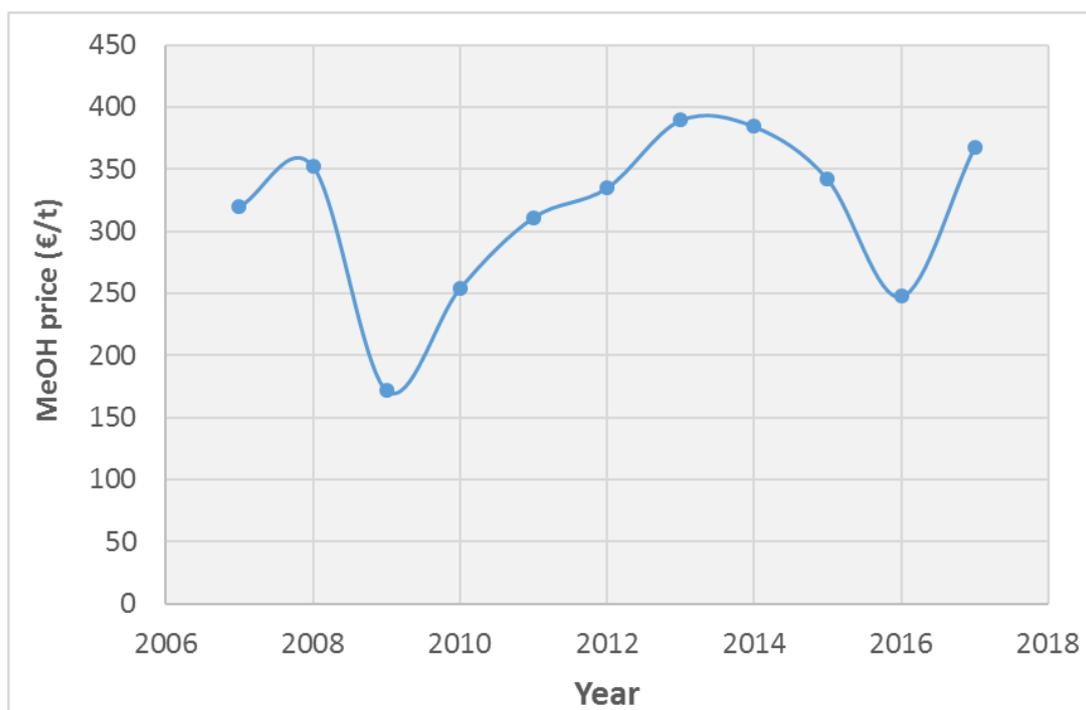


Figure 2. Historical yearly averaged MeOH prices, 2007 to 2017 [6]



Reference to TEA Guidelines

Consistent with **Guideline B.5**, the benchmark system was selected according to the product system function and targets similar group of users. MeOH from natural is the current best class system in the market that provides an identical product and targets similar group of users. As a futuristic scenario, someone may consider as benchmark bio-methanol produced by lignocellulosic wastes.

B.5.4 Technology maturity

The TRL-concept from DoE was used to identify technology maturity of the individual process units. Table 1 provides the TRLs of the individual process units. Membranes are yet to achieve pilot or large scale applications for high purity CO₂ recovery but since preliminary process design/simulation can be accomplished they are characterised as TRL 4. PEM electrolyser in the MW scale have recently become available, *i.e.* TRL 9 (Proton Onsite and Siemens) [8]. The TRL of the MeOH synthesis plant is based on the presence of a plant of the company CRI in Iceland that is operating since 2007; the capacity of the CRI Iceland plant is significantly lower than of a conventional MeOH plant, which is why a TRL 7 is attributed. For the MeOH process from H₂ and CO₂, Perez-Fortes and Tzimas estimate a TRL of 6-7 [9] and Buddenberg *et al.* a TRL of 8 [10].

Table 1. TRLs for each system elements of the process

Process units	TRL
PEM Electrolyser	9
Membrane CO ₂ capture	4
MeOH production	7
Overall plant	4

Even though there are more well-established technologies for CO₂ capture, a low TRL was selected here to showcase ways to treat immature systems.



Reference to TEA Guidelines

The importance of identifying the technology maturity of the investigated configuration is highlighted in this section. The TRL-concept was adapted and as described in **Guideline A.1**, each system element was attributed a TRL. Subsequently, the overall product system TRL equals the lower TRL of the individual units.

B.5.5 Assessment indicators, consistency and reproducibility

Process design along with mass and energy balances are conducted here by means of simulations. Process intensification and heat integration were also carried out in order to enhance the performance of the system. Based on the simulations, a thorough cost breakdown of the produced MeOH is presented, in order to indicate the influence of the capital and operational expenditures on the final MeOH production cost and therefore to identify the most crucial techno-economic parameters that determine the MeOH price. The latter was estimated *via* a typical dynamic profitability calculation; consequently, a break-even analysis was utilised to calculate the minimum MeOH price at which revenues received equals the costs. It should be noted that break-even analysis is a supply-side analysis; it does not analyse how demand may respond at different price levels. Apart from cost breakdown and dynamic profitability calculation, sensitivity and probabilistic analyses (in Matlab environment) on the most significant parameters were carried out, investigating which conditions further influence the MeOH price, aiming to the optimisation of the scheme. For the sake of clarity and reproducibility simulation and financial methods are clearly stated throughout all chapters and/or Appendices. Data for the study came from peer-reviewed academic papers and textbooks and is therefore secondary data.



Reference to TEA Guidelines

As stated in **Guideline B.6**, both economic and technical indicators were considered and calculated with the aim of providing robust answers to the assessment goal question. These comprise mass and energy efficiencies, CapEx, OpEx and MeOH selling price. **Guideline B.7** acknowledges the importance of providing consistent and reproducible ways/methods. To this direction, process conditions and configurations (see Chapter B.4) as well as economic assumptions and data (see Chapter B.5) are clearly and thoroughly described.

B.6. Inventory



Reference to TEA Guidelines

A checklist of items to be included in each section of the report is included in **Chapter B.8 Reporting** of the guidelines, which for the goal is as follows:

Inventory of the study

- State types and sources of the data including the quality
- State the technical in SI units a technical parameter list
- State economic data in an economic parameter list
- State all economic decisions and assumption made

B.6.1 Technical Data

This section presents the process design and simulation of the showcase MeOH plant. For this purpose, Aspen Plus V8.4 simulator was employed to execute the necessary mass and energy balances. Economic evaluations were performed in Excel while probabilistic analysis was carried out in Matlab. A sequential-modular approach was adapted in which the equations describing each process unit (module) were solved module-by-module in a stepwise manner. Iterative techniques were then used to solve the problems arising from the recycle of information. Four models have been employed to determine the thermodynamic properties: Redlich-Kwong-Soave with Huron-Vidal mixing rules for streams at high pressure (>10 bar), and NRTL-RK for streams at low pressure (<10 bar), electrolyte-NRTL for electrolysis reactions and typical steam tables for the CHP units.

For compressors and gas turbines mechanical and isentropic efficiencies are 95% and 90% respectively while for pumps overall efficiency is set equal to 70%. For high pressure, intermediate pressure and low pressure steam turbines, isentropic efficiencies are 92%, 94% and 88% respectively [11]. Pressure drop in heat exchangers is typically dependent on phase. Usually liquid phase pressure drop is absolute and does not depend on relative pressure of the liquid. Liquid phase pressure drop for cold and hot side was set equal to 0.3 bar and gas phase pressure drop for cold and hot side 3% [12]. In addition, a minimum temperature difference of 7°C was applied for heat transfer. If, during heating exchanging, phase change occurs, then at least three heat exchangers are used; two to account for the sensible heat above and below the temperature at which the phase change occurs and one for the latent heat. Open-recirculating cooling systems, that utilise the evaporation process to provide process or comfort cooling, were considered in the present study. A 5% loss was assumed to count for drift, evaporation and blow down losses. The temperature range of cooling water is typically 15-25°C. Furthermore, for a condensate return of 80% (of generated steam), the make-up rate of feed boiler water should be expected to be around 25% of the recirculated rate [13].



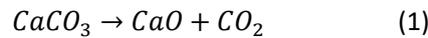
Reference to TEA Guidelines

The current subchapter provides typical process design configurations and assumption. Generic data derived from reliable process engineering literature (high quality) as described in **Guideline B.8** was utilised. This level of detail serves to avoid CapEx and OpEx miscalculations and is in line with **Guideline B.9**.

B.6.2 CO₂ capture

Among the different carbon capture processes, membrane separation generates growing interest. Membrane separation can handle low to moderate dilute CO₂ streams and achieve high purity of CO₂ capture streams. However, based on the existing materials performances, the target purity (>99.5%) cannot be attained by a single or multiple stages [14]. CO₂ can be also separated from other gases by cooling and condensation. Cryogenic separation is widely used commercially for streams that already have high CO₂ concentrations (typically >90%) but it is not used for more dilute CO₂ streams. This suggests that a synergy between the two separations processes is possible and their combination could lead to a possible optimum with a minimal overall requirement. The study focuses on decarbonising heavy industries such as cement production, rather than power generation units as the latter are projected to shut down in the next few years while the market for the former is expected to grow over the foreseeable future.

There are two distinct sources of CO₂ associated with a cement kiln. The first is the result of the direct combustion of fossil fuels for heat energy which produces a flue gas of 4–15 vol% CO₂ depending upon the source of fuel. The second is the by-product of calcining of limestone to produce the clinker material [15], where stoichiometric quantities of carbon dioxide are produced according to the reaction.



Both of these sources emit roughly equal amounts of CO₂ per tonne of cement produced and therefore when combined generate a flue gas that is more concentrated in CO₂ than in the power generation industry [16]. Typical total cement flue gas concentrations range from 22 to 28 mol%, with 900 kg of CO₂ emitted during production of 1000 kg of cement [14].

A single-stage membrane unit cannot achieve a high CO₂ capture ratio (>80%) and CO₂ purity (>95%) simultaneously, as reported by He et al. [17], and energy efficiency could be improved using multiple-stage membrane systems to reduce the irreversibility of the whole process, as documented by Zhang et al. [18]. Based on the design of He et al. [19], a two-stage cascade membrane system was considered in the present study. The first stage membrane unit is used for pre-concentration of CO₂ up to 50–70%. The concentrated CO₂ stream is then compressed to a certain pressure (e.g., 2–3 bar) and fed into the second stage membrane unit for ultimate CO₂ purification to achieve high CO₂ purity (>95%). In order to document the process and economic feasibility of membrane systems, the following assumptions were made:

- 1) A membrane gas separation process based on MRT (Membrane Research and Technology) Polaris™ membrane at a feed and permeate pressure of 2.5 bar and 250 mbar (optimal pressure reported in [17]) was employed,
- 2) a CO₂/N₂ and CO₂/O₂ selectivity of 50 and 20 were achieved [20]. Selectivity of CO₂/H₂O is assumed as unity and
- 3) pressure drop between retentate and feed streams in conjunction with temperature difference between feed and permeate sides due to the Joule–Thompson effect are assumed to be negligible.

For the assumptions made above, the governing equations for the membrane model are as follows.

$$F_f y_{i,f} = F_p y_{i,p} + F_r y_{i,r}, [i = 1, \dots, n] \quad (2)$$

$$\sum_{i=1}^n y_{i,f} = 1, \sum_{i=1}^n y_{i,p} = 1, \sum_{i=1}^n y_{i,r} = 1 \quad (3)$$

$$F_p y_{i,p} = A_m Q_i (P_f y_{i,f} - P_p y_{i,p}) \quad (4)$$

In the above equations, F_f , F_p and F_r are molar flow rates of feed, permeate and retentate streams, respectively; $y_{i,f}$, $y_{i,p}$ and $y_{i,r}$ are mole fractions of the i^{th} component in the feed, permeate and retentate

streams, respectively; P_f and P_p are feed-side and permeate-side pressures, respectively; A_m is the membrane area and Q_i is the permeance (= permeability/effective membrane thickness) of the i^{th} component through the membrane. For a gas mixture of n components, there are $2n + 3$ equations and $4n + 6$ variables in the above model. Of these variables, $F_f, y_{1,f}, \dots, y_{n,f}, P_f$ are for the feed stream and are assumed to be given; Q_1, \dots, Q_n are related to membrane-specific variables to be found from the literature. Then, the remaining number of variables is $2n + 4$ (namely, $F_p, F_r, y_{1,p}, \dots, y_{n,p}, y_{1,r}, \dots, y_{n,r}, P_p$ and A_m). As $y_{1,f}, \dots, y_{n,f}$ are known, mole fraction summation for feed in Eq. (3) becomes redundant. Now, there are $2n + 2$ equations in the model, and so the model has 2 degrees of freedom; for this, the user provides values of A_m and P_p . The algebraic equations in the above model were solved in Matlab environment by utilising the nonlinear system solver *fsolve*. The output component flowrates were transferred to the Aspen plus file via Microsoft's COM technology for software interaction and using Excel as intermediate.

The Matlab solver is a variant of the Powell dogleg method. This algorithm, similar to the well-known Newton-Raphson method requires astute initial guesses. The latter were adopted from Aspen Plus.

Flue gas is assumed to be delivered free of NO_x and SO_x components. Initially, flue gas is cooled below dew point, water is then removed and subsequently compressed to 2.5 bars. Before each membrane, flue gas is conditioned so as to meet the pressure conditions (via vacuum pumps and compressors), monitor temperature at 35°C and where necessary remove condensed water. The permeate stream exiting the second membrane has a 95% CO_2 purity. Vent streams are collected and due to the high pressure of the mixed stream, electricity can be generated in an expander. Prior to this step, heat integration is realised with the aim of increasing the temperature (and consequently the enthalpy content) of the vent stream. For this purpose, three heat exchangers are added as depicted in Figure 4 (red circles) to recover heat from the gas streams exiting, Compr1, Compr2 and Compr3 as well as VPump2. The integrated design reduces the cooling duties by 27.2% and the electricity demand by 32.9%. Furthermore, the total annual cost (TAC) - sum of annualised capital cost (ACC) Eq. (5) and operating expenditures - decreases by 31.6%. Table 2 presents the comparison of the integrated with the initial scenario.

$$ACC = CapEx \times \frac{i_d \times (1+i_d)^n}{-1+(1+i_d)^n}, \quad i_d \text{ is the discount rate and } n \text{ is project lifetime in years} \quad (5)$$

Table 2. Heat integration results for the CO_2 capture plant

	Initial design	Integrated design
Cooling duties (MW)	23.2	16.1
Electricity demand (MW)	11.7	8.97
CapEx heat exchanger and expander (M€)	3.06	3.6
OpEx electricity and cooling water (M€/y)	8.07	6.19
TAC (M€/y)	8.37	6.54

The next step of the carbon capture system configuration involves the cryogenic treatment of the CO_2 rich stream. According to Belaissaoui et al. [21], highly concentrated CO_2 streams (>90%), should be pressurised up to 11 bar before entering the economiser unit; a two stage compressor with intermediate cooling is utilised. Afterwards, the gas is cooled down by water and cooled down further

in the economiser to 73°C. The gas goes through a valve in order to reach the minimal necessary liquefaction pressure. The inlet pressure flash is set in order to have a CO₂ partial pressure of 5.4 bar



Reference to TEA Guidelines

As stated in the Guidelines document, the reported range of CO₂ price is very wide; in some cases it is considered free of charge or under an ETS scheme. To avoid such or similar misconceptions CO₂ capture is within the system boundaries pricing is determined by capture costs. As recommended in **Guideline B.10**, a detailed design is presented including technical information of capture technology and CO₂ source.

ensuring a proper liquefaction of CO₂ and non-frosting conditions. After being heated in the economizer, the incondensable gases are mixed with the retentate streams and sent to an expander to generate electricity while the CO₂ enters a flash unit to remove water and is sent to MeOH synthesis plant. Overall, a CO₂ purity of 99.6% is achieved along with a capture ratio of approximately 87%.

B.6.3 Renewable hydrogen production

Hydrogen is produced *via* electrolysis by passing electricity through two electrodes in water. The water molecule is split, generating oxygen at the anode and hydrogen at the cathode. A PEM electrolyser was selected over an alkaline process as it provides high efficiencies and also serves to separate the hydrogen and oxygen gases, as oxygen and hydrogen are produced on opposite sides of the membrane [22]. Additionally, PEM electrolysers operate very well under variable loads and further efficiency enhancements are anticipated in the coming decade [23]. The membrane allows the H⁺ ion to transfer from the anode side of the membrane to the cathode side, where it forms hydrogen. An efficiency of 70% based on H₂ LHV [24] was considered and the required electricity is derived for the base case study from the grid. An additional 10% of electricity is supplied to cover the demand of the balance of the electrolysis (BOE) section [25]; the overall electrical consumption is 52.4 kWh/kgH₂. The lifetime of the electrolyser is 80,000 h [26]. In order for the process to cause less environmental impacts renewable electricity is a prerequisite and therefore either renewable energy certificates and/or substantial penetration of renewable energy in the grid are required.

The PEM unit was modelled in Aspen Plus by employing two RSTOICH reactors both operating at temperature and pressure of 80°C and 35 bar respectively. The first represents the anode ($\text{H}_2\text{O} \rightarrow 1/2\text{O}_2 + 2\text{H}^+ + 2\text{e}^-$) and the second the cathode ($2\text{H}^+ + 2\text{e}^- \rightarrow \text{H}_2$). A 95% water conversion is realised while the electricity demand was calculated by FORTRAN statements based on the necessary amount of hydrogen. Additionally, it should be noted deionised water is utilised. Oxygen is separated *via* a common separator (100% efficiency was assumed), then cooling to 30°C is applied and subsequently condensed water is removed. Purified hydrogen is recovered and sent to the MeOH synthesis plant. It was assumed that electrolysis occurs on site and no transport is required. Additionally, for a constant electricity load, a constant H₂ production can be assumed and as such no buffering is included for the study.

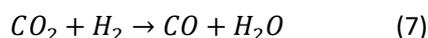
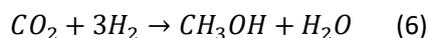


Reference to TEA Guidelines

Guideline 12 raises the issue of H₂ pricing. In this study, hydrogen production is within the system boundaries and thereby cost of production include both associated CapEx and OpEx. Hydrogen is produced via PEM electrolysis; a realistic and environmental friendly scenario as suggested in **Guideline B.11**.

B.6.4 MeOH synthesis

The process design for the MeOH plant is based on previous works conducted by Van-Dal and Bouallou [27], Perez-Fortes *et al.* [28] and Kiss *et al.* [29]. The plant produces MeOH at a rate of 42.78 t/h (1.02 kt/day). CO₂ is delivered to the plant at a flowrate of 62.08 t/h, pressure of 5.4 bar and temperature of 25°C. The required H₂ (3:1 H₂:CO₂ molar ratio) produced from the PEM unit is supplied at a pressure of 35 bar and temperature of 25°C. The first step of the process incorporates the conditioning of the reactant gases to the operating conditions of the synthesis reactor, *i.e.* 78 bar and 210°C. CO₂ is compressed *via* a multistage compressor consisting of two compressors with intermediate cooling. If the gas is cooled to the inlet or slightly higher temperature of the previous lower pressure compressor (~35°C), it can be shown that minimum specific work is consumed when the pressure ratios (3.8:1 in this case) of the compressors are equal [30]. A single stage compressor was utilised to raise H₂ pressure up to 78 bar. The two gases are mixed and then re-mixed (for convergence purpose) with the recycle stream. The stream is then heated to 210°C and injected into the fixed bed adiabatic reactor. The adiabatic reactor is packed with a fixed bed of 44.5 tonnes of Cu/ZnO/Al₂O₃ commercial catalyst and it assumed that it is renewed once a year [28]. For this catalyst, the model proposed by Vanden Bussche and Froment [31] with readjusted parameters of Mignard and Pritchard [32] is able to describe with good precision the reactions of MeOH production, Eq. (6) and the RWGS reaction, Eq. (7).

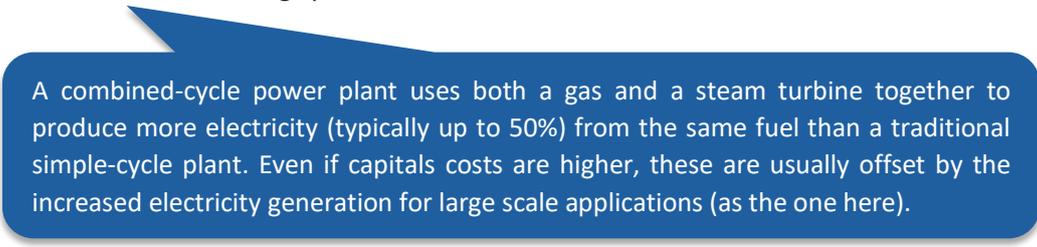


The kinetic model is described with Langmuir–Hinshelwood kinetics and it can be found in [27]. The model was directly implemented in Aspen Plus and pressure drop was calculated by the Ergun equation, already nested in the simulator. For a pellet density of 1,775 kg/m³ and a void fraction of 0.5 the volume of the reactor is approximately 50 m³. It was observed that steady state can be reached at lower volumes meaning that less amount of catalyst can result in same MeOH productivities, *i.e.* 35 tonnes. The gases leave the reactor at a temperature of 288°C and subsequently is divided into two streams: the first (60% of initial stream) is used to heat the fresh feed, while the second is used in the reboiler and also to heat the feed of the distillation column. The two streams are re-mixed and cooled to 35°C. Water and MeOH, after being condensed, are separated from the non-reacted gases in a knock-out drum. Some of the non-reacted gases (1.3%) are purged to minimise the accumulation of inerts and by-products in the reaction loop. The crude MeOH that leaves the reactor is a mixture of MeOH, water and residual gases (*i.e.*, H₂ and CO and CO₂). To remove the non-reacted gases, the stream is expanded to 1 bar using valves, and then separated in a flash tank. The remaining liquid is heated to 75°C and fed into the distillation column. The bottom product of the distillation column corresponds to water while the top product is mostly MeOH with some unreacted gases. The column was simulated with the rigorous model RadFrac. A rate-based calculation of the column is carried out, using the mass and heat transfer model from Billet and Schultes. The column has 50 stages and the reflux ratio is equal to 1.2. However, the column can be intensified by reducing the reflux ratio to 1; reboiler's heating duty is reduced to 16.5 MW from 19.5 MW while MeOH recovery and purity remain unchanged. MeOH is then compressed to 1.2 bar and cooled to 40°C proceeding to another flash that separates the gases (top outlet) from the MeOH product with 99% w/w in the bottom stream. Figure 5 provides the process flow diagram of the MeOH plant including the electrolysis unit.

B.6.5 Utilities

The streams containing unreacted gases are collected and sent to a gas turbine unit to generate electricity. The gas stream is compressed up to 6 bar and enters a gas turbine where it is burned with excess of pressurised air (15%) to produce electricity at a temperature of 1200°C. The required air is specified by a FORTRAN calculator according to the flows of carbon monoxide and hydrogen. The exhaust gas from the

gas turbine is recovered from a heat recovery steam generation (HRSG) system, composed of three heat exchangers (namely economizer, evaporator and superheater) where superheated steam (550°C, 100 bar) is produced. According to the present design, the economiser is placed before HE1 to extract heat from the pressurised CO₂. This way more steam can be produced as the exhaust gas provides heat only to the evaporator and the superheater. A series of high, intermediate and low pressure steam turbines are employed to generate electricity. Part of high pressure steam is utilised to provide heat to the deaerator. The combined Rankine-Brayton cycle generates 4MW of electricity. Furthermore, based on the design of Van-Dal and Bouallou [27], hot water at a temperature of 85°C that is raised from HE1, HE2, and H4 is mixed with the bottom stream of the distillation column and enters an Organic Rankine Cycle (ORC) unit. The working fluid is the R245fa and electricity of 1MW is generated. Finally, it should be noted that the wastewater unit and the cooling system were not simulated but considered in the economic evaluation.



A combined-cycle power plant uses both a gas and a steam turbine together to produce more electricity (typically up to 50%) from the same fuel than a traditional simple-cycle plant. Even if capitals costs are higher, these are usually offset by the increased electricity generation for large scale applications (as the one here).

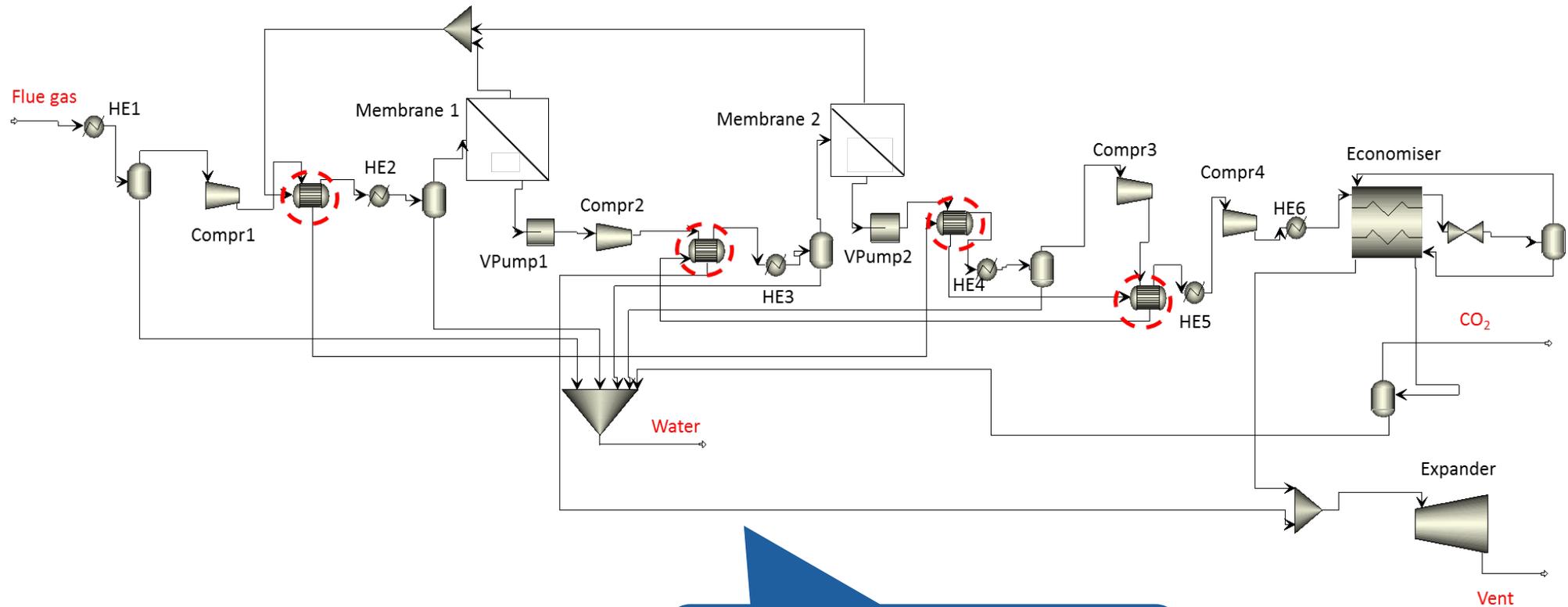


Figure 4. PFD for CO₂ capture through multistage membrane

The membranes were modelled as common separators and the output flowrates were defined in Matlab software as described in the main text. The economiser is fully integrated.

