

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29

MR. WESLEY A BICKFORD (Orcid ID : 0000-0001-7612-1325)

DR. DONALD R. ZAK (Orcid ID : 0000-0002-9730-7337)

Article type : Article

**Root endophytes and invasiveness: no difference between native and non-native *Phragmites* in the Great Lakes Region**

Wesley A. Bickford<sup>1,2,†</sup>, Deborah E. Goldberg<sup>2</sup>, Kurt P. Kowalski<sup>1,3</sup>, Donald R. Zak<sup>2,3</sup>,

<sup>1</sup>U.S. Geological Survey – Great Lakes Science Center, Ann Arbor, Michigan USA

<sup>2</sup>Department of Ecology and Evolutionary Biology, University of Michigan, Ann Arbor, Michigan USA

<sup>3</sup>School for Environment and Sustainability, University of Michigan, Ann Arbor, Michigan USA

Received 25 June 2018; revised 10 September 2018; accepted 11 October 2018. Corresponding

Editor: Karen Haubensak.

† E-mail: [wbickford@usgs.gov](mailto:wbickford@usgs.gov)

**Abstract**

Microbial interactions could play an important role in plant invasions. If invasive plants associate with relatively more mutualists or fewer pathogens than their native counterparts, then microbial communities could foster plant invasiveness. Studies examining the effects of microbes on invasive plants commonly focus on a single microbial group (e.g., bacteria) or measure only plant response to microbes, not documenting the specific taxa associating with invaders. We surveyed root microbial communities associated with co-occurring native and non-native lineages of *Phragmites australis*, across Michigan, USA. Our aim was to determine whether (1) plant lineage was a stronger predictor of root microbial community composition than environmental variables and (2) the non-native, lineage associated with more mutualistic and/or fewer pathogenic microbes than the native lineage. We used microscopy and culture-independent

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1002/ecs2.2526](https://doi.org/10.1002/ecs2.2526)

This article is protected by copyright. All rights reserved

1 molecular methods to examine fungal colonization rate and community composition in three  
2 major microbial groups (bacteria, fungi, and oomycetes) within roots. We also used microbial  
3 functional databases to assess putative functions of the observed microbial taxa. While fungal  
4 colonization of roots was significantly higher in non-native *Phragmites* than the native lineage,  
5 we found no differences in root microbial community composition or potential function between  
6 the two *Phragmites* lineages. Community composition did differ significantly by site, with soil  
7 saturation playing a significant role in structuring communities in all three microbial groups. The  
8 relative abundance of some specific bacterial taxa did differ between *Phragmites* lineages at the  
9 phylum and genus level (e.g., *Proteobacteria*, *Firmicutes*). Purported function of root fungi and  
10 respiratory mode of root bacteria also did not differ between native and non-native *Phragmites*.  
11 We found no evidence that native and