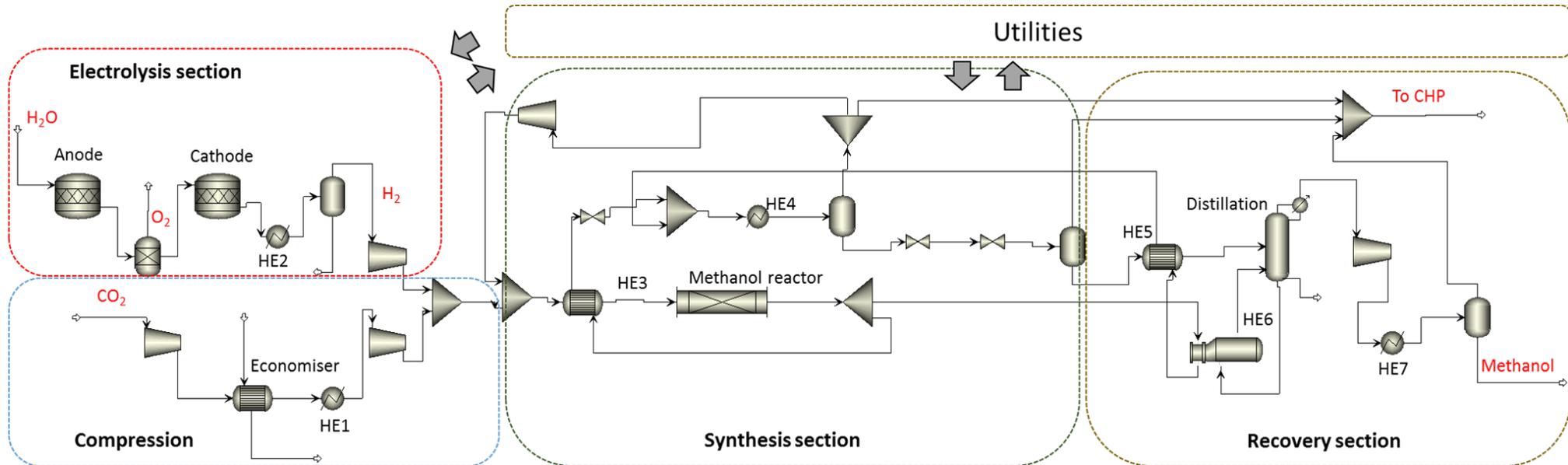


Figure 5. MeOH plant process flow diagram



Reference to TEA Guidelines

Generic data and sensible process design (see **Guideline B.8**) was employed to simulate the MeOH production and utilities system elements. Collection of high quality data (see **Guideline B.9**) enabled the intensification of the process. As recommended in **Guideline B.12**, technical parameters are documented in SI-Units in the Inventory section and throughout the manuscript.

B.7. Calculation of Indicators



Reference to TEA Guidelines

A checklist of items to be included in each section of the report is included in **Chapter 8 Reporting** of the guidelines, which for the goal is as follows:

Inventory of the study

- State calculation procedures
- Explain methodology of financial analysis
- Include results of technical assessment
- Include results of economic assessment

The previous chapter served to provide all the necessary data and assumptions for the assessment phase of the present study. Based on the simulations and the given process design material and energy streams were quantified and subsequently technical indices were calculated. As previously discussed, with the aim of determining the attractiveness of the investment a DCFA was carried out to estimate the MeOH price. Technical and economic indicators should go hand in hand. However, the economic indicators are those which support decision making; technical related results have the potential to identify hot spots and insights on why a project is profitable or not.

B.7.1 Technical assessment indicators

Given the process technologies modelled and integrated, mass and energy efficiencies were estimated for the proposed MeOH process configuration. The basic simulation outcomes can be found in Appendix A. As depicted in Table 3, some of the major indicators include specific energy consumptions, overall and per pass (around the MeOH reactor) CO₂ conversion. The overall and the per pass CO₂ conversions are defined as $(CO_{2,in} - CO_{2,out}) / CO_{2,in}$ around the MeOH plant and the reactor respectively. For a plant utilising 1,490 t/d of CO₂, a MeOH yield of 1,000 t/d is realised.

Table 3. Technical indicators

Indicator	Value	Unit
Overall CO ₂ conversion	94.65	%
Conversion factor	1.45	tCO ₂ /tMeOH
Per pass CO ₂ conversion	24.8	%
Specific energy consumption for CO ₂ capture	0.145	MWh/t CO ₂ captured
	0.21	MWh/t MeOH
Specific energy consumption for electrolysis	52.4	MWh/t H ₂
	10.52	MWh/t MeOH
Specific energy consumption for MeOH plant	0.11	MWh/t MeOH
Overall energy consumption	10.84	MWh/t MeOH

Since methanol is treated as chemical, it is more sensible to report energy consumptions rather than energy efficiency. The latter considers the energy content of the final product in terms of heating value and as so it is a suitable metric for fuels production. However, for the sake of comparability the energy efficiency of the system on a LHV basis is 58.1%.



Reference to TEA Guidelines

As suggested in **Guideline B.13**, this section provides the necessary formulas for calculating indicators, uses common units and presents metrics for both the entire system and the individual system elements. Furthermore, simulation data (input/output) is stored and presented in Appendix A.

B.7.2 Basis for cost estimation and DCFA

The current section presents the utilised methodology for the estimation of capital and operating expenditures (CapEx and OpEx). In addition, all the required data and assumptions for the conduction of a typical DCFA are presented.

The location of the proposed plant is Germany; a location factor of 1.1 [33] was utilised to represent the difference in cost between building a plant in Germany versus building a plant in the US Midwest. The reference year of the study is 2016 and the currency is EUR. Where capital costs were reported at different currency than EUR the value was first converted to EUR by using exchange rates of the respective reported year and were subsequently brought up to date by utilising the Chemical Engineering Plant Cost Indices (CEPCI).

The scaling exponent method is utilised to estimate equipment costs. Eq. (8) represents the economies of scale because buying a piece of equipment with twice the size or capacity is less than twice as expensive (when the exponent is less than 1.0).

$$C = C_0 \left(\frac{S}{S_0}\right)^f \quad (8)$$

Where C is the estimated actual cost of the unit, C_0 the base cost of the unit, S the actual size or capacity of the unit (extracted from the simulations), S_0 the base or capacity and f an empirical scaling exponent. The values of C_0 , S_0 and f can be found in Appendix B (Table B.1).

Then, the factorial method is employed to estimate CapEx. The latter includes ISBL, OSBL, indirect costs and working capital. As mentioned before, the OSBL counts only for the BOP and thereby a relatively lower cost value was attributed, *i.e.* 12% of the ISBL [34]. For the electrolysis unit a value of 1.5 M€/MW installed was utilised. This figure includes the balance of the electrolyser (BOE) and successively a factor of 1.524 [35] was employed for the fixed capital investment (both ISBL and OSBL). Based on confirmed information from sources in industry, the DoE reports a value between 0.6 and 0.7 for the cost exponent [36]; a value of 0.65 is considered here. The full methodology is provided in Appendix C (Table C.1).

The cost estimation described here incorporates what are known as “ n^{th} -plant” economics. The chief hypothesis involved in n^{th} of a kind (NOAK) economics is that the analysis does not designate a pioneer plant; as an alternative, several plants using same technologies have been already built and operating. It should be noted here that for the high TRL units of the plant (electrolysis and MeOH synthesis) NOAK costs were calculated directly while for the low TRL CO₂ capture unit and indirect method was implemented. The indirect method involves the following three steps (1) estimate the equipment and installation cost, (2) find the FOAK total plant cost costs by adding appropriate engineering costs, process and project contingencies, and (3) find the NOAK total plant cost using learning curves [37, 38]. The NOAK costs can be calculated from the FOAK using a single factor learning curve [39]:

$$C_{NOAK} = C_{FOAK} \left(\frac{N_{NOAK}}{N_{FOAK}}\right)^b \quad (9)$$

The learning rate coefficient b is a parameter that is calculated from $LR=1-2^b$, where LR is the learning rate. This single factor learning curve combines learning by doing, learning by searching (continued RD & D in the commercial stage of technology deployment), and scale factors. For MEA technology, a learning rate of 0.11 was proposed by Rubin [38]; since advances in membrane technologies are expected to be sharper than MEA, a higher LR was assumed, *i.e.* 0.15. N is the number of installed plants. According to Greig *et al.* [40] N_{FOAK} is typically less than 10 (a nominal value of 5 was used in the present study), early movers range between ten and twenty, and N_{NOAK} more than twenty (20 was selected).

In addition, an overall scaling factor (see Figure 6) was calculated by setting as design variable MeOH productivity in t/d. In this way, potential future designers can utilise the overall factor instead of individual factors. The limitation of this approach is that the technologies involved should be the same or at the very least similar. A scaling factor of 0.669 was calculated here which is in line with the figure reported by Couper [41], *i.e.* 0.7

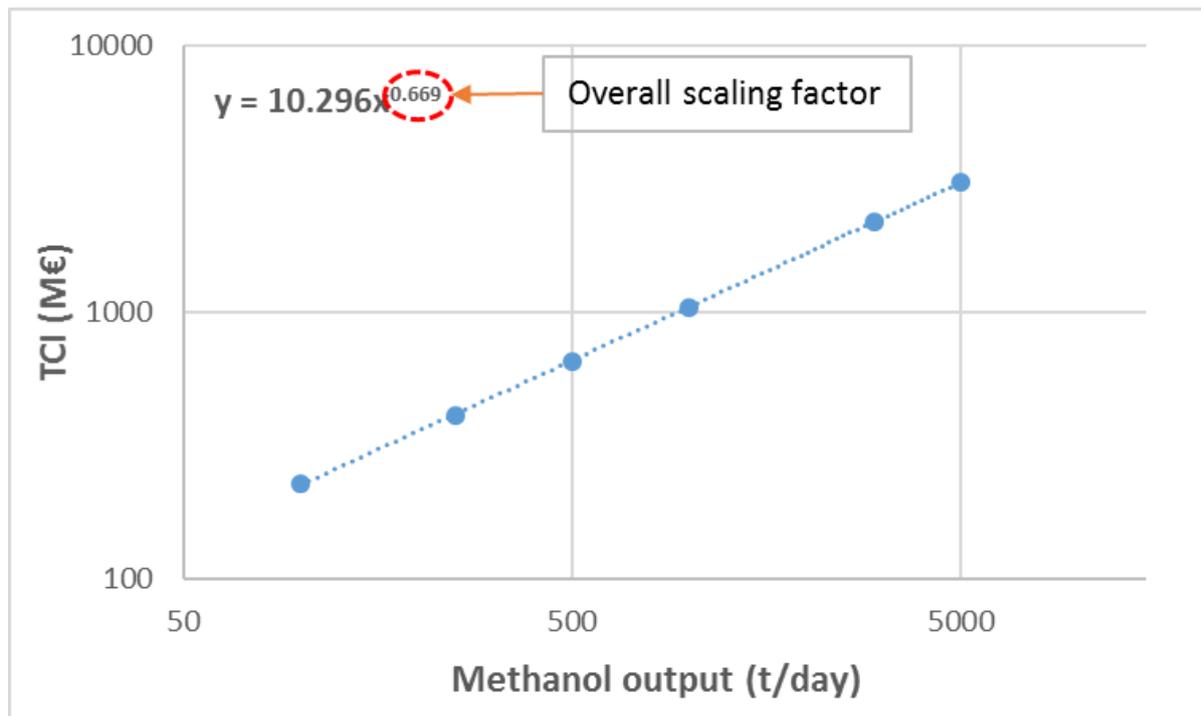


Figure 6. Economies of scale of the proposed power to MeOH plant (includes both ISBL and OSBL as defined in the present study)

The calculation of the overall scaling factor depends on the individual equipment factors and their contribution to CapEx.

OpEx comprise variable costs and fixed costs. The former expenditures refer to raw materials price, catalyst costs, utilities and waste disposal (see Appendix B, Table B.2) while the latter comprise labour, supervision, direct salary overhead, maintenance, insurance and general plant overhead (see Appendix B, Table B.3). For the labour requirement, the following correlation was utilised (refers to fully automated processes) [42]:

$$N_{OL} = (6.29 + 31.7P + 0.23N)^{0.5} \quad (10)$$

Where P the number of the solids handling steps and N is the number of nonparticulate processing steps and includes compression, heating and cooling, mixing, and reaction. For each of the N_{OL} operators per 8-

hour shift, approximately 5 operators must be hired for a plant that runs 24 hours per day, to account for 3 shifts per day as well as regular and sick annual leaves. An average annual salary of €85,000 is considered [43].

The plant is financed with a 50-50 debt-equity split, considering an interest rate of 6% for the debt and assuming a cost of equity of 12%. The weighted average cost of capital (WACC) is calculated as follows [44]:

$$WACC = DR \times i_d + [(1 - DR) \times i_e] \quad (11)$$

Where DR is the debt ratio, i_d interest rate due on debt and i_e cost of equity. The WACC can serve as a good proxy for the discount rate and this approach was adopted here for budgeting the project and carrying out the DCFA. The capital investment is assumed to be spent over a 3-year construction period, with 20% in the first year, followed by 50% and 30% for the second and third years respectively. Working capital is applied in the year before operation and recovered at the end of the plant life. A straight line depreciation was realised and assets are depreciated in ten years whereas the project life time is 20 years. Any salvage value was assumed to be fully offset by decommissioning expenses. The basic economic parameters and assumptions are presented in Table 4.

Table 4. Economic parameters and assumptions

Parameter	Value
Location	Germany
Base year	2016
Project lifetime (y)	20
Construction period (y)	3
Plant availability (h/y)	8,000
CHP capacity factor (%)	85
Tax rate (%) [45]	29.72
Equity/Debt (% / %)	50/50
Debt interest (%)	6
Cost of Equity (%)	12
WACC (%)	9
Depreciation	straight-line
Depreciation period (y)	10
Salvage value (€)	0

The capacity factor should not be confused with the plant availability. The capacity factor is the unitless ratio of units of electricity generated / units the plant can generate if in 100% operation. So in our case the nameplate capacity is $5/0.85=5.88\text{MW}$



Reference to TEA Guidelines

This section along with Appendices B and C is aligned with **Guidelines B.14 and B.15**. Consistent with Table 9 of the guidelines document, Table 4 presents all the necessary economic data. Furthermore, the full numerical methodology for estimating CapEx an OpEx is provided. For the low TRL section an indirect method for CapEx estimation was utilised while for the high TRL a direct one. Cost transformation was utilised for costing typical equipment (high quality data) and factorial methods for FCI and TCI (Appendix B and C respectively). Variable OpEx were estimated based on the quantification of material and energy streams while fixed OpEx follow a factored estimation.

B.7.3 Economic assessment indicators

The goal is to calculate CapEx and OpEx for each segment of the proposed plant, total production costs and the minimum selling price of MeOH. Results are presented in Table 5 and Table 6 depicts comparisons with the benchmark process. On the basis of the fixed capital investment cost and the cash flows at each successive year, the NPV of the project is calculated. Revenues are generated from MeOH and O₂. The two widely used measures for evaluating an investment are based on calculating NPV and the internal rate of return (IRR). The NPV and IRR are calculated by using the following equations:

$$NPV = -TCI + \sum_{n=1}^{20} \frac{CF}{(1+discount\ rate)^t} \quad (12)$$

$$\sum_{n=1}^{20} \frac{CF}{(1+IRR)^n} = 0 \quad (13)$$

$$CF = P(1 - t) + Dt \quad (14)$$

Where the *CF* values are the after tax cash flow for each year, *P* are gross profits, *t* is the tax rate and *D* the depreciation. The MeOH price is calculated for NPV=0 or equally IRR=discount rate. This value is also referred as minimum MeOH selling price (MMSP) and accounts for the price that can cover the full costs of the investment. The DCFA resulted in a MMSP of 1,402 €/t. Furthermore, Figures 7 and 8 provide detailed breakdown of purchased equipment costs and total operating costs respectively. Figure 9 is the breakdown of the major OPEX contributor, *i.e.* electricity, across the three unit blocks.

Table 5. Economic indicator results of the plant

	CO ₂ capture plant	Electrolysis	MeOH synthesis	Utilities	Total
CapEx (M€)	258.06	140.90	403.12	247.21	1,049.28
OpEx (M€/y)	18.08	328.89	20.67	9.23	376.87
CapEx per tonne of MeOH (€/t)	82.60	45.10	129.03	79.13	335.86
OpEx per tonne of MeOH (€/t)	52.82	961.00	60.39	26.97	1,101.17
Production cost	135.42	1006.10	189.43	106.09	1,437.04
Revenues from O₂ (€/t MeOH)	-	-	-		86.41
Production cost (€/t MeOH)	-	-	-		1,350.63
Minimum selling price (€/t MeOH)	-	-	-		1,402

Table 6. Comparison of technical and techno-economic indicators of the proposed plant with weighted-average conventional synthesis plant in Western Europe [9]

Indicator	Proposed plant	Conventional plant
Electricity needs (MWh/tMeOH)	10.84	0.147
Cooling water needs (tH₂O/tMeOH)	128	90
CapEx per tonne of MeOH (€/t/y)	3,066	862
OpEx per tonne of MeOH (€/t)	1,101	407



Reference to TEA Guidelines

As suggested in **Guideline B.3**, economic results are presented on a functional unit basis, *i.e.* metric tonnes of MeOH. For CO₂ capture (low TRL) energy consumption and subsequently OpEx can give better insights on the performance of the process. The full methodology, equations and assumptions to conduct an NPV analysis is presented here as recommended in **Guideline B.16**.

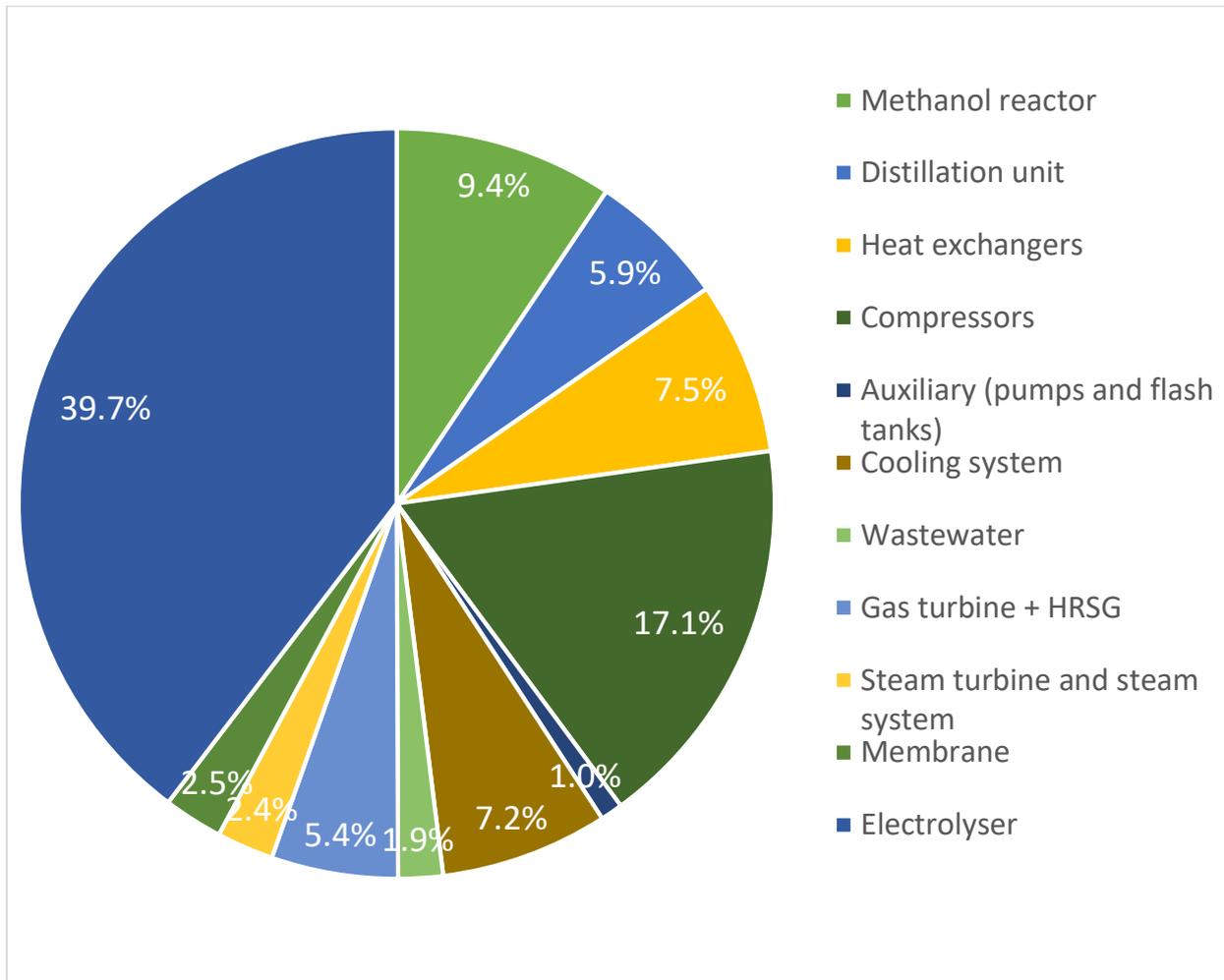


Figure 7. Purchased equipment cost breakdown

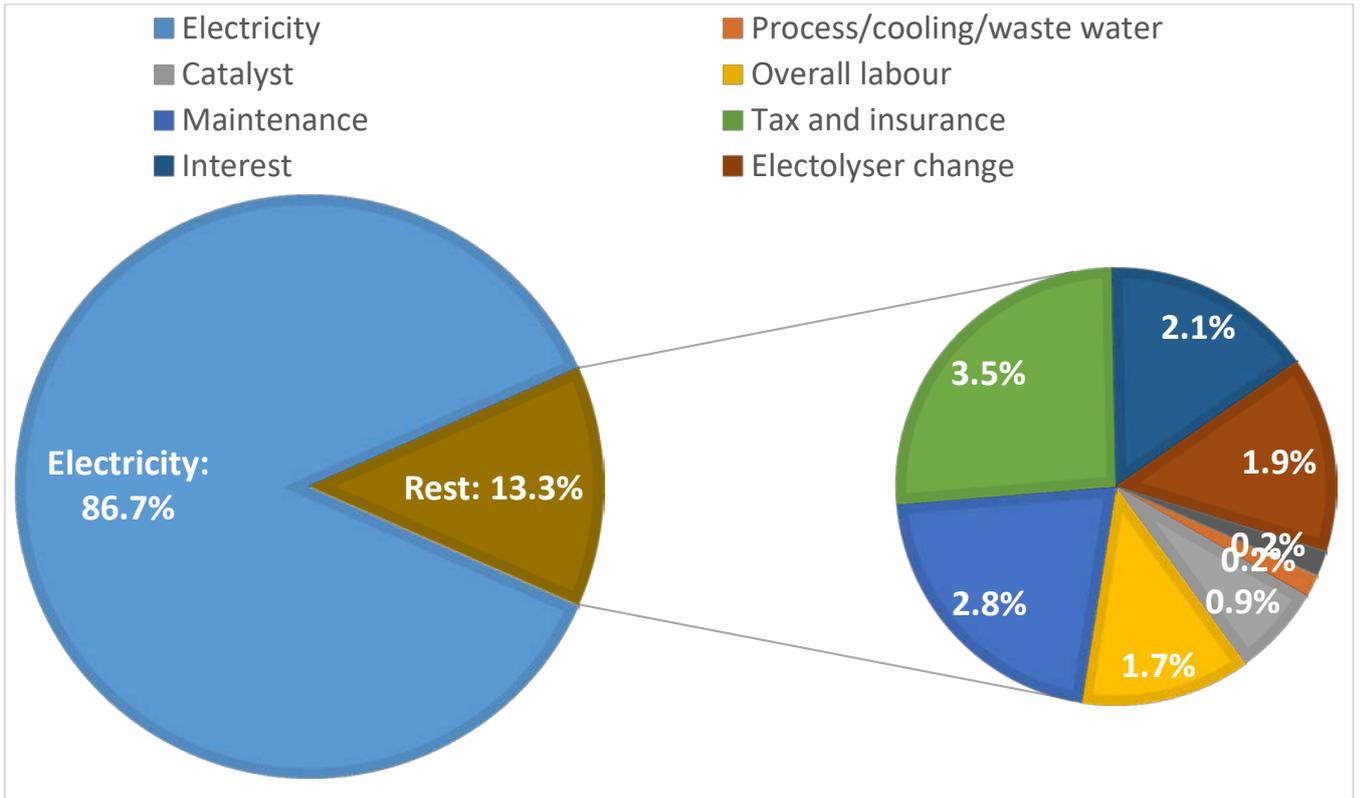


Figure 8. OPEX breakdown

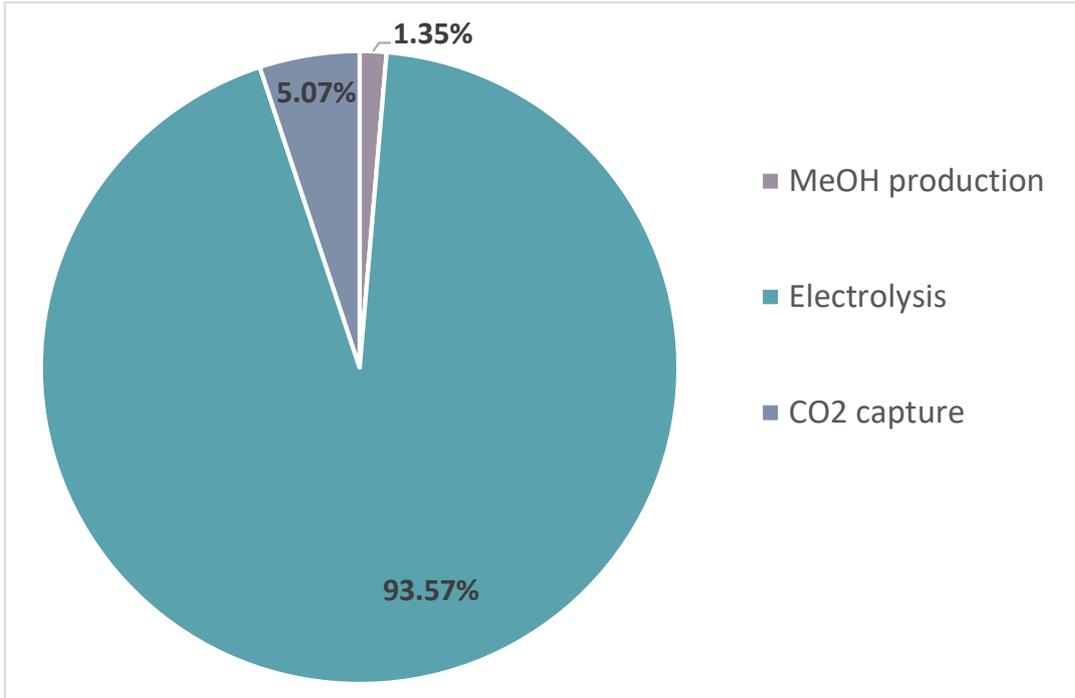


Figure 9. Electricity cost breakdown

The electrolyser is the main contributor to CapEx accounting for 40% followed by gas pressure change equipment. The cost of compressors is relatively high because of the high pressure ratios (up to 10) in the CO₂ capture unit. In addition, electricity costs dominate OpEx and it can be concluded that the major cost drivers are associated with the electrolysis unit.



Reference to TEA Guidelines

A checklist of items to be included in each section of the report is included in **Chapter 8 Reporting** of the guidelines, which for the goal is as follows:

Interpretation of the study

- Include and describe the results
- Include and describe uncertainty and sensitivity analysis
- State assumptions and limitation associated with the assumptions, methods and interpretation of results
- Include conclusions
- Include recommendations, if any

B.8. Interpretation

B.8.1 Sensitivity analysis

A local sensitivity analysis was carried out for the proposed plant configuration to identify and rank the plant parameters that most influence the MeOH minimum selling price. The parametric analysis is carried out by changing one input variable to its low or high value while keeping all other variables at their nominal values. The low and high values were selected based on nominal variations or where possible, estimates and projects. Table 7 illustrates the independent parameters along with their respective lower and upper bounds. A highly optimistic scenario was included for electricity price whereby electricity is supplied for free (e.g. to avoid renewable energy curtailments) and for oxygen price a $\pm 30\%$ fluctuation can reflect market uncertainties. Electrolyser cost is expected to drop in the near future [8, 22] down to 0.5 M€/MW and technological advances can increase the efficiency up to 85% [46]. In addition, as electrolyser is the hot spot of the process, the capital cost exponent factor of the electrolyser was fluctuated. Due to insufficient relevant data, the electricity demand for the BOE is also included. According to the AACE International classification, the present CO₂ capture plant is of “Class 5 estimate” and an accuracy in the range of -30% to +50% should be anticipated where the MeOH and utilities plants can be categorised as “Class 4 estimate” with an accuracy level of -20% to +30% [47]. Significantly, increased accuracy can only be achieved through acquiring capital cost data from a commercial plant which was not possible. Furthermore, the effect of typical economic parameters, such as tax and interest rates, was investigated.

Table 7. Selected parameters for sensitivity analysis and their bounds

Parameter	Low value	Nominal	High value	Unit
Electricity price	0	87.7	114	€/MWh
Electrolyser efficiency	60	70	85	%
Electrolyser scaling factor	0.55	0.65	0.8	-
BOP electricity demand	5	10	20	%
Oxygen price	37.94	54.2	70.46	€/T
CAPEX Electrolysis	0.5	1.5	2	M€
CAPEX MeOH	322.5	403.12	524.06	M€
CAPEX CO ₂ capture	180.64	258.05	387.08	M€
CAPEX Utilities	197.76	247	321.37	M€
Interest rate	5	6	8	%
Tax rate	0	29.72	40	%

As depicted in **Figure 10**, three levels of uncertainty can be observed, *i.e.* high, intermediate and low. Electricity price and electrolyser efficiency raises a significant level of uncertainty while overall capital investment along with the electricity demand for the BOE and the tax rate have a medium effect on the MMSP. Interest rate as well as oxygen price have negligible effect on the MMSP. Overall, it is shown that the most crucial block is the electrolysis unit; in fact, changes in electricity price cause a wide variation to MeOH price of -68.1% to +20.3% and electrolyser efficiency causes the price to range between -11.5% to +15.6% from the nominal value. If excess of electricity can be generated by renewable sources such as wind turbines and that excess is provided free of charge, then the MMSP reduces drastically (447€/t); however still ~ 1.8 times higher than the fossil MeOH price. On top of this, this scenario will raise competition with other technologies and should not be taken for granted.

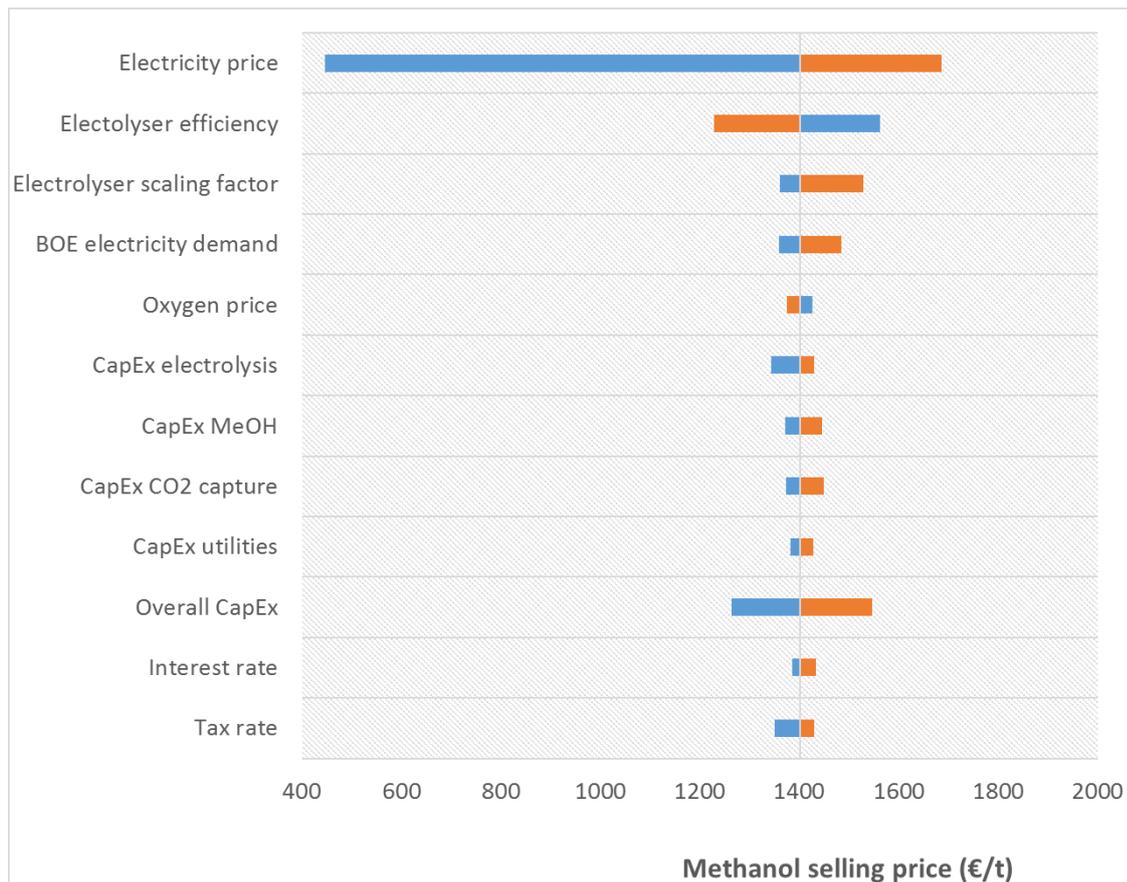


Figure 10. Sensitivities on MeOH price. When blue bar is on the left a positive correlation is depicted and vice versa

Rather than providing symmetric bounds to the investigated parameters, ranges were selected so as to provide realistic reflections of techno-economic estimates and/or projections

B.8.2 Uncertainty analysis

A limitation of local sensitivity analysis (LSA) is that uncertain variables typically are examined one at a time with all other parameters held constant. Thus, interactions among several uncertain parameters may be overlooked. In some cases, several parameter values may be changed simultaneously, for example, as a bounding analysis with parameters set to their maximum or minimum values. However, this provides no information on the likelihood of such extreme outcomes. A complementary approach to LSA is a probabilistic analysis in which distribution functions are assigned to multiple independent variables. The distributions are sampled repeatedly using Monte Carlo (or related) methods to yield a distribution function showing the probability of a specified outcome or result. After identifying the dominant parameters, a Monte Carlo simulation was applied to characterize the effect of uncertainty or variability of these parameters on MeOH price. The investigated parameters are those mentioned above with the greater/medium effect on MeOH price. In order to avoid probable correlations, CapEx for electrolysis and MeOH plant were excluded from the analysis in conjunction with BOE electricity demand; the latter was expressed as a portion of the electrolysis demand and as so it is correlated with the electrolyser efficiency.

Therefore, the manipulated variables are electricity price, electrolyser efficiency, tax rate and CapEx. Regarding CapEx, the electrolyser cost was included in the analysis in terms of its scaling factor and the CO₂ capital costs as it carries higher uncertainty due to low TRL. Based on the practitioner's understanding of the variables in question, the shape for the probability distribution can be selected. Uniform distribution suggests that all values have an equal chance of occurring, and the user simply defines the minimum and maximum; in the present study parameters related to manufacturing costs were assigned uniform distributions. Another option is the triangular distribution where the user defines the minimum, most likely, and maximum values. Values around the most likely are more likely to occur and the results of such distributions tend toward the normal or log-normal distributions (preferred by purists). Electricity price, electrolyser efficiency and tax rates were assigned triangular distributions as the deterministic values are the most probable as well.

Figure 11 depicts the cumulative probability functions for two cases; Case 1 reflects the assumed distributions for parameters such as the electricity price which is non-symmetric relative to the nominal deterministic value and is skewed to the left (lower prices) while Case 2 assumes symmetric bounds. The MMSP for Case 1 has mean value of 1,356 €/t and a 95% confidence interval of approximately 1,238 €/t to 1,448 €/t. The mean value of Case 2 is closer to the deterministic value, *i.e.* 1400 €/t and the range of a 95% confidence interval is narrower, 1,343 to 1,458 €/t. The steepness of the Case 1 curve suggests a smaller value of variance compared to Case 2. As a result, the probability that the MMSP is lower than the deterministic estimate is much higher for Case 1 than Case 2, *i.e.* 77% and 51% respectively.

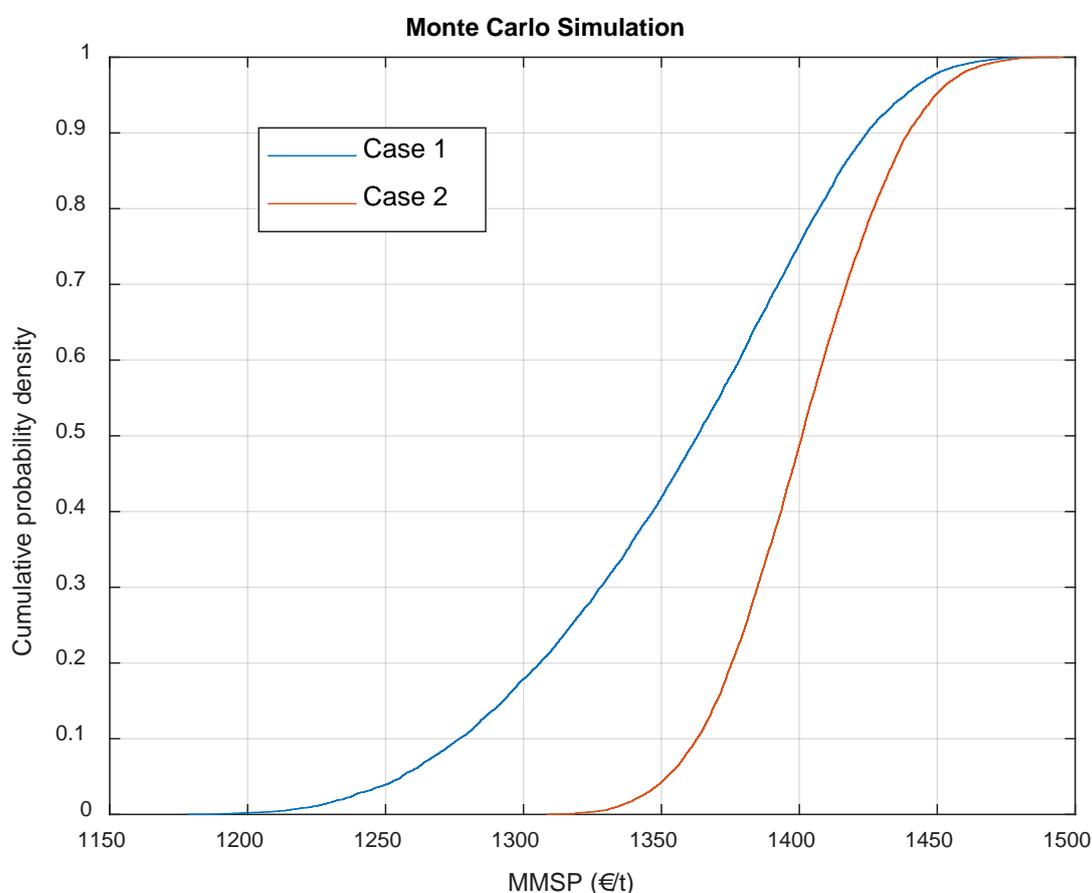


Figure 11. Monte Carlo simulation on MMSP



Reference to TEA Guidelines

As suggested in **Guideline B.18**, quantitative local sensitivity analysis was conducted in conjunction with uncertainty analysis (Monte Carlo simulations). The purpose of the former was to identify key variables and those were subsequently utilised in the UA. As design variable the MMSP was set and confidence intervals were identified.

B.8.3 Development of alternative scenarios

The base case scenario of the proposed Power to MeOH assumed a green electricity load from the grid equal to approximately 464 MW. Furthermore, sensitivity analysis revealed that electricity price is the chief cost contributor. Renewable energy in Germany is mainly based on wind and solar. As for 2016, the total installed nameplate capacity of wind energy in Germany was 49,600 MW while the respective number for solar power was 40,720 MW [48]. These figures imply that at the current status the proposed plant would have to consume approximately 2.4% of the wind energy or roughly 3.8% of the solar; assuming a capacity factor of 40% for wind turbines and 30% for photovoltaics [49]. Even if these values can be considered relatively low and can be covered by the existing renewable energy infrastructure, in order to avoid distortion of the current applications and competition with other markets, scenario analysis was performed by expanding the system boundaries. Therefore, two alternative scenarios were developed considering the production of renewable energy within the system boundaries. Renewable electricity is supplied in both cases from wind energy either from an onshore (Scenario 1) or an offshore (Scenario 2) farm. It should be highlighted that the turbines provide power only to electrolysis section (=450 MW) in order to avoid dynamic operation of the MeOH and CO₂ capture units. According to Table 8 (calculation basis can be found in [49]), the installed capacity for the onshore capacity should be 1,128 MW and for the offshore 1,095 MW. Furthermore, the electrolyser was resized in order to exploit the power generated at the maximum achieved capacity factor. A challenge for any Power to X system is the temporary storage of hydrogen necessitated by the fluctuating power supply and consequent intermittent operation of the electrolyser. According to Gahleitner [50], hydrogen in high pressure gas cylinders is the current and future method of choice. The storage unit was not designed or simulated but was attributed a cost. Rough cost estimations for storage tanks can be found in [8] and suggest that for steady state operation of the hydrogen utilisation plant, storage expenditures can be as high as 28% of the electrolyser cost.

Table 8. Summary of the Land-Based and Offshore wind farms. Costs were brought up to date considering an inflation rate of 2.5%

	Onshore	Offshore
Turbine capital cost (€/MW)	1,140,335	1,382,739
Balance of system (€/MW)	311,258	2,043,925
Financial costs (€/MW)	142,424	927,171
Normalised CapEx (€/MW)	1,594,016	4,353,834
CapEx (€)	1,797,762,914	4,766,971,839
Net capacity factor (%)	39.9	41.1
Maximum capacity factor (%)	95	95
Normalised OpEx (€/MW/y)	48,103	183,475
OPEX (€/y)	54,252,017	200,885,036

The data of Table 8 was integrated in the economic model of the base case scenario. As depicted in Figure 12, a trade-off between CapEx and OpEx can be observed between the base case study and the alternative scenarios. The MMSP for the Scenario 1 is marginally lower than the case study, *i.e.* 1380 €/t and 1402 €/t respectively, showcasing benefits associated with minimizing electricity costs; however, the high initial investment can be unattractive for decision makers and/or potential investors. The offshore wind energy scenario come at very high costs mainly due to immature technologies and development constraints. It should be noted that a crucial factor, that has not been taken into consideration here, for the penetration of wind energy is land availability.

This section serves to demonstrate the effect of expanding the system boundaries rather than provide a realistic scenario. In particular for Germany, a Power to X plant is more probable to be installed in the northern area as the existing power infrastructure might be unable to transport the energy and exploit offshore wind energy.

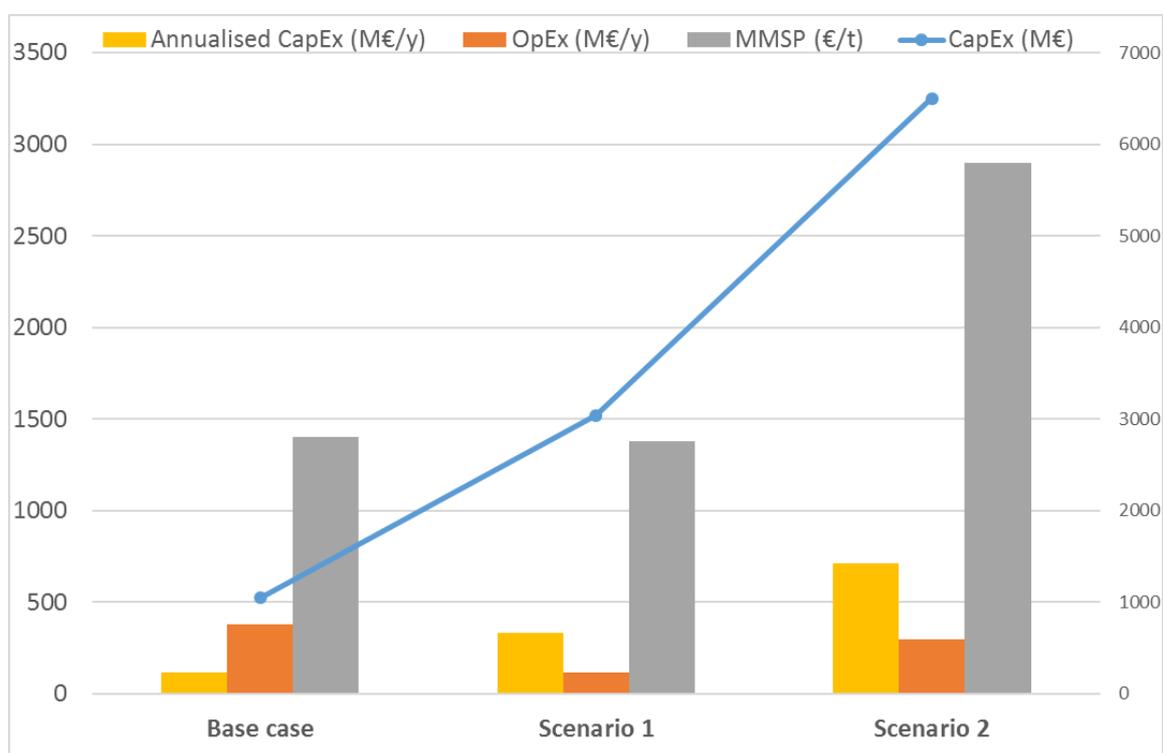


Figure 12. CapEx (right axis), Annualised CapEx OpEx and MMSP for the investigated scenarios



Reference to TEA Guidelines

As suggested in **Guideline B.2**, scenario analysis was conducted based on hot spot identification. Electricity price poses as the major cost driver and consequently alternative scenarios investigating different electricity supplies were developed.

B.8.4 Interpretation of Indicators

The designed plant in the present situation does not produce any profit. The revenue obtained from selling MeOH and O₂ does not cover the cost of the electricity needed for generating the required H₂. The proposed CCUS MeOH plant is not an attractive investment with a discount rate of 9%. The following sensitivity analyses verify the effect of changes in electricity price, electrolyser efficiency and overall CapEx on the financial aspect of the proposed plant. Comparing the bars in Figure 10 it can be concluded, that the electricity price has the greatest impact on the NPV. In order to have a NPV = 0 the market price of MeOH should increase approximately more than 5 times. The high electricity consumption indicates that efforts should be focused on improving the efficiency of the electrolyser unit. Solid oxide electrolysers, could be a more efficient way of obtaining the necessary H₂ as they consume less energy; however they are still at development stage.



Reference to TEA Guidelines

In accordance with **Guideline B.19**, scenario analysis was conducted based on hot spot identification. Electricity price poses as the major cost driver and consequently alternative scenarios investigating different electricity supplies were developed.

B.9. Concluding remarks

This report assessed the economic viability of a Power to MeOH concept through CO₂ hydrogenation. A generic modelling platform was successfully established in Aspen Plus using coherent data to establish mass and energy balances for a subsequent economic assessment. The deterministic MMSP is 1402 €/t and there is a confidence interval of 95% that the MMSP is in the range of 1,238 €/t to 1,448 €/t. These figures are 4.9-5.8 fold greater than the fossil MeOH price. Sensitivity analysis revealed that the process is OpEx intensive with electricity price to pose as the major cost component. In fact, the factors that mostly affect the MMSP are associated with the electrolysis unit, *i.e.* electricity price and electrolyser efficiency rather than with variables related to the CO₂ conversion plant. A free of charge electricity supply will result in a competitive MeOH price; however, this will probably be unrealistic due to competition with other technological options of Power to X and beyond (*e.g.* exporting energy). Generation of positive NPV is not possible under the current technological and market status. Nevertheless, several technology and market developments can significantly improve the economics of the modelled processes. If energy independence and security are prioritised and societies begin to place a meaningful monetary burden on carbon-intensive technologies, these developments could accelerate the adoption of sustainable products such as the investigated MeOH.



Reference to TEA Guidelines

Following **Guideline B.21**, the report states and analyses the entire spectrum of elements of the checklist. Overall, the present study highlights the importance of a standardised approach for conducting TEA for CCU technologies in a transparent and reproducible manner, enabling direct comparisons of the broad range of technologies that can result the same product. The study is a worked example that demonstrates how these guidelines can be followed.

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Appendix A – Simulation results

Table A.1. Data extracted from simulations

Parameter	Value	Unit
Electrolysis		
Deionised water	80.85	t/h
Oxygen	68.2	t/h
H ₂	8.594	t/h
Electricity for electrolysis	450.12	MW
Cooling duties	2.81	MW
Cooling water (circulated)	248	t/h
Cooling water (make up)	12	t/h
Waste water	4	t/h
MeOH plant		
CO ₂ input	62.09	t/h
CO ₂ output	3.32	t/h
MeOH output	42.78	t/h
Cooling duties	43.13	MW
Heating duties*	31.3	MW
Electricity consumption	9.73	MW
Electricity generation	5	MW
Net electricity demand	4.73	MW
Cooling water (circulated)	3806	t/h
Cooling water (make up)	190.3	t/h
Boiler water (circulated)	6.4	t/h
Boiler water (make up)	1.6	t/h
Air	20	t/h
Catalyst loading	35	t/y
Waste water	24	t/h
CO₂ Capture		
Flue gas	224.27	t/h
CO ₂ input	71.39	t/h
CO ₂ capture	86.979	%
Electricity consumption	16.86	MW
Electricity generation	7.89	MW
Net electricity demand	8.97	MW
Cooling duty	16.1	MW
Cooling water (circulated)	1421	t/h
Cooling water (make up)	71.05	t/h
Waste water	9	t/h
*Fully integrated		

Appendix B – Data for economic evaluation

Table B.1. Equipment cost data

Process unit section	Reference cost (M€)	Design variable	Unit	Reference size	Scaling factor	Reference year	Reference
MeOH reactor	7.69	Feed gas	t/h	87.5	0.6	2006	52
Heat exchangers	39.26	Heat duty	MWth	355	1	2007	53
Distillation unit	16.58	MeOH flow rate	t/h	87.5	0.7	2006	52
Compressors	12.08	Power	MWe	10	0.67	2006	54
Cooling system	49.6	Q rejected	MWth	470	0.67	2007	55
Pump	14.77	Power	MWe	47.61	0.67	2011	56
Flash tank	0.983	Gas feed	t/s	1	0.8	2000	57
Gas turbine	20.98	Power	MWe	16.1	0.8	2001	58
Steam turbine	4.44	Power	MWe	4.1	0.8	2001	58
Wastewater	7.15	Flowrate	Millions gallons	1	0.6174	2007	59
Electrolyser	1.5	Power installed	MW	1	0.65	2015	36,60
Membrane Housing	0.189	Area	m ²	2000	0.7	2015	61
Membrane cost	37.75 €/m ²						61

Table B.2. Prices for variable costs

Parameter	Price	Unit	Reference
Cooling water	0.025	€/t	62
Process water	1	€/t	62
Catalyst	95.2	€/kg	9
Electricity	87.7	€/MWh	63
Oxygen	54.2	€/t	9
Waste water disposal	0.4	€/t	64

Table B.3. Methodology for fixed costs [62, 65]

Parameter	Price
Supervision	0.25×Labour
Direct overhead	0.5×(Labour + supervision)
General overhead	0.65×(Labour + supervision + direct)
Maintenance materials	0.03×ISBL
Interest	Debt interest × WC
Insurance and tax	0.015×FCI

Appendix C – Methodology for CapEx

Table C.1. Total Capital Investment (TCI) estimation methodology [65,66,67]

Cost component	Lang factor
<i>Purchased Equipment Cost (PEC)</i>	1
<i>Purchased Equipment Installation</i>	0.39
<i>Instrumentation and controls</i>	0.26
<i>Piping</i>	0.31
<i>Electrical Systems</i>	0.1
<i>Buildings (including services)</i>	0.29
<i>Yard Improvements</i>	0.12
ISBL	2.47×PEC
OSBL	0.12×ISBL
<i>Engineering and Supervision</i>	0.32-0.5 × (ISBL+OSBL)
<i>Construction Expenses</i>	0.34 (ISBL+OSBL)
<i>Legal Expenses</i>	0.04× (ISBL+OSBL)
<i>Contractor's Fee</i>	0.19 × (ISBL+OSBL)
Indirect costs (IC)	0.89-1.07×(ISBL+OSBL)
<i>Project Contingency</i>	0.15-0.4× (ISBL+OSBL+IC)
<i>Process Contingency</i>	0.05-0.5× (ISBL+OSBL+IC)
Fixed Capital Investment (FCI)	Contingencies+ISBL+OSBL+IC
<i>Working Capital (WC)</i>	0.15×FCI
Total Capital Investment (TCI)	FCI+WC

When an interval is mentioned, the first value refers to the high TRL technologies whereas the second one to low TRL systems

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Reviewed by: Leonard Müller, Marvin Bachmann, André Sternberg, André Bardow

PART C LCA Worked Example

THIS LCA WORKED EXAMPLE USES A FICTIONAL MODEL PRODUCED SOLELY TO ASSIST IN THE USE AND INTERPRETATION OF THE ACCOMPANYING LCA GUIDELINES. THE FOLLOWING LCA WORKED EXAMPLE HAS THEREFORE NOT UNDERGONE AN EXTERNAL REVIEW IN ACCORDANCE WITH ISO 14040/14044 AND CONSEQUENTLY IT SHOULD NOT BE USED IN COMPARISONS OF OTHER LIFE CYCLE ASSESSMENTS OF CCU AND/OR METHANOL TECHNOLOGIES.

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C.1 Executive Summary



Reference to LCA Guidelines

The LCA guidelines do not have a specific section covering what should be included within the executive summary, because this is not particular to CCU. However, a checklist of items to be included in the executive summary is included in **Section C.8 Reporting** of the guidelines, which is as follows:

Checklist - Executive summary

Goal of the study

- State the intended application of the study
- State the reasons for carrying out the study
- State the intended audience of the study
- State whether the results are to be used in comparative assertions disclosed to public
- State unambiguously the research question(s)
- State the classification of the assessed CCU technology

Scope of the study

- State functional unit clearly and unambiguously according to guideline and report changes due to solving of multi-functionality
- State system boundaries according to guideline
- State relevant issues with data quality and assumptions
- State technology readiness level (TRL) of processes and sub-processes
- Report production or storage capacity
- Report geographical scope
- State software system (and version) and data library (and version) used
- State type of review and provide additional information about reviewers

Life cycle inventory and Life cycle impact assessment

- State main results of life-cycle inventory and life-cycle impact assessment
- If results are reported on a relative basis, report basis
- Describe uncertainty and sensitivity analysis and report results separately

Interpretation

- State any conclusions, recommendation and limitations

Whilst the executive summary of an LCA report on CCU technologies is not significantly different to an executive summary from any other LCA report, this section has been provided as an example.

Methanol has a large global market as a commodity chemical, acting as a feedstock for many chemical processes. It can be synthesized from CO₂ instead of from natural gas, which means it becomes possible to indirectly substitute the fossil carbon currently used in a multitude of plastics, polymers and carbon containing chemicals with carbon from CO₂. This would release the chemicals industry from its current dependency upon fossil oil feedstock.

The goal of this study is to compare the environmental impacts (with a focus upon greenhouse gas emissions) of producing methanol as a chemical feedstock in Germany synthesised via a CCU route using CO₂ captured from a cement plant, compared to the conventional route using natural gas. CO₂-based methanol which is identical in chemical structure and composition to its conventional counterpart is being produced for use as a chemical feedstock. The results are intended to allow policy-makers to compare the impacts of producing methanol using CCU versus conventional technology and to review the effect of different locations. The results will not be used in comparative assertions disclosed to the public.

System expansion was used to include the additional function of the CO₂ source to the functional unit - in this case a cement plant. To enable comparison, the boundary of the conventional methanol plant was expanded to include a cement plant producing the same quantity of cement. The basis for comparison between the two technologies (the functional unit) is the production of 1 t methanol and 1.96 tonnes Portland cement. The modelled plant has a production capacity of 1 kt methanol/day. The study is limited by the lack of infrastructure modelling for the CO₂-capture and electrolysis units, the membrane CO₂-capture technology used is at a low technology readiness level (TRL 4) whereas the PEM electrolyzer is at TRL 9 and the methanol plant is considered to be at TRL 7. The software used for modelling the system was Gabi version 8.6.0.20 and the database was ecoinvent v3.4.

A contribution analysis revealed that the environmental impacts of the CCU process are highly dependent upon the electricity used, much of the electricity is used for the electrolysis of water for H₂ production. A scenario analysis was conducted using different electricity mixes (status-quo, low decarbonized, high decarbonized, full decarbonized) to determine the effect of decarbonizing the electricity supply.

The life cycle impact assessment revealed that the CCU methanol process only reduces greenhouse gas (GHG) emissions if low-carbon electricity is used. The high electricity demand of electrolysis means that if current grid mix electricity is used, the GHG emissions of the CCU process are several times higher than the conventional process. However, GHG emissions of the CCU process were reduced by up to 90% by switching to using decarbonized electricity. This contrasted with the conventional process where switching to wind power led to a reduction of just 2% in GHG emissions. When the low decarbonized scenario is used, GHG emissions of the CCU process are slightly lower than the conventional process. Using the high decarbonized scenario leads to GHG emissions around 50% lower than the conventional process using the same electricity. If just wind power is used (full decarbonized scenario), then GHG emissions reductions of the order of 60% over the conventional process can be achieved. Some caution needs to be applied due to uncertainty introduced by study limitations.

The technique required to enable comparison of siting the process in different countries was discussed. An issue to be aware of is that the GHG emissions per main product of the CO₂ source (in terms of CO₂ output/t cement in this case) differs in different countries. When using system expansion of a multi-functional system, this leads to different amounts of cement/t methanol in the functional unit. When comparing siting a CCU plant in Germany versus Switzerland or USA, Switzerland comes out as being the most favourable location as the highly efficient Swiss cement plants produce more cement per quantity of CO₂ required for 1 t of methanol.

Note: the precision of results provided in the executive summary has been reduced due to caveats introduced by the study limitations discussed in **Sections C.4.2.2** and **C.7.2** of this worked example.

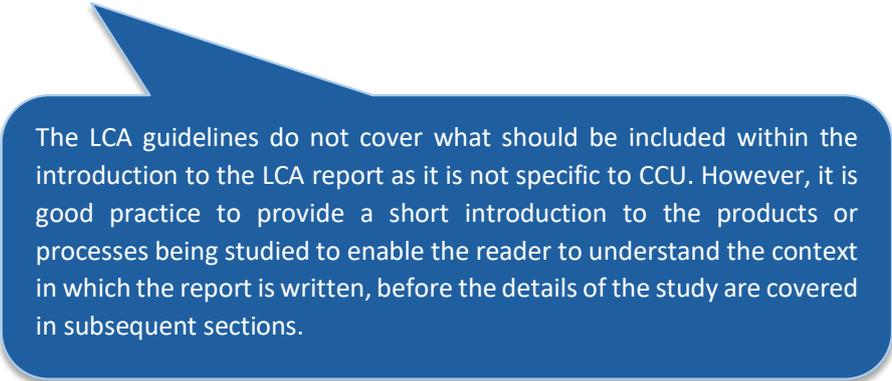
C.2 Technical Summary

GOAL	CCU Product	Methanol as a chemical feedstock									
	Goal	What would be the environmental consequences (with a focus on global warming impacts) of using methanol as a chemical feedstock synthesized via hydrogenation of carbon dioxide compared to methanol synthesis from natural gas?									
	Brief description	CO ₂ is captured via membrane capture from a cement plant, H ₂ is produced via PEM water electrolysis, subsequently methanol is produced via thermochemical synthesis.									
	Intended audience	LCA practitioner									
	Functional Unit	The production of 1 tonne of MeOH as a feedstock/chemical/solvent and 1.96 t cement.									
SCOPE	Limitations and Assumptions	<ul style="list-style-type: none"> Based on literature data and Ecoinvent 3.4 Cut off data – no use phase data included No infrastructure for H₂ production or carbon capture included Low TRL process for carbon capture 									
	Boundary (i.e. cradle to gate)	Cradle to gate									
	Location	Germany									
	Time frames	Current, 2030, 2050									
INVENTORY	Multi-functional approach	<input type="checkbox"/> Sub-division <input checked="" type="checkbox"/> System expansion <input type="checkbox"/> System expansion via substitution <input type="checkbox"/> Virtual sub-division <input type="checkbox"/> Mass allocation	<input type="checkbox"/> Energy allocation <input type="checkbox"/> Economic allocation <input type="checkbox"/> Closed loop scenarios <input type="checkbox"/> Other (specify).....								
	Data Source	<input type="checkbox"/> Primary sources <input checked="" type="checkbox"/> Secondary sources <input type="checkbox"/> Stoichiometric data	<input checked="" type="checkbox"/> Process modelling based data <input type="checkbox"/> Mixes sources <input type="checkbox"/> Other (specify).....								
	Energy sources (select all that apply)	<input checked="" type="checkbox"/> Grid mix <input type="checkbox"/> Power station with Carbon Capture <input checked="" type="checkbox"/> Wind <input type="checkbox"/> Solar	<input type="checkbox"/> Nuclear <input type="checkbox"/> Hydro <input type="checkbox"/> Future (see timeframes) <input type="checkbox"/> Other (specify).....								
	Main Sub-Processes and TRLS	<table border="1"> <thead> <tr> <th>SUB-PROCESS</th> <th>TRL</th> </tr> </thead> <tbody> <tr> <td>Carbon capture via membrane</td> <td>TRL 4</td> </tr> <tr> <td>PEM water electrolysis</td> <td>TRL 9</td> </tr> <tr> <td>Methanol Synthesis</td> <td>TRL 7</td> </tr> </tbody> </table>	SUB-PROCESS	TRL	Carbon capture via membrane	TRL 4	PEM water electrolysis	TRL 9	Methanol Synthesis	TRL 7	
	SUB-PROCESS	TRL									
Carbon capture via membrane	TRL 4										
PEM water electrolysis	TRL 9										
Methanol Synthesis	TRL 7										
Database & Software	ecoinvent v3.4 and Gabi v8.6.0.20										
ASSESSMENT	LCIA Method	<input checked="" type="checkbox"/> ILCD 1.09a <input type="checkbox"/> ReCiPe OTHER IMPACT METHODS <input type="checkbox"/>	SINGLE CATEGORIES: <input type="checkbox"/> Global warming <input type="checkbox"/> CED <input type="checkbox"/> use TOX								
	Highlighted Results (graphical, text or tabular format)	<ul style="list-style-type: none"> GHG emissions of the modelled CCU methanol process sited in Germany are higher than the conventional SMR process unless low-carbon electricity is used Switching to 100% wind power reduces the GHG emissions of the modelled CCU process by 90%, but the same switch only reduces the emissions of the conventional SMR methanol process by 2% GHG emissions reductions of the order of 60% can be achieved by changing from the conventional SMR methanol production to the modelled CCU production process 									
INTERPRETATION	Main Conclusions	<ul style="list-style-type: none"> GHG emissions of the CCU process are highly dependent upon the source of electricity used for the electrolysis which dominates the electricity use The CCU methanol process only reduces GHG emissions if low carbon energy is used If wind power is used, modelled GHG emissions can be reduced by around 40% over the conventional process 									
	Sensitivity Analysis	<input type="checkbox"/> No <input checked="" type="checkbox"/> Yes (please specify below) Sensitivity of location - Germany compared to Switzerland and USA. Switzerland would offer the lowest global warming impacts per functional unit as more cement is produced per quantity of CO ₂ required to produce 1 t methanol.									

C.3 Introduction

Methanol can be used as a liquid fuel, either directly or after dehydration to dimethyl ether, but currently its main use is as a commodity chemical where it acts as a feedstock for many chemical processes. The market for methanol is large, with global demand reaching 70 Mt in 2015, supplied by over 90 methanol plants worldwide [1]. Methanol can be synthesized from CO₂ and is itself used as a chemical building block in the synthesis of many different compounds. By producing methanol using carbon capture and utilization (CCU) technologies from waste CO₂, rather than from steam methane reformation as per the conventional route, it becomes possible to indirectly substitute fossil carbon with carbon derived from CO₂ into a multitude of plastics, polymers and carbon containing chemicals. In this way, the dependency of the chemicals industry upon fossil oil feedstock is significantly reduced.

This report assesses the environmental impacts (with a focus upon global warming impacts) of a CCU methanol plant capable of producing around 1 Kt methanol per day.



The LCA guidelines do not cover what should be included within the introduction to the LCA report as it is not specific to CCU. However, it is good practice to provide a short introduction to the products or processes being studied to enable the reader to understand the context in which the report is written, before the details of the study are covered in subsequent sections.

C.4 Goal and scope of the Study

C.4.1 Goal of the study



Reference to LCA Guidelines

A description of what should be included within the Goals section of an LCA report is provided in **Section C.3 Goal definition** of the guidelines. Specific guidelines of what shall or should be included are listed in **Section C.3.2**.

In addition, a checklist of items to be included in each section of the report is included in **Section C.8 Reporting** of the guidelines, which for the goal is as follows:

Goal of the study

- State the intended application of the study
- State the reasons for carrying out the study
- State the intended audience of the study
- State whether the results are to be used in comparative assertions disclosed to public
- State unambiguous research question(s)
- State the classification of the assessed CCU technology
- State limitations due to the assumptions and methods, e.g. if study is preliminary
- State the commissioner of the study and other influential actors
- State technology readiness level (TRL) of processes and sub-processes
- Report production or storage capacity
- State review process and review experts, if any

The following study has the goal:

To compare the environmental impacts (with a focus upon global warming impacts) of producing methanol for use as a chemical feedstock in Germany, synthesised via two routes: the hydrogenation of CO₂ captured from a cement plant vs methanol synthesised using the conventional steam methane reforming process from natural gas.

The results are intended to be used to compare the environmental impact of producing methanol using CCU technology versus conventional technology. CO₂-based methanol which is identical in chemical structure and composition to its conventional counterpart and is being produced for use as a chemical feedstock rather than as a fuel. The modelled plant has a production capacity of 1 Kt methanol/day. The study is limited by the lack of infrastructure modelling for the CO₂-capture and electrolysis units, also the compression energy required to pressurise the CO₂ prior to transportation was not modelled. Further limitations are imposed by the membrane CO₂-capture technology utilised being at a low technology readiness level (TRL 4) whereas the PEM electrolyser is at TRL 9 and the methanol plant is considered to be at TRL 7.

The goal of the study is to provide information to assist policy-makers in comparing the environmental impacts of CCU technologies with conventional technologies. However, it is not intended that the results of this study will be used for marketing purposes or to make comparisons between products to the public.

This study was commissioned by The Global CO₂ Initiative/CO₂ Sciences and was conducted by the University of Sheffield with assistance from RWTH Aachen University. This report has undergone review by a panel of independent experts with a knowledge of CCU consisting of (*names and affiliations*).

As mentioned in the overall introduction to this set of worked examples and in the preface to this report, this document is a fictional worked example produced to help in the use and interpretation of the guidelines. Therefore, the true goal of this study is rather different than that described above and the report has not undergone external review as required by a published LCA report according to ISO 14040/14044.

C.4.2 Scope of the study



Reference to LCA Guidelines

A description of what should be included within the Scope section of an LCA report is provided in **Section C.4 Scope definition** of the guidelines. The specific guidelines of what shall, should or may be included are provided as **Guidelines C.2 to C.6**.

In addition, a checklist of items to be included in each section of the report is included in **Section C.8 Reporting** of the guidelines, which for the scope is as follows:

Scope of the study:

- State functional unit clearly and unambiguously according to the guideline, report changes due to solving of multi-functionality
- State performance characteristics, any omission of additional function in comparison and how performance is measured (might apply for products different in chemical structure and composition to their conventional counterparts)
- State system boundaries according to guideline and cut-off criteria including a system boundaries diagram
- State omitted life cycle stages and processes (might apply for products different in chemical structure and composition to their conventional counterparts)
- State relevant issues with data quality and assumptions
- State method(s) to solve multi-functionality
- State impact assessment methods
- State data quality needs and how energy and material inputs and outputs are quantified
- State software system (and version) and data library (and version) used
- State type of review and provide additional information about reviewers

C.4.2.1 Product characteristics

The product under assessment is CO₂-based methanol which is identical in chemical structure and composition to its conventional counterpart. In this study it is being assessed as a chemical rather than a fuel. The functional unit for the methanol component is 1 tonne methanol produced. **Section C.4.2.4 Solving multi-functionality and defining the functional unit** below further defines the functional unit in

terms of the cement produced as this is determined by how the multi-functionality of the system is resolved.

The classification is important as it determines the nature of the functional unit. For instance, chemicals produced by CCU generally have an identical chemical structure and composition to those produced by conventional techniques. They can therefore be compared by mass with the functional unit in Kg. However, CCU fuels such as synthetic gasoline or diesel may have different energy contents than the conventional equivalents, in which case the functional unit needs to reflect the energy service such as heat that it provides. Fuels with identical chemical structure and composition would be compared using energy content in the functional unit. More information is available in the LCA Guidelines.



Reference to LCA Guidelines

As described in **Section C.4.1 Product system, its function, functional unit and reference flow** of the guidelines, the functional unit description provided follows the recommendation for CO₂-based chemicals. In this situation, **mass** is used for comparison since the end-product has the same molecular structure and composition in both of the process routes assessed.

The process routes considered for this assessment are:

- a. MeOH produced from steam methane reforming (SRM) of natural gas - the conventional route.
- b. MeOH derived from CO₂ captured from a cement plant and H₂ produced from the electrolysis of water - the CCU route.

C.4.2.2 System diagrams, completeness requirements and related cut-offs

All processes in this study are assessed from cradle-to-gate as shown diagrammatically in Figures 1 and 2. System diagrams are presented for both the CCU route and the conventional (reference) route.

This example considers only the production of methanol and not its use. This is because CO₂-based methanol is identical in its structure and composition to fossil-based methanol, so the usage and end-of-life treatment will be identical. It is therefore not necessary to consider a cradle-to-grave scenario, a cradle-to-gate analysis is sufficient.



Reference to LCA Guidelines

The use of a cradle-to-gate boundary is in line with the recommendation in **Section C.4.2** of the guidelines when the chemical structure and composition of the CO₂-derived product is identical to the conventionally-derived product.

All processes in this case study are assessed from cradle-to-gate as shown in Figures 1 and 2.

a) Methanol from natural gas as a chemical feedstock:

The reference/benchmark process uses steam methane reforming (SMR) to produce syngas which is then converted to methanol (MeOH). The process involves inputs of energy (heat, electricity), raw materials (natural gas, water, catalyst), plant infrastructure etc., and results in emissions to air/water from the process as indicated in Figure 1 below (source: ecoinvent version 3.4).

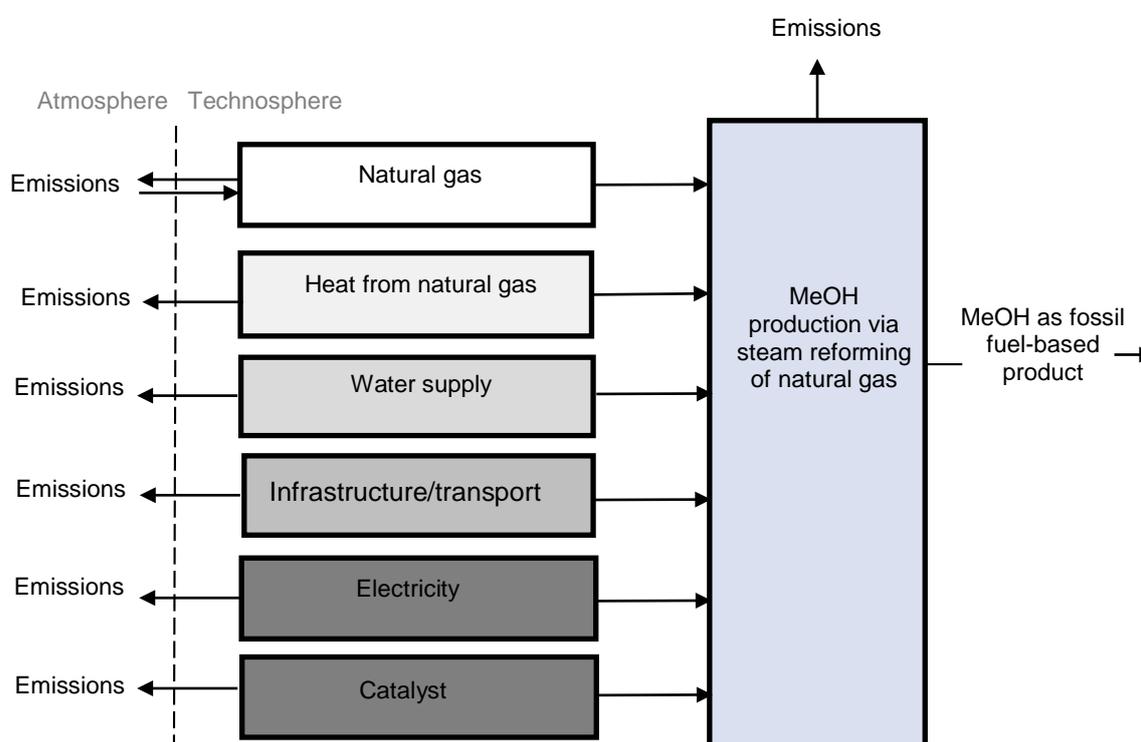


Figure 1. System diagram for methanol synthesis using natural gas as feedstock

b) Methanol derived from captured CO₂ and H₂ as a feedstock chemical:

A hydrogenation process is used for the synthesis of methanol using captured CO₂ and H₂ (Figure 2). A multistage membrane system is used to separate the CO₂ from the flue gas. Hydrogen is produced via electrolysis using a proton exchange membrane (PEM) electrolyser. Inputs include energy (electricity), raw

materials (water, catalyst), methanol plant infrastructure, and emissions to air/water result from the process. Source: the LCI for this process.

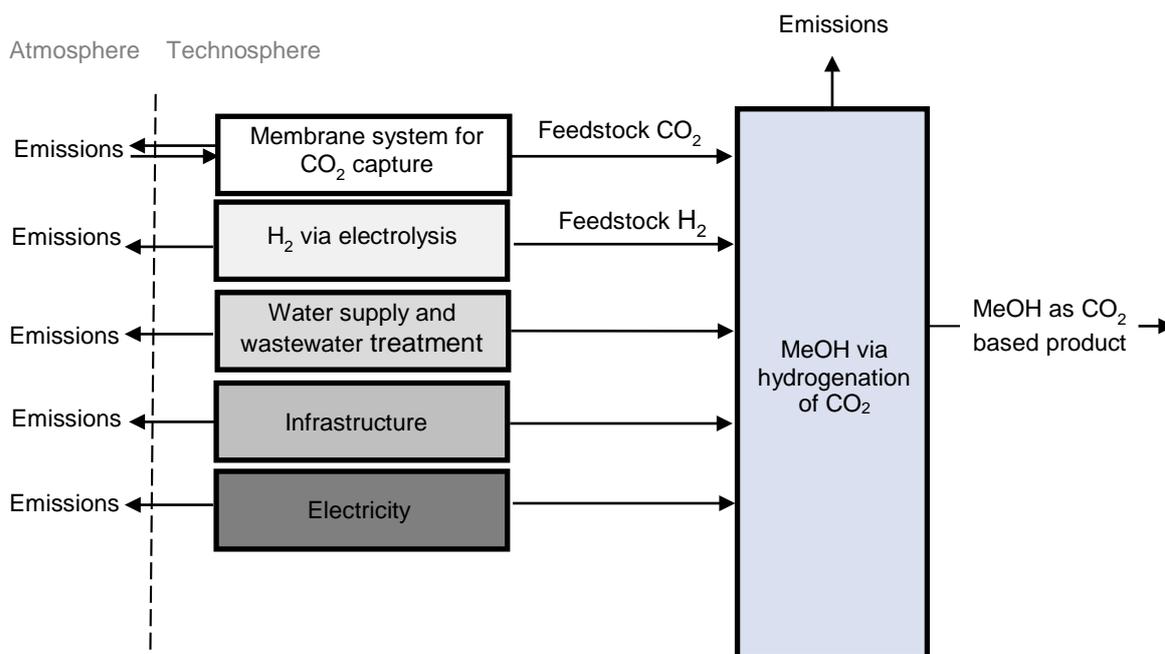


Figure 2. System diagram for methanol synthesis using CO₂ and H₂ as feedstock.



Reference to LCA Guidelines

The description of the reference process is included as recommended in **Section C.4.2** of the guidelines. This ensures that the avoided impacts of the conventional (reference) process are transparent and measurable. The CO₂ source, CO₂ purification (not required in this process) and CO₂ transport are included for the CCU process and are shown within the system boundaries of the process in **Figure 2** as suggested in **Section C.4.2** of the guidelines.

The limitations of the study, including data cut-offs, data availability and other uncertainties are summarized in Table 1.

Limitation	Description
Data cut-offs	Infrastructure for CO ₂ capture and H ₂ production was not included in the modelling. The energy requirement for CO ₂ compression prior to transportation by road was not modelled.
Data availability	Some of the datasets used from ecoinvent version 3.4 were several years out of date, as described further in Table 2 below.
Other uncertainties	The CO ₂ capture unit modelled is at a low technology readiness level.

Table 1. Limitations of the LCA

Limitations of the study include the data cut-offs used: infrastructure was not modeled for either the CO₂ capture step or H₂ production and the energy required for CO₂ compression prior to transport was also not modelled. The study was limited by data timeliness as some of the data sets used from the ecoinvent

database version 3.4 are several years out of date, as presented in Table 2 below. Other uncertainties were created by the use of a carbon capture membrane separation technology which is currently at a TRL assessed to be between 3 and 4.

C.4.2.3 Geographical and time related representativeness

Region of study:

The main location assessed in this study is Germany, but a sensitivity analysis is conducted to test the sensitivity of the environmental impacts of the process if it had been located in Switzerland or the United States. Energy and other inventories were adapted to the conditions of these locations as closely as possible, as indicated in Table 2. When data for the specific country was not available, data for the regional average (RER for Europe) was used and if this was not available then global average (GLO) data was used. Whilst this limits the regional representativeness of the data, the technology representativeness can still be achieved through the use of such secondary data as proxy data.

A comparison of locations is included as an example sensitivity assessment in **Section C.7 Interpretation** of this report, because it highlights one of the challenges of performing LCA on CCU technologies.

Transportation:

Once captured at a cement plant, the CO₂ requires transportation from the point of capture to the point of use at the CCU methanol production facility. Since the necessary pipeline infrastructure does not yet exist in most locations, it has been assumed that the CO₂ is transported by road freight. A distance of 50 Km was used for all countries so that the transportation impacts could be included within the model but an individual country would not benefit from any arbitrary variations in projected/assumed transport distances. The energy requirement for CO₂ compression prior to transportation by road was not modelled.

Since a cradle-to-factory-gate system boundary has been adopted, the use-phase transportation of the methanol was not modeled in this study.



Reference to LCA Guidelines

Section C.4.4 of the guidelines covers data quality. Complete data availability may be limited to full LCA studies, so clarity and transparency in explaining the sources of data used is important. **Guideline C.9** in **Section C.5.5** explains that the reference (conventional) process should use the current best available technology because the conventional technologies in future when CCU is implemented is likely to be unknown. In this way, the perception of bias is avoided.

Process inputs	Location (database time period)		
	Germany	Switzerland	USA
Natural gas	DE: natural gas production (1996-2000)	DE: natural gas production (1996-2000)	US: natural gas production (2010-2010)
Methanol production: <i>Energy inputs adapted to region of study</i>	GLO: methanol production (1994 to 2001)	GLO: methanol production (1994 to 2001)	GLO: methanol production (1994 to 2001)
Tap water	Europe without Switzerland: tap water production, underground water without treatment (2012 to 2012)	CH: tap water production, underground water without treatment (2012 to 2012)	GLO: tap water production, underground water without treatment (2012 to 2012)
Deionised water	Europe without Switzerland: water production, deionised, from tap water, at user (1992 to 2002)	CH: water production, deionised, from tap water, at user (1992 to 2002)	GLO: water production, deionised, from tap water, at user (1992 to 2002)
Treatment of wastewater	Europe without Switzerland: treatment of wastewater, average, capacity 1E9 l/year (1994-2000)	CH: treatment of wastewater, average, capacity 1E9 l/year (1994-2000)	GLO: treatment of wastewater, average, capacity 1E9 l/year (1994-2000)
Transport, lorry	RER: transport, freight, lorry 3.5-7.5 metric ton (2009-2013)	RER: transport, freight, lorry 3.5-7.5 metric ton (2009-2013)	GLO: transport, freight, lorry 3.5-7.5 metric ton (2009-2013)
Electricity grid mix	DE: electricity high voltage, production mix	CH: electricity high voltage, production mix	US: WECC, US only, electricity high voltage, production mix
Electricity, wind power	DE: electricity production, wind >3 MW, turbine, onshore (2012-2012)	CH: electricity production, wind >3 MW, turbine, onshore (2016-2018)	US: WECC US only, electricity production, wind >3 MW, turbine, onshore (2012-2012)
Heat	Europe without Switzerland: market for heat, district or industrial, natural gas (2011-2011)	CH: market for heat, district or industrial, natural gas (2011-2011)	GLO: market for heat, district or industrial, natural gas (2011-2011)
Aluminium oxide	GLO: aluminium oxide production (2012-2012)	GLO: aluminium oxide production (2012-2012)	GLO: aluminium oxide production (2012-2012)
Copper oxide	RER: copper oxide production (2000-2020)	RER: copper oxide production (2000-2020)	GLO: copper oxide production (2000-2020)
Zinc oxide	RER: zinc oxide production (2005-2020)	RER: zinc oxide production (2005-2020)	GLO: zinc oxide production (2005-2020)
Molybdenum	RER: molybdenum production (2000-2003)	RER: molybdenum production (2000-2003)	GLO: molybdenum production (2000-2003)
Nickel 99.5%	GLO: market for nickel 99.5% (2011-2011)	GLO: market for nickel 99.5% (2011-2011)	GLO: market for nickel 99.5% (2011-2011)
Zinc	GLO: market for zinc (2011-2011)	GLO: market for zinc (2011-2011)	GLO: market for zinc (2011-2011)
Methanol factory	GLO: market for methanol factory (2011 to 2011)	GLO: market for methanol factory (2011 to 2011)	GLO: market for methanol factory (2011 to 2011)

Table 2. Databases used for process input with time representativeness.

Database time representativeness:

The time representativeness for each process used in this case study for both production routes is shown in Table 2. All process inputs apart from the IEA electricity scenarios were obtained from ecoinvent version 3.4. The impact of this will be discussed during the interpretation of the results.

To prevent erroneous comparisons, the conventional routes have been taken in an un-aggregated form and adapted to the same conditions as the CCU process. This mainly includes energy supply, feedstocks, water supply and treatment and the same chemical facility for infrastructure assumptions.

C.4.2.4 Solving multi-functionality and defining the functional unit

In the system under investigation, CO₂ is captured from a cement plant which then becomes a feedstock for methanol production as a CO₂-based chemical.

To allow the environmental impacts of producing the CO₂ to be included in the assessment, the cement plant is included in the analysis. This results in the overall system being multi-functional in that multiple products (cement and methanol) are produced. To enable the comparison of the conventional route and the CCU route under equal conditions, this multi-functionality needs to be solved. Depending upon where the system boundaries are drawn, solving the multi-functionality can lead to having a single product or multiple products within the functional unit.

**Reference to LCA Guidelines**

Section C.4.3.2 Solving multi-functionality describes a hierarchy of methods for solving multi-functionality. System expansion is the method that should be used to solve multi-functionality in this instance. For this reason, cement production is included within the system boundary of the conventional methanol production process.

Sub-division should be the first approach to solving multi-functionality, but in this instance the impacts of producing the cement cannot be separated from the impacts of producing the CO₂ because cement cannot be produced without producing CO₂. Since CO₂ is an essential feedstock of the methanol, it is therefore impossible to separate the impacts of cement production from those of methanol production in this way.

The second approach to solving the multi-functionality of the system is to use system expansion to draw the system boundary to include both the cement plant and the methanol plant. Figure 3 below aims to present this diagrammatically. With the system expansion approach, cement is considered to be produced as a joint product along with the methanol, meaning that the functional unit contains both products. As long as a cement plant is added within the system boundary of the conventional methanol plant, then this approach allows the comparison of the conventional route and the CCU route under equal conditions.

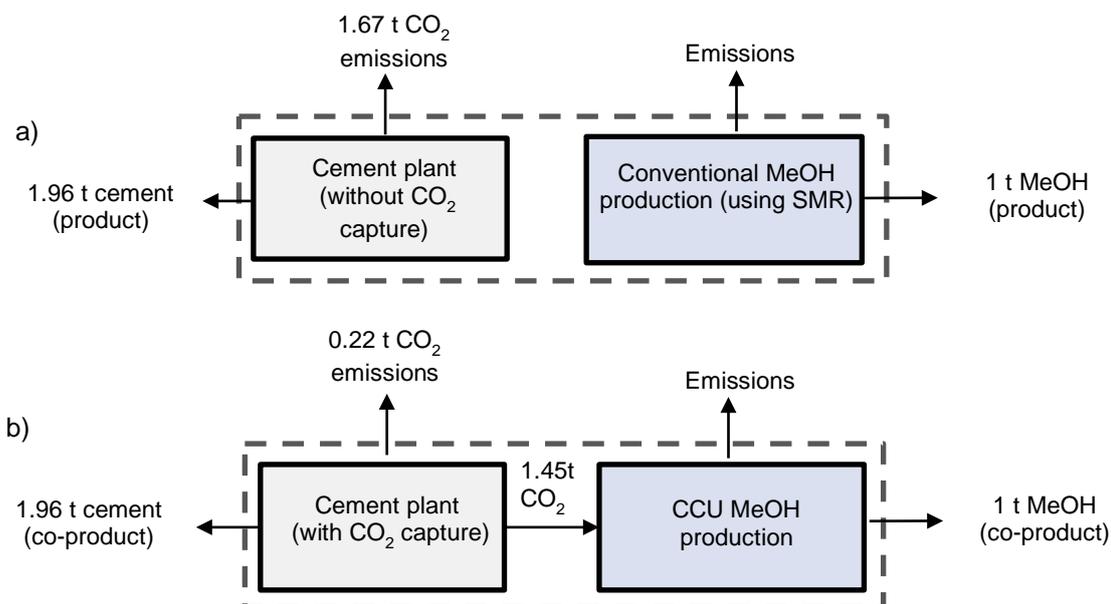


Figure 3. System expansion to enable comparison between a) the conventional production system of methanol from steam methane reforming and b) CCU based methanol. The dotted lines indicate the system boundary.

In the system described, according to ecoinvent 3.4, a Portland cement plant in Germany produces 0.852 Kg CO₂/Kg cement, meaning that such a plant would produce 1.959 t of cement to produce the 1.669 t CO₂ required to produce 1 t methanol. At the modelled capture efficiency of 86.98%, a feed of 1.669 t CO₂ is required to capture the 1.451 t CO₂ needed to produce 1 t methanol. Therefore, in this case the functional unit required to compare CCU production with conventional production is 1 t methanol and 1.96 t cement.

For the product specific assessment of methanol as a chemical feedstock, the basis for comparison used is mass, therefore the functional unit used is 1 tonne of methanol and 1.96 tonnes of cement.

C.5 Inventory data



Reference to LCA Guidelines

A description of what should be included within the inventory section of an LCA report is provided in **Section C.5 Life cycle inventory (LCI)** of the guidelines, with **Guidelines C.7 to C.9** listing the specific items which need to be included.

In addition, a checklist of items to be included in each section of the report is included in **Section C.8 Reporting** of the guidelines, which for the life cycle inventory is as follows:

Life cycle inventory

- Include flow diagram of assessed process system(s)
- State types and sources of required data and information
- State calculation procedures
- State all assumption made
- Describe sensitivity analysis for refining system boundaries
- Include calculated full LCI results (if this does not contradict with confidentiality)
- State data representativeness and appropriateness of LCI data
- If results are reported on a relative basis, report basis
- State results obtained from scenario analysis (including scenarios) and threshold values, if any

The main process inputs (electricity, water use, wastewater treatment etc.) into the three main stages of the CCU process (CO₂ capture, H₂ production via electrolysis, and methanol production from hydrogenation of CO₂) is provided in Table 3 below.

Table 3 is provided to enable the reader of this worked example to see, at-a-glance, the most relevant inputs to the CCU methanol process. The rest of the inventory data, together with the explanations of the processes and process flow diagrams are provided in the Appendix of this document. This was done in an attempt to avoid distracting the reader from working through the LCA examples by providing too much information at this stage. However, it is usual to provide all inventory data within the Inventory chapter of the report rather than in an appendix.

Parameter	Scaled to factory output		Scaled to 1 t methanol	
	Value	Unit	Value	Unit
CO₂ Capture				
CO ₂ input	71.39	t/h	1.669	t/h
CO ₂ capture efficiency	86.979	%	86.979	%
CO ₂ output	62.09	t/h	1.451	t/h
Net electricity demand	8.97	MW	0.209	MW
Cooling water (make up)	71.05	t/h	1.661	t/h
Waste water	9	t/h	0.210	t/h
Electrolysis				
Deionised water	80.85	t/h	1.89	t/h
H ₂ output	8.594	t/h	0.2	t/h
Electricity for electrolysis	450.12	MW	10.522	MW
Waste water	4	t/h	0.1	t/h
Methanol synthesis				
CO ₂ input	62.09	t/h	1.451	t/h
CO ₂ output	3.32	t/h	0.077	t/h
MeOH output	42.78	t/h	1	t/h
Net electricity demand	4.73	MW	0.111	MW
Cooling water (make up)	190.3	t/h	4.448	t/h
Boiler water (make up)	1.6	t/h	0.037	t/h
Catalyst	4.375	Kg/h	0.102	Kg/h
Waste water	24	t/h	0.56	t/h

Table 3. The main process inputs and outputs from the three main stages of production of methanol by the hydrogenation of CO₂.

C.6 Life Cycle Impact Assessment



Reference to LCA Guidelines

A description of what should be included within this section of an LCA report is provided in **Section C.6 Life cycle impact assessment** of the guidelines, with **Guidelines C.10** and **C.11** specifying the items which need to be included.

In addition, a checklist of items to be included in each section of the report is included in **Section C.8 Reporting** of the guidelines, which for the life cycle inventory is as follows:

Life cycle impact assessment

- Include results of life cycle impact assessment
- State if impact categories coverage is reduced, e.g. in case of carbon footprinting
- If results are reported on a relative basis, report basis
- State if delayed emissions occur and include emission time profile if needed
- If applied, state discounting method and discounted results

C.6.1 Approach taken

Gabi ts software (version 8.6.0.20) from thinkstep, utilizing the ecoinvent 3.4 database, was used in this study. A model of methanol production utilizing CO₂ as the carbon source was compared to methanol production using the conventional technique of steam methane reforming (SMR).

Note: when constructing the model in Gabi using ecoinvent 3.4, the input flow of CO₂ into the methanol plant was treated as an elementary flow (using Carbon dioxide [inorganic emission to air]) so that the environmental impact of capturing the CO₂ emissions were included in the assessment. Treating the CO₂ flow as a technical flow (using Carbon dioxide [inorganic intermediate product]) enables determination of the environmental impact of the process without distortion caused by the captured CO₂ but does not include the beneficial impact of avoiding the CO₂ emissions that were captured.

The results arising from the life cycle impact assessment were obtained using the ILCD/PEF recommendation method v1.09 due to this method being recommended by the European Commission's Joint Research Centre as appropriate for a study based in Germany. Although impacts upon global warming are the main category of interest in this report, for transparency, all of the ILCD impact categories at midpoint level are presented in tabular form.



Reference to LCA Guidelines

The method used to analyze the results of the LCIA are described in **Section C.6.1** of the guidelines and specific guidelines of what shall, should or may be included as Guideline **C.10**.

A contribution analysis was conducted to determine the contribution that the different processes which make up the CCU process (CO₂ capture, CO₂ transport, electrolysis, methanol production) make towards the overall environmental impacts. The contribution that electricity usage makes to the impacts of each of these individual processes is calculated.



Reference to LCA Guidelines

As described in **Section C.7.3** of the guidelines, environmental impacts in CCU can be very sensitive as they are often closely linked to energy requirements. Showing only a dependency on clean feedstocks limits interpretation and can leave to ambiguity. Presenting alternative energy scenarios facilitates comparison between similar assessments and prevents overstating the environmental benefits of CO₂-based products. The standard inventory data-sets for electricity scenarios, as introduced in **Section C.7.3** (and described in greater detail in **Appendix C.9.1**) of the guidelines, have therefore been used.

Electricity usage was found to be the major cause of the environmental impacts caused by the CCU process, so a scenario analysis of the effect of using different electricity scenarios was conducted as part of the impact assessment. For both the conventional methanol production process and the CCU production process, the status-quo is compared to a low decarbonized (IEA 2030 electricity mix scenario), a high decarbonized (IEA 2050 electricity mix scenario) and a fully decarbonized scenario where all electricity used is provided by onshore wind power. The low and high decarbonized scenarios were created based upon data published in the IEA Energy Technology Perspectives 2017 report [18]. They are not specific to any country, but simply reflect a lower carbon intensity and other environmental impacts as specified by the IEA per kWh of electricity produced. The wind power used for the decarbonized scenario was created using current wind power data for > 3MW onshore wind turbines in Germany and therefore reflects the efficiencies of wind power today, rather than projected efficiencies in 2030, 2050 or beyond.

It is recognized that LCA practitioners may wish to evaluate the environmental impact of changing the country in which processes are modelled in order to identify the best location for a new development. A CCU-specific challenge to having a consistent functional unit which arises due to differences in the carbon efficiency of the CO₂ source is highlighted and methods to avoid it are discussed in *Section C.7 Life Cycle Interpretation*.

Additional sensitivity analyses to assess the sensitivity of the system to changes to the electrolyser efficiency, the CO₂ capture efficiency and methanol reactor pressure were also conducted. As it is recognized that many LCA practitioners wish to integrate the results of an LCA together with a TEA, these sensitivity analyses have been included in the section of this document specifically dealing with the integration of LCA and TEA data, as an example of how the environmental and economic impacts of such analyses can be presented.

C.6.2 Results of methanol production impact assessment

C.6.2.1 Conventional methanol vs CCU methanol production in Germany

Life cycle impact assessment was conducted, involving the utilisation of the four different electricity scenarios discussed in *Section C.6.1* (status-quo, low decarbonized, high decarbonized and fully

decarbonized). The data is presented following system expansion to enable the direct comparison of conventional methanol production (Table 4) with CCU methanol production (Table 5) where the CO₂ used for the CO₂-derived methanol is provided by a cement plant.

The first observation which can be made from Table 4 is that the greenhouse gas emissions of the conventional steam methane reforming (SMR) process of methanol production is only modestly dependent upon the source of the electricity. Attempting to decarbonize the conventional process by switching from using electricity from the current German grid to electricity sourced from 100% onshore wind turbines leads to a reduction in the greenhouse gas emissions (as determined by the impact category: climate change midpoint, excluding biogenic carbon) of just 2%. Clearly the greenhouse gas emissions arising from this process arise from the SMR process itself rather than through electricity use.

In contrast, Table 5 reveals that the impacts of the CCU process are highly dependent upon the source of the electricity, reflecting that the CCU process has a much higher reliance upon electricity. A reduction of 90% in greenhouse gas emissions is obtained if the status-quo electricity (current grid mix) is switched to fully decarbonized electricity (onshore wind turbines). The low and high decarbonized scenarios reflect the impacts part-way through the transition between the current and the fully decarbonized scenarios.

Impact categories (ILCD/PEF recommendation v1.09)	Conventional Methanol Plant			
	Status-quo	Low decarbonised	High decarbonised	Fully decarbonised
Acidification midpoint (v1.09) [Mole of H+ eq.]	7.3	7.1	7.1	7.1
Climate change midpoint, excl biogenic carbon (v1.09) [kg CO2 eq.]	2260	2223	2217	2214
Climate change midpoint, incl biogenic carbon (v1.09) [kg CO2 eq.]	2290	2250	2242	2242
Ecotoxicity freshwater midpoint (v1.09) [CTUe]	3354	2901	2901	4279
Eutrophication freshwater midpoint (v1.09) [kg P eq.]	0.25	0.18	0.18	0.18
Eutrophication marine midpoint (v1.09) [kg N eq.]	1.26	1.23	1.23	1.23
Eutrophication terrestrial midpoint (v1.09) [Mole of N eq.]	15.0	14.3	14.3	14.3
Human toxicity midpoint, cancer effects (v1.09) [CTUh]	0.00003	0.00003	0.00003	0.00003
Human toxicity midpoint, non-cancer effects (v1.09) [CTUh]	0.0002	0.0002	0.0002	0.0002
Ionizing radiation midpoint, human health (v1.09) [kBq U235 eq.]	72.3	69.6	69.5	63.7
Land use midpoint (v1.09) [kg C deficit eq.]	1801	1803	1806	1795
Ozone depletion midpoint (v1.09) [kg CFC-11 eq.]	0.0001	0.0001	0.0001	0.0001
Particulate matter/Respiratory inorganics midpoint (v1.09) [kg PM2.5 eq.]	0.45	0.44	0.44	0.45
Photochemical ozone formation midpoint, human health (v1.09) [kg NMVOC eq.]	5.3	5.3	5.3	5.2
Resource depletion water, midpoint (v1.09) [m ³ eq.]	7.8	10.0	9.8	7.1
Resource depletion, mineral, fossils and renewables, midpoint (v1.09) [kg Sb eq.]	0.03	0.03	0.03	0.03

Table 4. Impact assessment per 1 tonne of methanol and 1.96 tonnes of cement produced in Germany in a conventional methanol plant (using system expansion to include the impacts associated with a cement plant).

Impact categories (ILCD/PEF recommendation v1.09)	CCU Methanol Plant			
	Status-quo	Low decarbonised	High decarbonised	Fully decarbonised
Acidification midpoint (v1.09) [Mole of H+ eq.]	36.5	7.7	6.8	7.9
Climate change midpoint, excl biogenic carbon (v1.09) [kg CO2 eq.]	7376	2009	1042	700
Climate change midpoint, incl biogenic carbon (v1.09) [kg CO2 eq.]	7849	2018	859	725
Ecotoxicity freshwater midpoint (v1.09) [CTUe]	70067	3732	3736	205606
Eutrophication freshwater midpoint (v1.09) [kg P eq.]	9.7	0.19	0.19	0.76
Eutrophication marine midpoint (v1.09) [kg N eq.]	5.8	2.3	2.0	1.59
Eutrophication terrestrial midpoint (v1.09) [Mole of N eq.]	128.9	23.9	20.9	17.1
Human toxicity midpoint, cancer effects (v1.09) [CTUh]	0.001	0.00003	0.00003	0.0002
Human toxicity midpoint, non-cancer effects (v1.09) [CTUh]	0.002	0.0003	0.0003	0.001
Ionizing radiation midpoint, human health (v1.09) [kBq U235 eq.]	1342	943	934	81.9
Land use midpoint (v1.09) [kg C deficit eq.]	3123	3474	3920	2293
Ozone depletion midpoint (v1.09) [kg CFC-11 eq.]	0.0003	0.0001	0.0001	0.00008
Particulate matter/Respiratory inorganics midpoint (v1.09) [kg PM2.5 eq.]	1.64	0.49	0.47	0.83
Photochemical ozone formation midpoint, human health (v1.09) [kg NMVOC eq.]	9.4	5.9	5.1	4.7
Resource depletion water, midpoint (v1.09) [m ³ eq.]	106.3	432.7	390.8	8.4
Resource depletion, mineral, fossils and renewables, midpoint (v1.09) [kg Sb eq.]	0.05	0.05	0.06	0.09

Table 5. Impact assessment per 1 tonne of methanol and 1.96 tonnes produced in Germany in a CCU methanol plant (using system expansion to include the impacts associated with the cement plant which provided the CO₂).

Figure 4 below presents the greenhouse gas emissions graphically, making clear the limited impact that the electricity scenarios have upon the conventional process but the significant role that the carbon intensity of electricity plays in the greenhouse gas emissions from the CCU process.

The same figure allows direct comparison of the greenhouse gas emissions from the CCU process compared to the conventional SMR process. If current grid mix electricity (status-quo) is used for both processes, the greenhouse gas emissions of the CCU process are several-fold higher than the emissions from the conventional process. However, if low-carbon electricity is used, then the CCU process has the potential to reduce greenhouse gas emissions significantly.

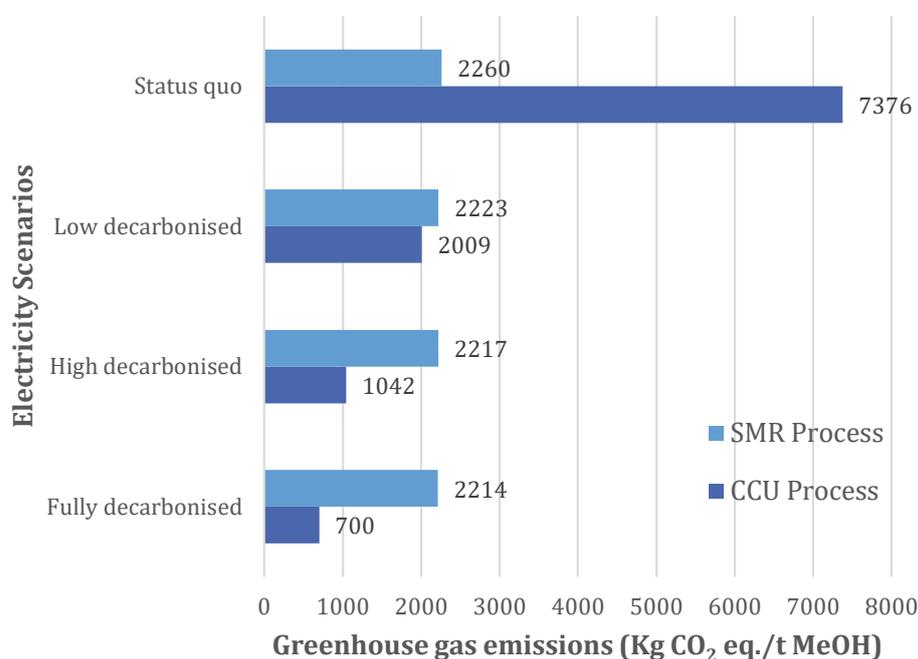


Figure 4. Greenhouse gas emissions of the conventional SMR process compared to the CCU methanol process using different electricity grid scenarios.

Comparing the impacts of the two processes reveals that the greenhouse gas emissions arising from the CCU process are 9.6% lower than the conventional process when the low decarbonized scenario is used for both processes. If the high decarbonized scenario is used for both, then the CCU process shows 53% lower greenhouse gas emissions than the conventional process. If fully decarbonized electricity (modeled as onshore wind turbines) is used for both processes, then greenhouse gas emissions from the CCU process are 68% lower than if the conventional process is used.

This worked example does not aim to provide a template for how to write an LCA report, it attempts to focus upon problems and issues specific to CCU. Since the most common reason for implementing CCU technologies is to reduce the greenhouse gas emissions of the underlying process, the emphasis of the impact assessment presented is upon the climate change impact category. The other environmental impacts have been mentioned here only briefly as examples of what would need to be highlighted and discussed in an LCA report. However, as with any LCA, an assessment needs to be made of which impact categories need to be discussed and analysed and the reasons given.

Comparison between the conventional SMR and the CCU processes when utilising the status-quo grid mix electricity reveals that the CCU process results in higher impacts across most of the environmental impact categories.

The effect upon the environmental impacts of both processes of changing the electricity scenarios is varied. When using status-quo scenario electricity, the acidification midpoint was 7.3 mol H⁺ eq. for the SMR process but 36.5 mol H⁺ eq. for the CCU. However, fully decarbonising the electricity brought the figure obtained for the CCU process down to a similar level as for the SMR process. The same pattern of response can be seen for terrestrial eutrophication.

Water usage (resource depletion water impact category) shows a similar dependency, with the SMR process having an impact of 7.8 m³ eq. compared to 106.3 m³ eq. for the CCU process. After switching to electricity sourced from onshore wind turbines the water use of the SMR process reduced slightly to 7.1 whereas that for the CCU process fell dramatically to a more similar 8.4 m³ eq. However, the interpretation is complicated by the fact that water usage of the CCU process rises to 432.7 and 390.8 m³ eq. when the low and high decarbonized scenarios are used. This likely reflects the lower water use efficiency of some of the renewable technologies included within these scenarios.

Freshwater ecotoxicity for the SMR process was 3,354 CTUe (comparative toxic units, ecosystems) when using the status-quo electricity scenario, this reduced to 2,901 when the low and high decarbonized scenarios were applied but increased to 4,279 when the system was modelled using electricity from onshore wind. The CCU process resulted in higher freshwater ecotoxicity values of 70,067 CTUe when using grid mix electricity, which reduced to more comparable values of 3,732 and 3,736 when the 2030 and 2050 scenarios were applied. However, when the electricity source was switched to 100% onshore wind the value increased significantly to 205,606 CTUe.

C.6.2.2 Contribution analysis

An analysis was conducted in order to reveal the contribution of each of the four main processes that form the CCU methanol process (CO₂ capture, CO₂ transport, H₂ production by electrolysis and methanol production) made towards the overall environmental impact. The proportion of these impacts that was due to electricity usage as opposed to the other main operational inputs into the processes such as water use, wastewater treatment and catalysts was also evaluated.

The electricity modelled for the analysis was 100% wind power, because, as will be shown, the use of current (2018) grid mix electricity to run the processes involved with a CCU methanol plant would be unlikely.

This contribution analysis includes only the indirect emissions. The direct CO₂ inputs and emissions are not included in the following charts, for example the 1669 Kg CO₂/t methanol input into the CO₂ capture unit (which would otherwise have to be treated as a negative emission) or the emission of 217 Kg CO₂/t methanol of CO₂ which was not captured (reflecting the 87% capture efficiency modelled) are not included. Similarly, the emission of 77.7 Kg CO₂/t methanol arising from the methanol plant which reflects the 94.5% conversion efficiency of CO₂ to methanol within the plant was not included. The contribution analysis was done in this way to enable the climate impacts of the electricity and other inputs to be determined alone, rather than allowing the input and emission of captured CO₂ through the system to distort the analysis. Figure 5 overleaf presents a pie chart reflecting the contribution of the four main steps involved in the modelled CCU process.

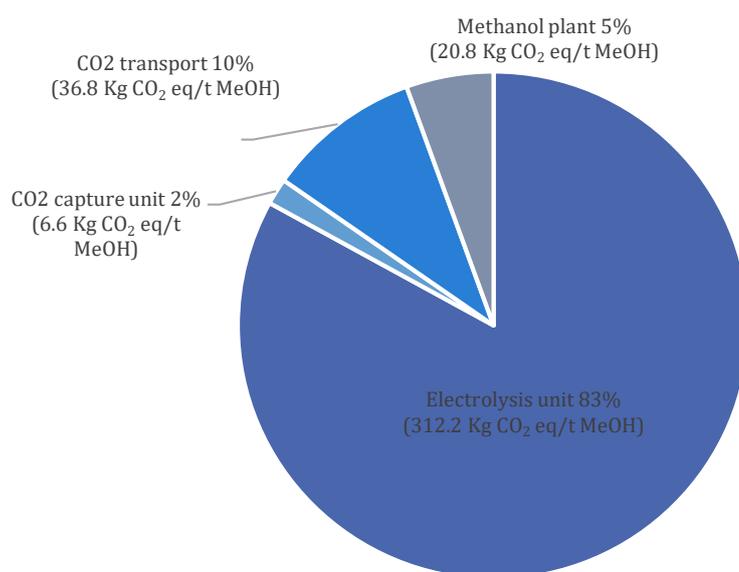


Figure 5. The indirect global warming impacts of the four main processes identified as components of the CCU methanol production process.

Despite the fact that renewable electricity was used, the contribution made by electricity usage compared to the other inputs was dominant, as presented in Figure 6 below.

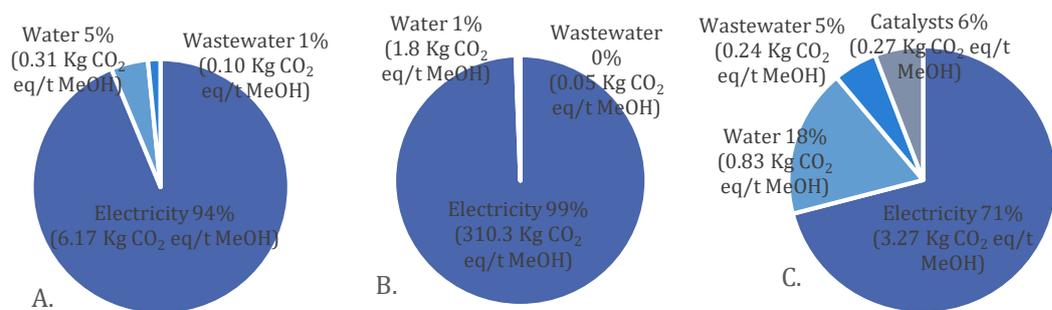


Figure 6. The contribution that the main process inputs make to the indirect global warming impacts of the component processes of the CCU methanol production. A. CO₂ capture unit, B. electrolysis unit and C. methanol plant.

C.7 Life Cycle Interpretation



Reference to LCA Guidelines

A description of what should be included within this section of an LCA report is provided in **Section C.7 Life cycle interpretation** of the guidelines, with **Guidelines C.12** and **C.13** specifying what is required.

In addition, a checklist of items to be included in each section of the report is included in **Section C.8 Reporting** of the guidelines, which for the life cycle interpretation is as follows:

Life cycle interpretation

- Include and describe the results
- Negative emission in cradle-to gate studies shall not be interpreted as CO₂ sinks if life does not end with permanent carbon fixation
- Emission reductions due to substitution effects shall be interpreted as environmental benefits but not as negative emissions
- Describe uncertainty and sensitivity analysis and report results separately
- Include completeness check
- Include consistency check
- State assumptions and limitation associated with the interpretation of results
- Include conclusions
- Include recommendations, if any

It is clear from the data presented in Figure 4 that switching methanol production from the conventional SMR process to the CCU process reduces the global warming impacts only if low carbon electricity is used. If that is the case, then significant savings can be made.

No attempt is made here to explain the variations in results of the impact categories obtained when the electricity scenarios were changed, since as mentioned previously, the focus is on issues specific to CCU. Such explanations are readily available in other LCA comparisons of electricity sources.

C.7.1 Sensitivity analysis – location of the facility

It is common practice when performing LCA of industrial processes to determine the sensitivity of the environmental impacts of the process to the country where the facilities are located. In this case, the baseline process was evaluated in Germany, so it would make sense to compare the same process in other countries. However, when utilizing CO₂ derived from another industry (such as a cement or power plant) the LCA practitioner needs to be aware that the efficiency of the CO₂ source may differ in different countries. This has implications, since it means that the amount of co-product produced per unit CO₂ varies. The following section provides a worked example of how this impacts LCA.

The CCU methanol plant modeled requires 1.45 t CO₂ to produce 1 t methanol. At a capture efficiency of 86.98%, this means that 1.669 t CO₂ per hour is required from the cement plant. Cement plants in the ecoinvent 3.4 database differ in efficiency between countries. For instance, Portland cement produced in Switzerland results in 0.729 Kg CO₂/Kg cement, but in Germany this is 0.852 and in USA it is 0.883 Kg CO₂/Kg cement. This means that a cement plant in Germany will manufacture 1.959 t cement while producing the required 1.669 t CO₂ (as was used in the functional unit of this report) but a similar plant in Switzerland would manufacture 2.289 t cement while producing the same quantity of CO₂ whereas a plant in the USA would produce only 1.890 t cement when producing the same quantity of CO₂.

The differing productivity of co-products per tonne CO₂ produced in different countries means that there needs to be some consideration of how a consistent functional unit can be obtained if doing such a comparison. If the same 1.959 t cement that was produced in Germany is manufactured in Switzerland and USA, then the highly efficient Swiss cement plant would not produce enough CO₂ to produce the 1 tonne methanol required by the functional unit, whereas enough CO₂ would be produced in USA to produce more than 1 tonne methanol.

Since a consistent functional unit is required to enable comparison of the impacts of the CCU processes in different countries, the suggested approach is to model the production of the mass of cement in the most efficient country which produces the required 1.669 t CO₂. In this example Switzerland has the most efficient cement plants, where 2.289 t cement needs to be produced to generate the 1.669 t CO₂. Therefore, a functional unit of 1 t methanol and 2.289 t cement is used, which allows all countries to produce enough CO₂ for the methanol production.

However, due to their lower efficiencies, producing this quantity of cement generates an additional 281 Kg CO₂ in Germany and 352 Kg CO₂ in USA, over and above that required to produce the 1 t methanol. This “additional” CO₂ is treated as being emitted to air by the cement plant because the same 1.669 t CO₂ is required to enter the CO₂ capture unit irrespective of the location.

Table 6 presents the LCIA results of this analysis, which assumes that the electricity used for all processes is decarbonized as used previously. The same system expansion approach was used to include the impacts of the cement plant as used previously. Figure 7 presents the climate change impact category (excluding biogenic carbon) resulting from this analysis as a graph to enable comparison of the conventional (SMR) versus CCU process in the three countries studied. This shows that the climate impact of the SMR process was 21% lower if the process was modelled in Switzerland rather than in USA. Similarly, the climate change impact of the CCU process was almost 25% lower if sited in Switzerland than in Germany (due to German onshore wind having a higher carbon intensity than US onshore wind, according to ecoinvent 3.4). Therefore, the production of methanol does not only change over time but is also depended on the regional setting.

The reduction in climate impacts achieved by switching from conventional steam methane reforming to the CCU methanol process was 63% when located in USA or Switzerland and 61% when located in Germany.

Impact categories (ILCD/PEF recommendation v1.09)	Germany		Switzerland		USA	
	SMR	CCU	SMR	CCU	SMR	CCU
Acidification midpoint (v1.09) [Mole of H+ eq.]	7.8	8.5	6.5	7.8	5.5	7.4
Climate change midpoint, excl biogenic carbon (v1.09) [kg CO2 eq.]	2501	986	2029	741	2582	948
Climate change midpoint, incl biogenic carbon (v1.09) [kg CO2 eq.]	2532	1016	2133	842	2613	977
Ecotoxicity freshwater midpoint (v1.09) [CTUe]	4646	205972	4292	229369	6438	141612
Eutrophication freshwater midpoint (v1.09) [kg P eq.]	0.21	0.78	0.21	0.85	0.37	0.74
Eutrophication marine midpoint (v1.09) [kg N eq.]	1.40	1.76	1.11	1.54	1.47	1.66
Eutrophication terrestrial midpoint (v1.09) [Mole of N eq.]	16.3	19.1	13.6	17.0	16.2	17.7
Human toxicity midpoint, cancer effects (v1.09) [CTUh]	0.00003	0.0002	0.00002	0.0002	0.00004	0.0001
Human toxicity midpoint, non-cancer effects (v1.09) [CTUh]	0.0002	0.001	0.0001	0.001	0.0002	0.001
Ionizing radiation midpoint, human health (v1.09) [kBq U235 eq.]	73.5	91.7	125.2	148.6	74.7	84.3
Land use midpoint (v1.09) [kg C deficit eq.]	1916	2413	1269	1975	2110	2013
Ozone depletion midpoint (v1.09) [kg CFC-11 eq.]	0.0001	0.00008	0.0001	0.00005	0.0001	0.0001
Particulate matter/Respiratory inorganics midpoint (v1.09) [kg PM2.5 eq.]	0.49	0.88	0.33	0.78	0.88	1.16
Photochemical ozone formation midpoint, human health (v1.09) [kg NMVOC eq.]	5.7	5.1	5.0	4.6	4.9	4.7
Resource depletion water, midpoint (v1.09) [m ³ eq.]	7.9	9.1	5.2	8.0	9.5	7.5
Resource depletion, mineral, fossils and renewables, midpoint (v1.09) [kg Sb eq.]	0.03	0.10	0.02	0.10	0.03	0.07

Table 6. Impact assessment per tonne methanol produced in a conventional (SMR) methanol plant compared to a CCU methanol plant using electricity generated by wind turbines. System expansion was used to include the impacts associated with the cement plant which provided the feedstock-CO₂.

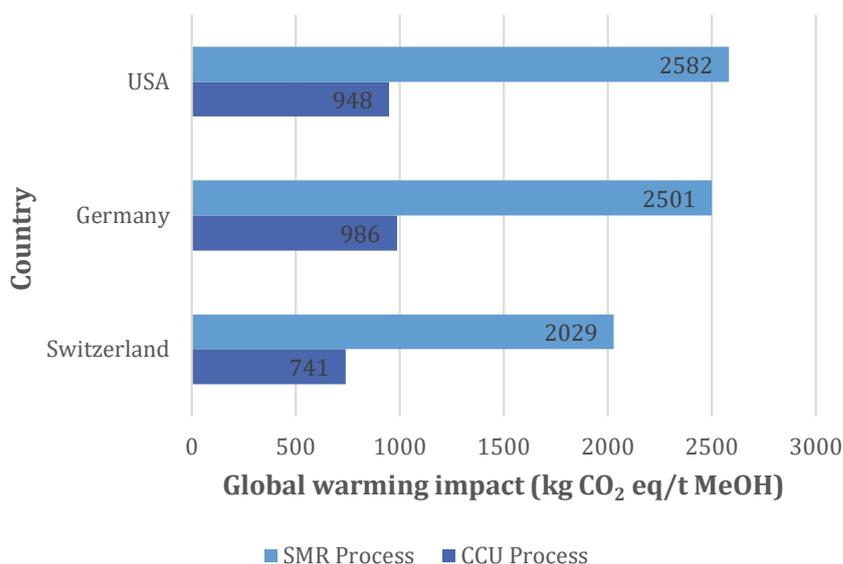


Figure 7. Comparison of the global warming impacts of the conventional SMR process vs the CCU process in USA, Germany and Switzerland. Wind power was used for all processes.

It should be noted that the global warming impact results for both the SMR and CCU processes in Germany presented in **Figure 7** are higher than those presented previously in **Figure 4** in **Section C.6** of this document. This is due to 2.289 t cement used in the functional unit in this part of the study, as opposed to 1.959 t used previously. This additional mass of cement was previously calculated to account for an extra 281 Kg CO₂, reflected by the climate impact increase of 286 Kg CO₂ eq. This discrepancy has been retained to enable the reader to see the impact of using a different functional unit within different parts of the same study.

For the reasons highlighted above, only one functional unit should be used in an LCA study. If an analysis involving comparisons between different countries is to be used, then all sections of the analysis need to be conducted using the same functional unit, i.e. in this report the functional unit involving 2.289 t cement would need to be used from the start. However, to justify the additional CO₂ production apparent at the start of the assessment, the country comparison would need to be made one of the goals of the study, rather than just an example as provided here.

Additionally, the use of the countries from different regions as comparator country has impacts upon the impact assessment methodology used for the study. ILCD is appropriate for studies within Europe, but if the USA is to be included as a comparator country then the CML assessment methodology should be used.

C.7.2 Limitations

The accuracy of this study was limited by the lack of modelling of the infrastructure for either the CO₂ capture unit or the electrolysis unit (as recorded in Section C.4.2.2 and Table 1). An infrastructure

component was added to the model for the CCU methanol plant itself. The database values for the conventional (SMR) methanol plant that the CCU plant is being compared to will be included within the ecoinvent 3.4 data used. This will result in a slightly unequal comparison, with a lower infrastructure contribution to the CCU process than the SMR process. The results should therefore be considered to be limited in their accuracy.

All weaknesses in modelling or data availability should be identified to enable the reader to determine their significance upon the interpretation of the results.

In *Section C.4.2.3* and Table 2 of this report it was highlighted that some of the reference process data from ecoinvent v3.4 utilized in modeling is several years old. This adds a level of uncertainty since the reference processes used in reality today may be slightly more efficient than those used in the models. If this is the case, then the observed reduction in global warming impacts gained by switching from conventional production to the CCU process may be smaller than the results suggest. However, the scale of the reductions in global warming impacts (in Germany, CCU production caused climate impacts up to 68% lower than conventional SMR if both processes use electricity from wind power) are so great that a small increase in efficiency of the conventional process would only slightly reduce the large CO₂ eq. reductions obtained by switching to the CCU process.

Another area of uncertainty was that caused the use of low TRL membrane capture system. However, a sensitivity analysis (described in the Integration worked example provided as Part D of this document) revealed the relatively small effect upon global warming impacts that changes to the efficiency of the capture system cause.

In addition, the focus of this worked example analysis has been the effect of the CCU process upon the global warming impacts. The results of the other impact categories have been presented as tabulated data but not presented graphically and analyzed in any detail. If a full life-cycle assessment was to take a similar approach and only present global warming impacts, this could potentially mislead decision-making since trade-offs between environmental impact categories cannot be identified by such a study.

C.7.3 Critical review of the study goal

The goal for this study was:

To compare the environmental impacts (with a focus upon global warming impacts) of producing methanol for use as a chemical feedstock in Germany, synthesised via two routes: the hydrogenation of CO₂ captured from a cement plant vs methanol synthesised using the conventional steam methane reforming process from natural gas.

The goal of this study was achieved. The environmental impacts of the two processes located in Germany were compared. Contribution analysis revealed that the most significant contribution to the impacts was

found to be electricity use, so different electricity scenarios were tested – comparing the status-quo with a low decarbonized, high decarbonized and a decarbonized scenario using 100% wind power.

As explained previously, if a comparison between countries is going to be conducted (either as part of the main study or just as part of a sensitivity analysis) then this should be captured within the goal of the study.

C.7.4 Interpretation and recommendations from a CCU perspective

The production of methanol from CO₂ does not provide a long-term net CO₂ sink as the chemical feedstock being produced has a limited lifetime before it is oxidized back to CO₂ and re-emitted to atmosphere. However, if low-carbon electricity is used, a significant reduction in global warming impacts is achieved if conventional (SMR) production is switched to CCU production. There are also no environmental impact reduction benefits from the CCU processes when comparing between the conventional SMR and the CCU processes if the current grid mix electricity is used. This may be expected from a technology with a very high electricity demand where most of the environmental impacts are related to the supply of electricity to the electrolyzer.

There are approximately 90 large-scale methanol plants globally [1]. Switching them from SMR-based processes to CCU-based processes to achieve the aims described above would lead to a significant increase in demand for low carbon/renewable hydrogen. Currently, producing low-carbon hydrogen requires large amounts of low-carbon electricity. If the efficiency of this process is increased, the impact reductions of switching to CCU processes would be even greater. However, assessing the impacts of a large-scale switch to CCU methanol was beyond the scope of this study.

In conclusion, life cycle assessment using system expansion was used to compare the environmental impacts of producing methanol from CO₂ captured from a cement plant with producing methanol from SMR. Modelling indicates that the global warming impact of CO₂-derived methanol is of the order of 60% lower than SMR-derived methanol if both processes use wind power. Comparison between countries reveals that the impacts are lowest if the plant is sited in a country where less CO₂ is emitted at the CO₂ source per main product, as more co-product is produced per quantity of CO₂ required to produce 1 t of methanol.

The results revealed that the global warming impacts were 68% lower than the impacts of conventional SMR production if electricity sourced from wind turbines was used for both processes. However, the limitations of the study discussed in **Section C.7.2** mean that it may not be wise to be too exact with the numerical results when reporting the final conclusions, as there is a degree of uncertainty imposed by the study limitations. A similar level of caution, reflecting the uncertainty in the results, was also applied in the executive summary. An uncertainty analysis would need to be conducted to enable the scale of this uncertainty to be determined.

Following from the above conclusions it is recommended that further analysis is done to reduce the recognized limitations of this study. Modelling of the impacts due to the infrastructure requirements of both the electrolysis unit and the CO₂ capture unit would be a useful step towards this goal. Low TRL membrane capture technology was modelled in the current study. It is therefore recommended that modelling is done using different (higher TRL) capture techniques to determine the sensitivity of the observed environmental impacts of the CCU methanol process to the CO₂ capture technology used.

C.8 References

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C.9 Appendix - Inventory

C.9.1 Process design and simulation basis

This section presents the process design and simulation of the showcase methanol plant. For this purpose, Aspen Plus simulator was employed to execute the necessary mass and energy balances. A sequential-modular approach was adapted in which the equations describing each process unit (module) were solved module-by-module in a stepwise manner. Iterative techniques were then used to solve the problems arising from the recycle of information. Four models were employed to determine the thermodynamic properties: Redlich-Kwong-Soave with Huron-Vidal mixing rules for streams at high pressure (>10 bar), and NRTL-RK for streams at low pressure (<10 bar), electrolyte-NRTL for electrolysis reactions and typical steam tables for the CHP units.

For compressors and gas turbines, mechanical and isentropic efficiencies are 95% and 90% respectively, while for pumps overall efficiency is set equal to 70%. For high pressure, intermediate pressure and low pressure steam turbines, isentropic efficiencies are 92%, 94% and 88% respectively [2]. Pressure drop in heat exchangers is typically dependent on phase. Usually liquid phase pressure drop is absolute and does not depend on relative pressure of the liquid. Liquid phase pressure drop for cold and hot side was set equal to 0.3 bar and gas phase pressure drop for cold and hot side 3% [3]. In addition, a minimum temperature difference of 7°C was applied for heat transfer. When, during heating exchanging, phase change occurs then at least three heat exchangers were used; two to count for the sensible heat above and below the temperature at which the phase change occurs and one for the latent heat. Open-recirculating cooling systems, that utilise the evaporation process to provide process or comfort cooling, were considered in the present study. A 5% loss was assumed to count for drift, evaporation and blow down losses. The temperature range of cooling water is typically 15-25°C. Furthermore, for a condensate return of 80% (of generated steam), the make-up rate of feed boiler water should be expected to be around 25% of the recirculated rate [4].

The system to produce methanol is divided into three main areas:

1. CO₂ capture
2. Hydrogen production
3. Methanol synthesis

C.9.2 CO₂ capture

Among the different carbon capture processes, membrane separation generates growing interest. Membrane separation can handle low to moderate dilute CO₂ streams and achieve high purity of CO₂ capture streams. However, based on the existing materials performances, the target purity (>99.5%) cannot be attained by a single or multiple stages [5]. CO₂ can be also separated from other gases by cooling and condensation. Cryogenic separation is widely used commercially for streams that already have high CO₂ concentrations (typically >90%) but it is not used for more dilute CO₂ streams. This suggests that a synergy between the two separations processes is possible and their combination could lead to a possible optimum with a minimal overall requirement. The study focuses on decarbonising heavy industries rather than power generation units as the latter are projected to shut down in the next few years while the market for the former is expected to grow over the foreseeable future.

There are two distinct sources of CO₂ associated with a cement kiln. The first is the result of the direct combustion of fossil fuels for heat energy which produces a flue gas of 4–15 vol% CO₂ depending upon the

source of fuel. The second is the by-product of calcining of limestone to produce the clinker material [6], where stoichiometric quantities of carbon dioxide are produced according to the reaction.



Both of these sources emit roughly equal amounts of CO₂ per tonne of cement produced and therefore when combined generate a flue gas that is more concentrated in CO₂ than is generally found in the power generation industry [7]. Typical total flue gas concentrations range from 22 to 28 mol%, with 900 kg of CO₂ emitted during production of 1000 kg of cement [5].

A single-stage membrane unit cannot achieve a high CO₂ capture ratio (>80%) and CO₂ purity (>95%) simultaneously, as reported by He et al. [8], and energy efficiency could be improved using multiple-stage membrane systems to reduce the irreversibility of the whole process, as documented by Zhang et al. [9]. Based on the design of He et al. [10], a two-stage cascade membrane system was considered in the present study. The first stage membrane unit is used for pre-concentration of CO₂ up to 50–70%. The concentrated CO₂ stream is then compressed to a certain pressure (e.g., 2–3 bar) and fed into the second stage membrane unit for ultimate CO₂ purification to achieve high CO₂ purity (>95%). In order to document the process and economic feasibility of membrane systems the following assumptions were made: 1) A membrane gas separation processes based on MTR (Membrane Research and Technology) Polaris™ membrane at a feed and permeate pressure of 2.5 bar and 250 mbar (optimal pressure reported in [8]) was employed, 2) a CO₂/N₂ and CO₂/O₂ selectivity of 50 and 20 were [11]. Selectivity of CO₂/H₂O is assumed as unity and 3) pressure drop between retentate and feed streams in conjunction with temperature difference between feed and permeate sides due to the Joule–Thompson effect are assumed to be negligible.

Flue gas is assumed to be delivered free of NO_x and SO_x components. Initially, flue gas is cooled below dew point, water is then removed and subsequently compressed to 2.5 bars. Before each membrane flue gas is conditioned so as to meet the pressure conditions (via vacuum pumps and compressors), monitor temperature at 35°C and where necessary remove condensed water. The permeate stream exiting the second membrane has a 95% CO₂ purity. Vent streams are collected and due to the high pressure of the mixed stream, electricity can be generated in an expander. Prior to this step, heat integration is realised with the aim of increasing the temperature (and consequently the enthalpy content) of the vent stream. For this purpose, three heat exchangers are added as depicted in **Figure 4** (red circles) to recover heat from the gas streams exiting, Compr1, Compr2 and Compr3 as well as VPump2. The integrated design reduces the cooling duties by 27.2% and the electricity demand by 32.9%. The next step of the carbon capture system configuration involves the cryogenic treatment of the CO₂ rich stream. According to Belaissaoui et al. [12], highly concentrated CO₂ streams (>90%), should be pressurised up to 11 bar before entering the economiser unit; a two-stage compressor with intermediate cooling is utilised. Afterwards, the gas is cooled down by water and cooled down further in the economiser to -73°C. The gas goes through a valve in order to reach the minimal necessary liquefaction pressure. The inlet pressure flash is set in order to have a CO₂ partial pressure of 5.4 bar ensuring a proper liquefaction of CO₂ and non-frosting conditions. After being heated in the economizer, the incondensable gases are mixed with the retentate streams and sent to an expander to generate electricity while the CO₂ enters a flash unit to remove water and is sent to the methanol synthesis plant. Overall, a CO₂ purity of 99.6% is achieved along with a capture ratio of approximately 87%.

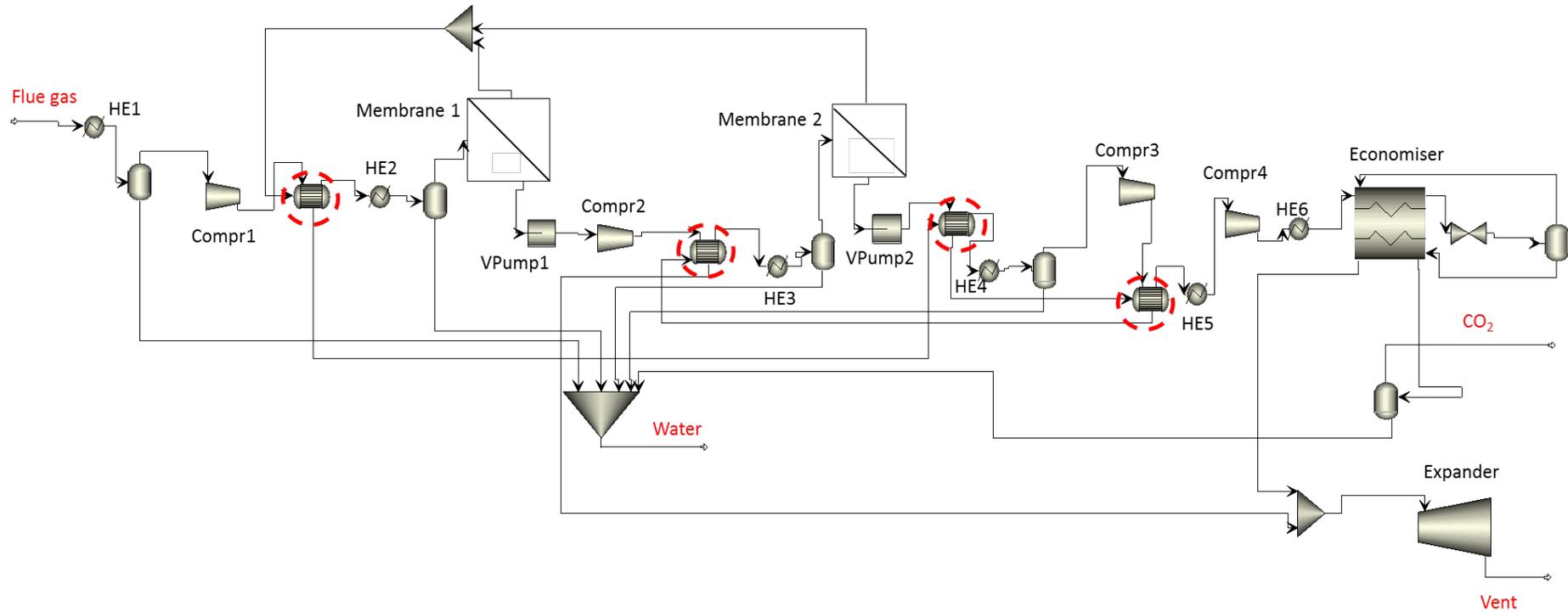


Figure 4. Process flow diagram for CO₂ capture through a multistage membrane system.

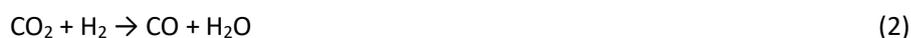
C.9.3 Hydrogen production

Hydrogen is produced via electrolysis by passing electricity through two electrodes in water. The water molecule is split and produces oxygen at the anode and hydrogen at the cathode. A proton exchange membrane (PEM) electrolyser was selected over an alkaline process as it provides higher efficiencies and also serves to separate the hydrogen and oxygen gases, as oxygen is produced at the anode on one side of the membrane and hydrogen is produced on the opposite side of the membrane [13]. Additionally, PEM electrolysers operate very well under variable loads further efficiency enhancements are anticipated in the coming decade [14]. The membrane allows the H^+ ion to transfer from the anode side of the membrane to the cathode side, where it forms hydrogen. An efficiency of 70% based on H_2 LHV [15] was considered and the required electricity is derived for the base study from the grid. An additional 10% is supplied to cover the demand of the balance of the electrolysis (BOE) section [16]; the overall electrical consumption is 52.4 kWh/kg H_2 . The lifetime of the electrolyser is 80,000 h [17]. In order for the process to be sustainable, it should be noted that green electricity is a prerequisite and therefore either renewable energy certificates and/or substantial penetration of renewable energy in the grid are required.

The PEM unit was modelled by employing two RSTOICH reactors both operating at temperature and pressure of 80°C and 35 bar respectively. The first represents the anode ($H_2O \rightarrow 1/2 O_2 + 2H^+ + 2e^-$) and the second the cathode ($2H^+ + 2e^- \rightarrow H_2$). A 95% water conversion is realized while the electricity demand was calculated by FORTRAN statements based on the necessary amount of hydrogen. Additionally, it should be noted deionized water is utilized. Oxygen is separated via a common separator (100% efficiency was assumed), then cooling to 30°C is applied and subsequently condensed water is removed. Purified hydrogen is recovered and is sent to methanol production plant. It was assumed that electrolysis occurs on site and no transport is required. Additionally, for a constant electricity load a constant H_2 production can be assumed and as so no buffering is included for the base case study.

C.9.4 Methanol synthesis

The process design for the methanol plant is based on previous works conducted by Van-Dal and Bouallou [18], Perez-Fortes et al. [19] and Kiss et al. [20]. The plant produces MeOH at a rate of 42.78 t/h (1.02 kt/day). CO_2 is delivered to the plant at a flowrate of 62.08 t/h, pressure of 5.4 bar and temperature of 25°C. The required H_2 (3:1 $H_2:CO_2$ molar ratio) produced from the PEM unit is supplied at a pressure of 35 bar and temperature of 25°C. The first step of the process incorporates the conditioning of the reactant gases to the operating conditions of the synthesis reactor, i.e. 78 bar and 210°C. CO_2 is compressed via a multistage compressor consisting of two compressors with intermediate cooling. If the gas is cooled to the inlet or slightly higher (~35°C) temperature of the previous lower pressure compressor it can be shown that minimum specific work is consumed when the pressure ratios (3.8:1 in our case) of the compressors are equal [21]. A single stage compressor was utilised to raise H_2 pressure up to 78 bar. The two gases are mixed and then re-mixed (for convergence purpose) with the recycle stream. The stream is then heated to 210°C and injected into the fixed bed adiabatic reactor. The adiabatic reactor is packed with a fixed bed of 44.5 tonnes of Cu/ZnO/ Al_2O_3 commercial catalyst. For this catalyst, the model proposed by Vanden Bussche and Froment [22] with readjusted parameters of Mignard and Pritchard [23] is able to describe with good precision the reactions of methanol production, Eq. (1) and the RWGS reaction, Eq. (2).



The kinetic model is described with Langmuir–Hinshelwood kinetics and it can be found in [18]. The model was directly implemented in Aspen Plus and pressure drop was calculated by the Ergun equation, already nested in the simulator. For a pellet density of 1,775 kg/m³ and a void fraction of 0.5 the volume of the

reactor is approximately 50 m³. It was observed that steady state can be reached at lower volumes meaning that less amount of catalyst can result in same methanol productivities, i.e. 35 tonnes. The gases leave the reactor at a temperature of 288°C and subsequently is divided into two streams: the first (60% of initial stream) is used to heat the fresh feed, while the second is used in the reboiler and also to heat the feed of the distillation column. The two streams are re-mixed and cooled to 35°C. Water and methanol, after being condensed, are separated from the non-reacted gases in a knock-out drum. Some of the non-reacted gases (1.3%) are purged to minimise the accumulation of inerts and by-products in the reaction loop. The crude methanol that leaves the reactor is a mixture of methanol, water and residual gases (i.e., H₂ and CO and CO₂). To remove the non-reacted gases, the stream is expanded to 1 bar using valves, and then separated in a flash tank. The remaining liquid is heated to 75°C and fed into the distillation column. The bottom product of the distillation column corresponds to water while the top product is mostly methanol with some unreacted gases. The column was simulated with the rigorous model RadFrac. A rate-based calculation of the column is carried out, using the mass and heat transfer model from Billet and Schultes. The column has 50 stages and the reflux ratio is equal to 1.2. However, the column can be intensified by reducing the reflux ratio to 1; reboiler's heating duty is reduced to 16.5 MW from 19.5 MW while methanol recovery and purity remain unchanged. Methanol is then compressed to 1.2 bar and cooled to 40°C proceeding to another flash that separates the gases (top outlet) from the methanol product with 99% w/w in the bottom stream.

Table 5 provides the simulated requirements of the CO₂ capture, electrolysis and methanol production units and **Figure 6** provides a process flow diagram of the methanol plant including the electrolysis unit.

C.9.5 Utilities

The streams containing unreacted gases were collected and sent to a gas turbine unit to generate electricity. The gas stream is then compressed up to 6 bar and enters a gas turbine where it is burned with excess of pressurised air (15%) to produce electricity at a temperature of 1200°C. The required air is specified by a FORTRAN calculator according to the flows of carbon monoxide and hydrogen. The exhaust gas from the gas turbine is recovered from a heat recovery steam generation (HRSG) system, composed of three heat exchangers (namely economizer, evaporator and superheater) where superheated steam (550°C, 100 bar) is produced. According to the present design, the economiser is placed before HE1 to extract heat from the pressurised CO₂. This way more steam can be produced as the exhaust gas provides heat only to the evaporator and the superheater. A series of high, intermediate and low-pressure steam turbines are employed to generate electricity. Part of high pressure steam is utilised to provide heat to the deaerator. The combined Rankine-Brayton cycle generates 4MW of electricity. Furthermore, based on the design of Van-Dal and Bouallou [18], hot water at a temperature of 85°C that is raised from HE1, HE2, HE3, HE4 and HE6 is mixed with the bottom stream of the distillation column and enters an Organic Rankine Cycle (ORC) unit. The working fluid is the R245fa and electricity of 1MW is generated. Finally, it should be noted that the wastewater unit and the cooling system were not simulated but considered in the economic evaluation.

Parameter	Value	Unit
CO₂ Capture		
Flue gas	224.27	t/h
CO ₂ input	71.39	t/h
CO ₂ capture	86.979	%
CO ₂ output	62.09	t/h
Electricity consumption	16.86	MW
Electricity generation	7.89	MW
Net electricity demand	8.97	MW
Cooling duty	16.1	MW
Cooling water (circulated)	1421	t/h
Cooling water (make up)	71.05	t/h
Waste water	9	t/h
Electrolysis		
Deionised water	80.85	t/h
Oxygen	68.2	t/h
H ₂	8.594	t/h
Electricity for electrolysis	450.12	MW
Cooling duties	2.81	MW
Cooling water (circulated)	248	t/h
Cooling water (make up)	12	t/h
Waste water	4	t/h
Methanol synthesis		
CO ₂ input	62.09	t/h
CO ₂ output	3.32	t/h
MeOH output	42.78	t/h
Cooling duties	43.13	MW
Heating duties (fully integrated)	31.3	MW
Electricity consumption	9.73	MW
Electricity generation	5	MW
Net electricity demand	4.73	MW
Cooling water (circulated)	3806	t/h
Cooling water (make up)	190.3	t/h
Boiler water (circulated)	6.4	t/h
Boiler water (make up)	1.6	t/h
Air	20	t/h
Catalyst	35	t/y
Waste water	24	t/h

Table 5. Simulation results for the CO₂ capture, PEM and methanol units.

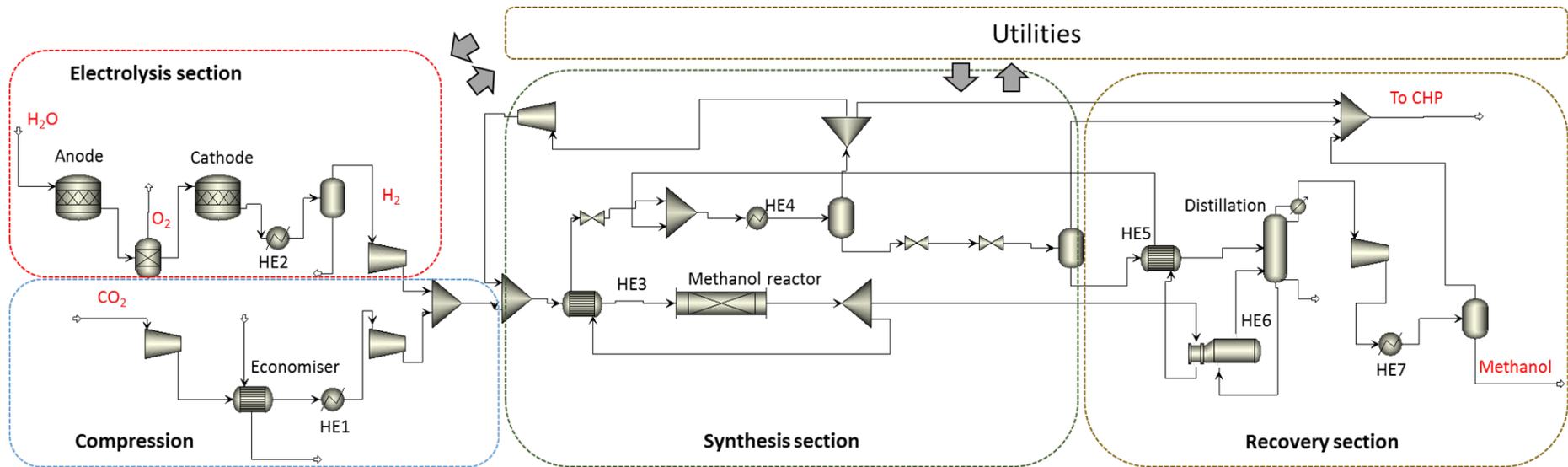


Figure 6. Methanol plant process flow diagram, showing the PEM electrolysis unit, conditioning of the resultant H_2 and captured CO_2 to the required temperature and pressure (78 bar, 210°C) prior to entry into the synthesis reactor. The CO_2 is compressed by two compressors with intermediate cooling prior to injection into the fixed bed adiabatic reactor containing $Cu/ZnO/Al_2O_3$ catalyst. The methanol containing gas leaving the reactor is used to heat the fresh feed and the feed to the distillation column. Methanol and unreacted gases are condensed and separated in a knock-out drum. Unreacted gases are separated by expanded to 1 bar using valves then separated in a flash tank before the liquid methanol is fed into the distillation column. The top product is mostly methanol while the bottom product is mostly water. The methanol is compressed and cooled to 40°C before proceeding to another flash tank to separate gases from methanol.

C.9.6 Inventory data sets for the main process inputs

In accordance with the functional unit, all inputs/outputs have been scaled to a final production of 1 t methanol and 1.96 t cement as final products. **Table 6** provides the input/output data for the CCU methanol production route whilst **Table 7** provides data for the conventional SMR from natural gas. Additionally, **Figure 7** presents the input/output data for the CCU methanol process in the form of a block diagram. The catalyst composition is taken to be CuO 60.1%, ZnO 28.3% and Al₂O₃ 11.3% [24].

It was assumed that the CO₂ was transported by road freight a distance of 50Km, giving the figure of 72.5 tKm per tonne methanol produced (i.e. 1.45 t CO₂ x 50 Km = 72.5 tKm).

Input flows	Amount
CO₂ capture	
Electricity (MWh)	0.209
Tap water (t)	1.66
Wastewater treatment (t)	0.21
Hydrogen production	
Deionised water (t)	1.89
Electricity (MWh)	10.52
Wastewater treatment (m ³)	0.1
Transport of CO₂ to industrial plant	
CO ₂ feedstock (t)	1.45
Assumed CO ₂ feedstock transport (tkm)	72.5
MeOH synthesis	
Tap water (t)	4.49
Electricity from grid mix/renewables (MWh)	0.111
Chemical plant (pcs)	9.8 x 10 ⁻⁸
Wastewater treatment (m ³)	0.56
Aluminum oxide (kg)	0.012
Copper oxide (kg)	0.062
Zinc oxide (kg)	0.029
Output flows	
CO₂ capture	
CO ₂ feedstock (t)	1.45
CO ₂ direct emissions (t)	0.22
H₂ production	
H ₂ feedstock (t)	0.20
Methanol synthesis	
Methanol (t)	1
CO ₂ direct emissions (t)	0.078

Table 6. Inputs/output data for CO₂-based methanol production system, based on 1 t methanol.

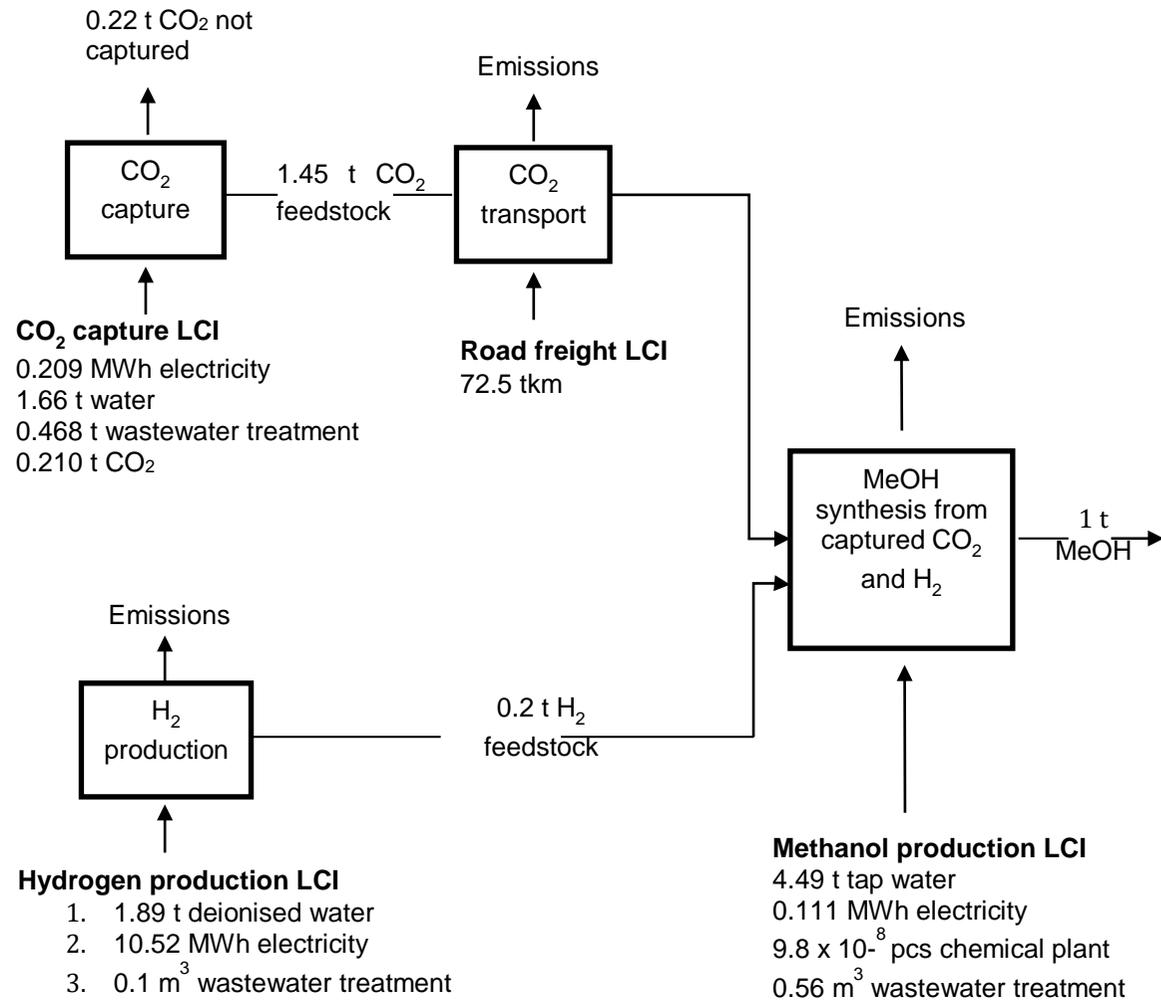


Figure 7. Block diagram of inputs and outputs for the CO₂-based methanol production system.

Input flows	Amount
Zinc (kg)	0.03
Natural gas (m ³)	0.652
Electricity from grid mix/renewables (MWh)	0.07
Methanol factory (pcs)	3.7 x 10 ⁻⁸
Copper oxide (kg)	0.1
Aluminium oxide (kg)	0.2
Heat, district or industrial natural gas (MJ)	6930
Molybdenum (kg)	0.01
Nickel (kg)	0.02
Deionised water (kg)	850
Output flows	Amount
Methanol (t)	1

Table 7. The input and output flows used by ecoinvent v 3.4 for conventional methanol production from natural gas.

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PART D Integration Worked Example

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D.1 Introduction

This section provides the combined results of the techno-economic assessment (TEA) and life cycle assessment (LCA) for the investigated sensitivity analysis of the CCU methanol process. The selected indicators are the minimal methanol selling price (MMSP) (€/t MeOH) and the climate change impact (kg CO₂/t MeOH) for the TEA and LCA respectively. The goal is to provide a holistic approach and deliver combined economic and environmental results. The presented methodology demonstrates the technical variables with the highest overall impact from both an economic and environmental perspective.

D.2 System boundaries

D.2.1 Options for dealing with the multi-functionality of the system

In the LCA worked example provided in **Part C** of this document, the multi-functionality of the system was solved by using system expansion. One feature of using system expansion with a multi-functional system is that it results in the functional unit with multiple functions (i.e. products). In the life cycle impact assessment described in Part B of this document, the functional unit was 1 t methanol and 1.96 t cement.

For a stand-alone LCA report, such a functional unit does not present a problem as LCA is commonly used in systems with multiple products. However, if the LCA study is to be integrated with a techno-economic assessment (TEA), problems can arise. TEA does not commonly use system expansion to include upstream processes, because the focus of TEA is the cost or market prices of the raw materials and products at the factory gates. If system expansion was used for TEA, it would necessitate the modelling of the processes which produce the CO₂ supply, requiring a significant increase in time and effort and access to data from another industrial plant which may not be available. To ease the integration of LCA and TEA, a method of achieving a functional unit with a single product is required. The following section describes the two options available for achieving this in accordance with the LCA Guidelines.

In the LCA worked example provided as **Part C** of this document, system expansion was used solve the multi-functionality of the system. However, whilst this approach is preferable for an LCA analysis, it results in a functional unit with two functions or products: 1 t methanol plus 1.96 t cement. Having cement in addition to methanol within the functional unit potentially makes the TEA challenging, so to enable a more straightforward integration of aspects of the LCA and TEA to be conducted, a second LCA analysis needed to be performed with cement excluded from the functional unit. This then allows the same functional unit to be used for the techno-economic analysis as for the environmental impact analysis. Two approaches to solving multi-functionality are discussed.



Reference to LCA Guidelines

Section C.4.3.2 Solving multi-functionality describes a hierarchy of approaches. According to this hierarchy, after system expansion, the next option is to try system expansion with substitution.

System expansion via substitution follows the same process for system expansion as described in Part B of this report, however when using this approach credit is given for the substitution of cement production without carbon capture as this enables the avoidance of the burdens caused by those CO₂ emissions. In practice, this means subtracting (or substituting) the impacts caused by the conventional cement

production from the system expansion model. The diagram in Figure 1 below shows a conceptual plan of how the impacts of a conventional cement plant are subtracted from the system expansion model. The system boundary of the resulting model includes only the methanol plant, meaning that the only product within the functional unit is methanol.

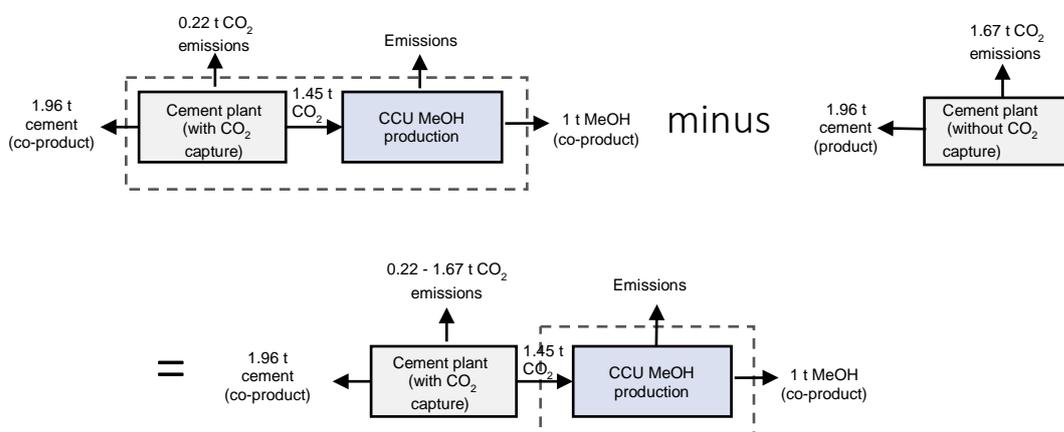


Figure 1. Representation of system expansion with substitution, whereby the impacts of a conventional cement plant are subtracted from the impacts of the system expansion model of a CCU methanol plant (including a cement plant with CO₂ capture), to give the impacts arising from the CCU methanol plant alone. The dotted lines indicate the system boundary.

By performing this subtraction, the impacts of cement production are effectively removed from the system boundary and therefore cement is removed from the functional unit. However, one outcome of the substitution of conventional cement production without carbon capture, with cement production with carbon capture is that once the life-cycle impact assessment (LCIA) is run the methanol process may appear to have a negative CO₂ emission. This occurs when the CO₂ emissions avoided by using carbon capture are greater than the CO₂ emitted by the methanol production process. We know from the LCA worked example in Part C of this document that such an outcome is highly likely. Such a “negative emission” requires clear interpretation to avoid misleading the reader.



Reference to LCA Guidelines

According to the hierarchy of approaches provided in **Section C.4.3.2 Solving multi-functionality**, the next approach to solving the multi-functionality following system expansion via substitution is allocation. Here the impacts of the CO₂ source (cement production in this case) are distributed between the CCU product (methanol) and the CO₂ production according to an underlying physical relationship or simply the mass, energy content or economic value of the products produced. In this way, the system can be viewed as being mono-functional, producing just one product.

An alternative approach to system expansion with substitution is allocation whereby the environmental impacts of the system are distributed amongst the functions. In practice, in this case this means allocating the impacts associated with the CO₂ production to the cement plant rather than to the methanol plant. By creating a model containing the CO₂ capture unit, the electrolysis unit and the methanol plant within the system boundary, but the cement plant is outside the boundary, we have essentially done just that. The results obtained by this technique are the same as those achieved by system expansion via substitution, but the conceptual process of how to get there is different, as representation by the diagram of the model as shown in Figure 2 below.

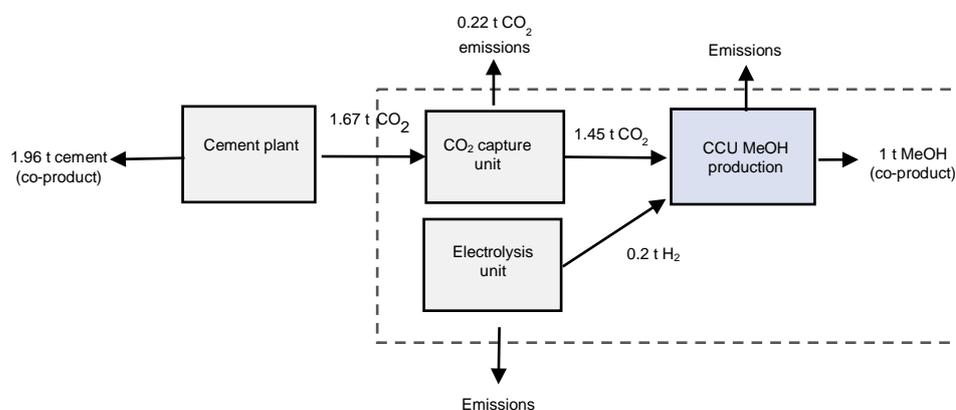


Figure 2. Representation of the system boundary (dotted line) when the impacts of producing the CO₂ from a cement plant are excluded from the system consisting of the CO₂ capture unit, the electrolysis unit and the CO₂-derived methanol plant.

By changing the system boundary in this way to include the CO₂ capture unit but not the source of the CO₂, this approach effectively allocates all the burdens of the CO₂ production to the cement plant and none to the methanol plant. It simply assumes that the CO₂ produced by the cement plant is a waste product which is an environmental burden “belonging” to the cement plant.

D.2.2 Outcome of the chosen method

System expansion via substitution was conducted using the data obtained by modelling the CCU methanol plant in Germany used in the LCA worked example (*Part C* of this document). The results of this approach are presented below in Table 1 overleaf.

The table allows comparison of the results obtained from system expansion (the preferred method of determining the life cycle impacts of CCU processes), with those results obtained from system expansion with substitution (which may be used to ease LCA integration with TEA). It can be seen that all impact categories show a reduction in impacts when the substitution method is used, but the climate change impacts show the most significant reduction and the substitution method gives a negative value.

When interpreting the results, it is important to remember that when using system expansion, the impact values reflect the impacts of both the methanol plant and the cement plant whilst only those from the

methanol plant are reflected when using system expansion with substitution. In addition, the latter technique assumes that the CO₂ emissions avoided at the cement plant by utilizing CO₂ capture are allocated solely to the methanol plant. In such a scenario, the cement plant cannot also claim to be producing low-carbon cement without “double-counting” any emissions reductions.

The negative climate impact value for the CCU methanol plant obtained by using the substitution approach is due to the technique effectively comparing between the process with and without carbon capture. The negative value obtained shows that the CCU process has a lower impact than the conventional process, but it does not mean that the CCU process is carbon negative (or even carbon-neutral). Rather, the value obtained is negative due to subtracting the avoided CO₂ emissions from the lower emissions resulting from the CCU plant itself. This needs to be made clear to avoid misunderstandings by policy-makers or other readers of the LCA report.

Impact categories (ILCD/PEF recommendation v1.09)	Impact value		
	System expansion (SE)	Cement plant (without capture)	SE with substitution
Acidification midpoint (v1.09) [Mole of H+ eq.]	7.9	3.9	4.0
Climate change midpoint, excl biogenic carbon (v1.09) [kg CO ₂ eq.]	700	1698	-998
Climate change midpoint, incl biogenic carbon (v1.09) [kg CO ₂ eq.]	725	1724	-999
Ecotoxicity freshwater midpoint (v1.09) [CTUe]	205606	2178	203428
Eutrophication freshwater midpoint (v1.09) [kg P eq.]	0.76	0.15	0.61
Eutrophication marine midpoint (v1.09) [kg N eq.]	1.59	1.01	0.58
Eutrophication terrestrial midpoint (v1.09) [Mole of N eq.]	17.1	11.8	5.3
Human toxicity midpoint, cancer effects (v1.09) [CTUh]	0.0002	0.00002	1.8E-04
Human toxicity midpoint, non-cancer effects (v1.09) [CTUh]	0.001	0.00013	8.7E-04
Ionizing radiation midpoint, human health (v1.09) [kBq U235 eq.]	81.9	58.1	23.8
Land use midpoint (v1.09) [kg C deficit eq.]	2293	717	1576
Ozone depletion midpoint (v1.09) [kg CFC-11 eq.]	0.00008	4.76E-05	3.24E-05
Particulate matter/Respiratory inorganics midpoint (v1.09) [kg PM _{2.5} eq.]	0.83	0.28	0.55
Photochemical ozone formation midpoint, human health (v1.09) [kg NMVOC eq.]	4.7	2.89	1.81
Resource depletion water, midpoint (v1.09) [m ³ eq.]	8.4	4.2	4.2
Resource depletion, mineral, fossils and renewables, midpoint (v1.09) [kg Sb eq.]	0.09	0.010	0.08

Table 1. Results of the LCIA of the CCU methanol plant sited in Germany, presented following system expansion with substitution. Impacts are presented per t methanol.

The LCIA conducted using Gabi and ecoinvent as part of the LCA worked example presented in **Part C** determined the climate change impact category, excluding biogenic carbon (which is measured in CO₂ equivalent and includes CO₂ plus other greenhouse gases such as methane and nitrous oxide) to be 700 Kg CO₂ eq. for the system expansion model. The data for all impact categories is presented again in Table 1 above so that it can be used for the system expansion via substitution calculation.

The same table also provides the LCIA results for a conventional cement plant sited in Germany. It can be seen that it emits 1698 Kg CO₂ eq. per 1959 Kg cement produced (i.e. the amount of cement manufactured, which produces the quantity of CO₂ required to produce 1 t methanol via CCU).

To obtain the results for system expansion via substitution, the LCIA results obtained from the conventional cement plant are subtracted from the LCIA results obtained when using system expansion. If we focus upon the climate change (excl. biogenic carbon) impact category as an example: $700 - 1698 = -998$ Kg CO₂ eq. / t MeOH.

An advantage of removing cement from the functional unit (in addition to enabling a more straightforward integration with TEA) is that the problem encountered with comparing CO₂-generating processes located in different countries as described in **Section C.7.1 Sensitivity analysis** of the LCA worked example is also avoided. Any additional CO₂ emitted by the cement plant (or whatever CO₂ source is used) due to differences in efficiency in different countries does not alter the results of the methanol plant results.

The drawback of this approach is that it effectively allocates all of the CO₂ emission savings obtained due to using carbon capture to the methanol plant and none to the cement plant. In this way, a negative greenhouse gas emission value is obtained for the methanol plant, but the cement plant in such a scenario has to be considered to emit the same emissions as a conventional cement plant, despite the fact that a CO₂ capture is being used. The negative value attributed to the methanol plant actually reflects the difference compared to the conventional system rather than reflecting a truly carbon negative methanol process. This needs to be made very clear in the conclusions drawn from the analysis to avoid misinterpretation by both the report author and the reader.

In summary, to allow an integrated analysis, the same functional unit needs to be used for the environmental analysis and the economic analysis. When conducting LCA of CCU processes, the widely accepted approach is to use system expansion to solve the multi-functionality associated with utilising CO₂ arising from one process to supply a separate CCU process. However, TEA does not commonly use system

expansion to include upstream processes as TEA tends to focus upon the cost or market prices of the raw materials and product at the factory gates. If system expansion was used for TEA, it would require modelling the processes which produce the CO₂. To avoid this, a method of achieving a functional unit with a single product was used – namely system expansion with substitution.

D.3 Integrated sensitivity analysis

Contribution and sensitivity analyses conducted as part of the separate LCA and TEA studies provided in *Parts B and C* of this document revealed that electrolyser efficiency is of crucial importance and a key contributor to both the economic and environmental performance of the CCU methanol process. Additionally, it is recognized that both the CO₂ capture ratio and the reactor operating pressure during methanol synthesis play a role in determining the methanol production efficiency. An integrated environmental and economic assessment of changes to the efficiency of these three processes has therefore been conducted.

The approach taken is the “alignment and combined indicators integration” as described in the **Integration Chapter** which is included within **Part A** of the associated Guidelines document.

D.3.1 Approach taken

To enable such an assessment to take place, the inventory data used for the economic and environmental analyses must be aligned and consistent and the functional units used must be the same to allow direct comparison of the two sets of results.



Reference to General Assessment Principles

Section A.5.2 Types of Study describes the different levels of TEA and LCA integration.

Section A.5.3 Alignment describes the data equivalency requirements to enable integration of TEA and LCA data and **Section A.5.4 Multi-functionality and system boundaries** suggests how to best align the system boundaries of the economic and environmental assessments.

In the original model, the efficiency of the PEM electrolysis unit was taken to be 70%, so the effect of this dropping to 60% or rising to 85% is modelled. Likewise, the CO₂ capture efficiency was taken to be 87%, so the effect of this reducing slightly to 85% or rising to 92% is evaluated. An additional analysis was conducted on the impact of changing the operating pressure of the methanol synthesis reactor. The baseline pressure was 78 bar, so this was varied from 50 to 100 bar to determine the effect of this upon the climate change impacts and the MMSP. Data obtained from Aspen modelling of these scenarios is provided in **Table 2**.

Electrolysis Efficiency	60%	70%	85%
Electrolyser electricity use (MW)	477.4	409.2	337.0
Balance of electrolysis plant (MW)	47.7	40.9	33.7
Overall electricity use (MW)	525.2	450.2	370.7
CO₂ Capture Efficiency	85%	87%	92%
Net electricity use (MW)	8.6	8.97	9.87
Cooling water (make-up) (t/h)	66.64	71.05	76.79
Membrane operating pressure (bar)	2	2.5	3
MeOH Synthesis Operating Pressure	50 bar	78 bar	100 bar
Electricity (MW)	3.76	4.73	5.4
MeOH Flowrate (t/h)	42.08	42.78	43.1
CO ₂ emissions (t/h)	4.1	3.32	3.02

Table 2. Changes in inputs resulting from changes to the modelled efficiency of the PEM electrolysis and membrane CO₂ capture systems, and changes to the operating pressure of the methanol synthesis reactor.

In the following section the results of the sensitivity analyses will be presented initially with the analysis of the greenhouse gas (GHG) emissions and the cost impact upon the methanol selling price separately. The integrated results will then be presented as an example of how such data could be presented in a single graphical representation.

D.3.2 Electrolyser efficiency

The contribution analysis in *Section C.6.2.2 of the LCA Worked Example* revealed that even when utilizing low-carbon electricity derived from wind turbines, H₂ production by electrolysis contributes 83% of the GHG emissions of the entire process due to its high electricity demand. It is therefore to be expected that modelled changes to the efficiency of the PEM electrolysis unit would have a significantly greater impact than the modelled changes to the efficiency of the CO₂ capture unit.

This is indeed what is found, with Figure 3 showing that GHG emissions decrease by 5.5% (or 55 Kg CO₂ eq/t MeOH) when the electrolyser efficiency is increased from 70% to 85%. When the electrolyser efficiency is decreased from the baseline 70% to 60%, the GHG emissions increase by an almost identical 5.2% (corresponding to 52 Kg CO₂ eq/t MeOH). Overall, the increase in electrolyser efficiency from 60 to 85% results in a decrease in GHG emissions of 11.3% (or 107 Kg CO₂ eq/t MeOH).

Similar to the assessment of GHG emissions, the assessment of minimal methanol selling price (MMSP) for the same sensitivity analysis was carried out. Figure 4 reveals that increasing the electrolyser efficiency from 60% to 85% results in an even more significant reduction in the MMSP of 21.4%.

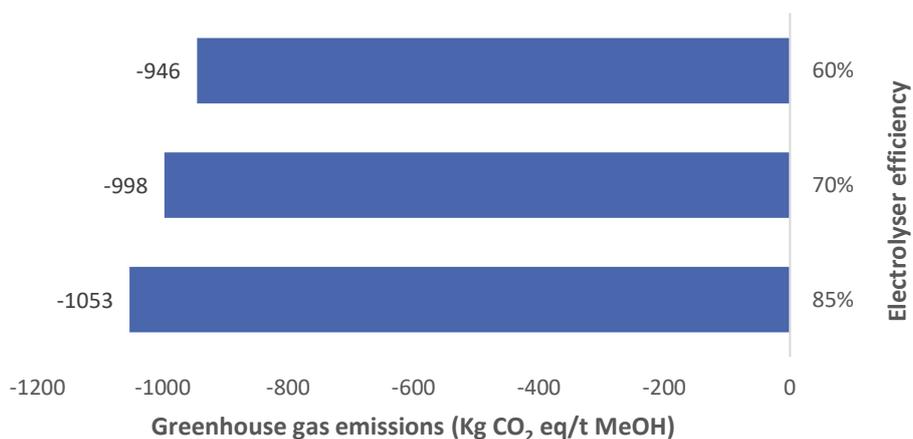


Figure 3. Effect upon the greenhouse gas emissions of the CCU methanol plant as the efficiency of the PEM electrolysis unit increases. Using system expansion with substitution.

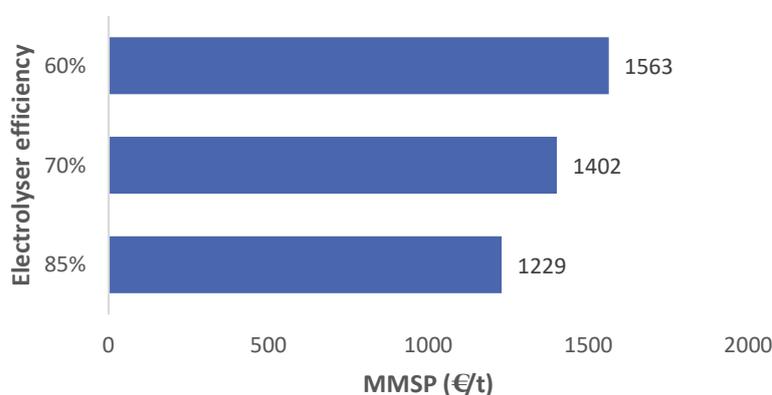


Figure 4. Effect on the MMSP of the CCU methanol plant as the efficiency of the PEM electrolysis unit is increased.

D.3.3 CO₂ capture efficiency/ratio

The small changes in capture efficiency of the membrane capture technology which were modelled resulted in very small changes in the climate change impact of less than 0.5 Kg CO₂ eq/t MeOH. However, the increase in efficiency caused a small increase in GHG emissions (rather than the expected decrease) whereas a decrease in efficiency resulted in a small decrease in emissions, as shown in Figure 5 overleaf. This can be explained by the fact that when using membrane capture technology, increases in capture efficiency (or the capture ratio) are achieved by using more energy. However, it should be noted that the increases modelled are very small, reflecting the fact that only around 2% of the overall GHG emissions of the entire CCU methanol process are due to the CO₂ capture unit.

As depicted in Figure 6, enhancing the CO₂ capture efficiency by elevating the operating pressures of the membranes does not translate into economic improvements and the MMSP is increased slightly (by 0.6%) as the efficiency increases from 85% to 92%.

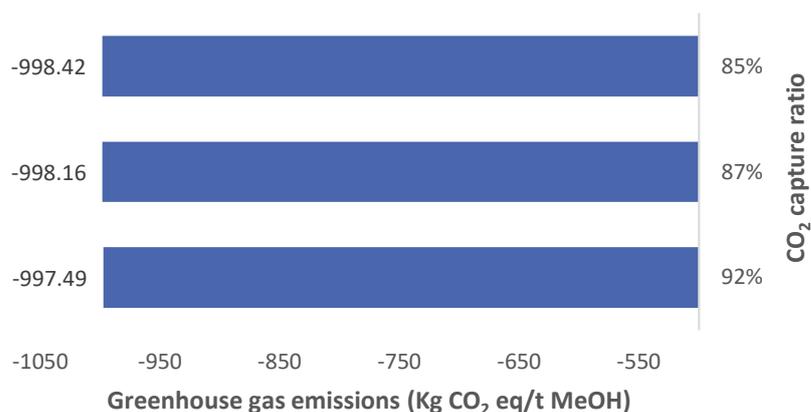


Figure 5. Effect upon the greenhouse gas emissions of a CCU methanol plant as the efficiency of the CO₂ capture unit increases. Using system expansion with substitution.

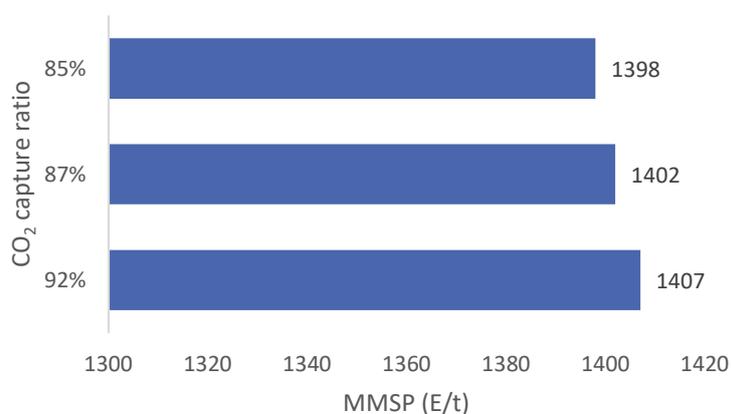


Figure 6. Effect on the MMSP of the CCU methanol plant as the efficiency of the CO₂ capture unit is increased.

As with the GHG emissions, we see that the MMSP increases with increased efficiency of the electrolyser, but decreases with increased efficiency of the membrane capture unit. This reflects the decreased electricity use in the electrolyser but increased electricity use in the capture unit to achieve the higher capture efficiency. Note that only the electricity costs are changing the MMSP, potential changes to the equipment required by the changing efficiencies were not modelled.

D.3.4 Reactor operating pressure

In the third and final sensitivity analysis, the baseline operating pressure of the methanol reactor was 78 bar, so this was varied from 50 to 100 bar to determine the effect of this upon the GHG emissions and the MMSP. Whilst the increased pressure leads to higher methanol yield and so productivity, it also increases the electricity consumption of the synthesis plant. The opposite effect is observed when lower pressures are modelled.

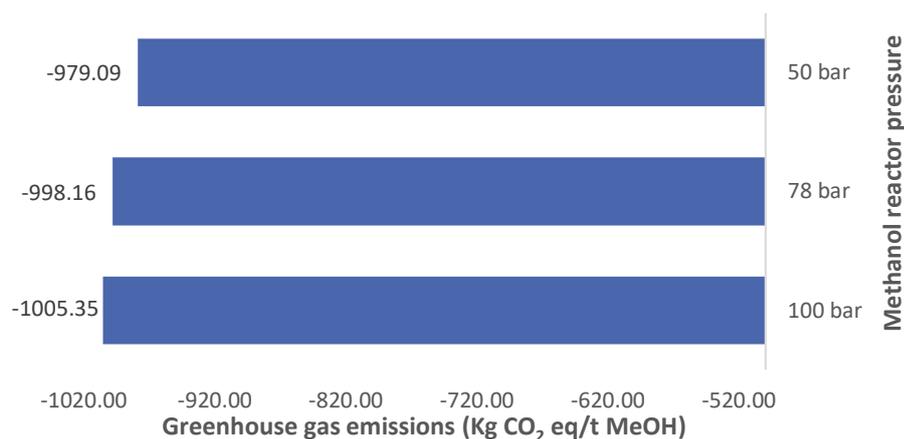


Figure 7. Effect upon the greenhouse gas emissions of a CCU methanol plant as a function of the operating pressure of the methanol reactor. Using system expansion with substitution.

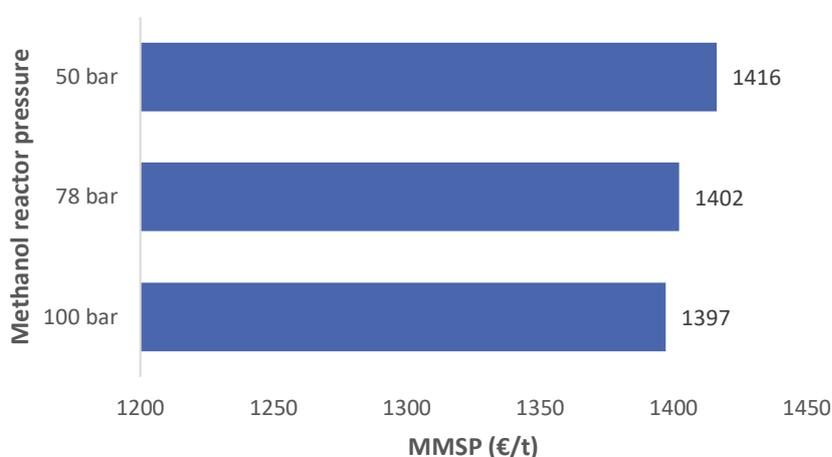


Figure 8. Effect upon the MMSP of a CCU methanol plant as a function of the operating pressure of the methanol reactor.

The data in Figures 7 and 8 reveals that the increased productivity overrides the higher electricity demand, so increasing pressures from 50 to 100 bar decreases GHG emissions by 2.6% and decreases MMSP by 1.3%.

D.3.5 Integrated results

The modelled effects upon GHG emissions and economic data provided above were displayed on separate graphs, but the same data can be shown on a single graph, in the form of a nested box chart. The results combined in this way are presented in Figure 9. The benefit of this form of presentation is that it enables the scale of the sensitivity of both MMSP and GHG emissions to changes in the three variables (electrolyser efficiency, CO₂ capture ratio and methanol reactor pressure) to be compared relative to each other.

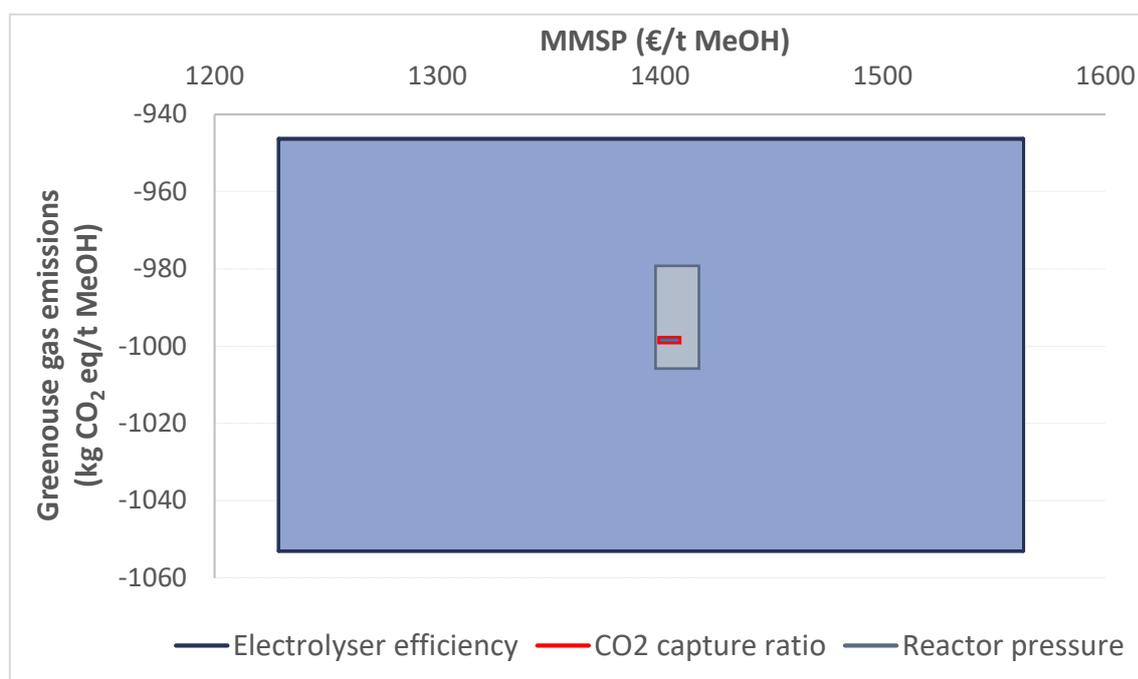


Figure 9. Combined LCA and TEA results considering the effect of the electrolyser efficiency, MeOH reactor pressure and the CO₂ capture ratio

The blue box indicates the sensitivity of the CCU methanol system to the electrolyser efficiency. The range (or scale) in both axes shows that the system is far more sensitive to changes to this variable than to the other two variables – both in terms of impacts upon the economics and GHG emissions.

The yellow box indicates the more limited sensitivity of the system to changes in the methanol reactor operating pressure of the scale modelled, both in terms of MMSP and GHG emissions. In this type of presentation, the shape of the boxes (along with the scales of the axis) is significant as it allows the comparison of the relative importance of a variable to the two indicators. In this case, the results suggest that the MMSP is relatively more sensitive to changes in the electrolyser efficiency than are the GHG emissions. However, the MMSP is relatively less sensitive to changes in reactor pressure than the GHG emissions which show a larger change (as the yellow box is tall and narrow compared to the blue box). The sensitivity of the system to changes in the CO₂ capture ratio (red box) can be considered negligible.

It is important to recognize that the scale of the sensitivity observed may be simply reflecting the scale of the changes modelled, so some caution is required when interpreting such a diagram. However, the shape of the boxes does allow a useful comparison of the sensitivity of the system to the variable relative to the other variables plotted.

D.4 Combined indicator



Reference to General Assessment Principles

Section A.5.6 Calculating combined economic and environmental indicators provides an example of the calculation of CO₂ abatement costs, as an example of a combined indicator of relevance to CCU processes.

A combined indicator that is used widely is the abatement cost of CO₂, C_{abated} (€/t CO₂ abated). Equation 1 shows that different technologies should be compared with each other against a reference system.

$$C_{abated} = \frac{C_{CCU} - C_{ref}}{GHG\ emissions_{ref} - GHG\ emissions_{CCU}} \quad (1)$$

As was used in the separate LCA and TEA assessments, the reference or benchmark technology is conventional methanol production from steam methane reforming. The GHG emissions of SMR is 516.4 kg CO₂ eq./t MeOH (taken from the LCA worked example *Part C*) and the average market price of methanol in 2016 was 248 €/t.

Usually, the abatement cost is calculated to compare different conversion systems. In our case, however, since only one technology was examined, abatement costs were calculated for different electrolyser efficiencies as previously described. As expected, the abatement cost reduces as the efficiency of the electrolyser increases, as displayed in Figure 10. An electrolyser efficiency of 85% therefore corresponds to a higher economic efficiency in respect to emission savings. For all cases, positive costs are observed, due to the difference between the prices of the conventional methanol and the CO₂-derived methanol.

In this case, selling prices instead of production costs were used due to data availability. Nevertheless, the prices include the costs of producing the methanol.

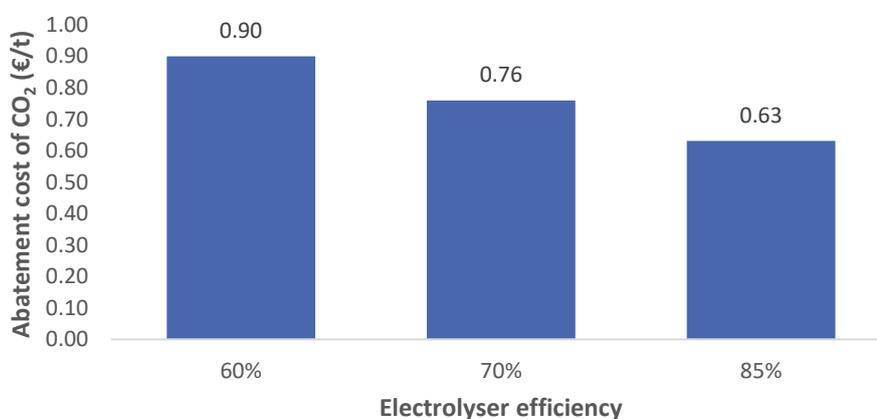


Figure 10. Abatement costs of CO₂ for different electrolyser efficiencies

D.5 Conclusions

In conclusion, the sensitivity analyses described here determined that increasing the efficiency of the electrolysis process and the methanol reactor operating pressure yields positive outcomes from both an environmental and an economic standpoint. However, increasing the CO₂ capture ratio by increasing the operating pressure of the membranes was found to have slightly negative environmental and cost outcomes which can be attributed to the increased pressures that have to be applied in order to enhance the membrane separation efficiency. The consequent increased electricity consumption cannot be offset by the reduced CapEx and reduced emissions of CO₂ due to the reduced volume of flue gas treated.

The depiction of the system sensitivity to the three variables tested using a box chart enabled the relative scale of the resulting impacts to be appreciated. Clearly, changes to the electrolyser efficiency would likely dwarf any changes to the reactor operating pressure or CO₂ capture ratio, both in terms of financial impact (MMSP) and environmental impacts (GHG emissions).

This chapter does not aim to provide a template or a definitive guide to performing an integrated TEA/LCA assessment of a CCU process. Rather, it hopes to give some initial direction and signpost practitioners towards possible routes towards integration of these two forms of assessment.

The approach taken enables the comparison of the CO₂ abatement costs achieved by producing CO₂-derived methanol with other CO₂ abatement technologies. However, the benefits of CCU are not solely CO₂ abatement, but also the ability to move the chemicals and wider process industries away from their current reliance upon fossil-derived carbon feedstock.

As explained in the LCA worked example presented in **Part C**, the examples of environmental impact provided have focused upon the climate change impact category which provides a calculation of the GHG emissions. This was done because this impact category is of direct interest (and is often a driver) to those conducting CCU processes. However, a focus upon the GHG emissions alone risks masking other environmental impacts resulting from a CCU process which would have been identified by analysis of the other environmental impact categories. For this reason, it must be emphasised that when conducting an LCA of CCU process, whether integrated with TEA or not, all impact categories need to be assessed.

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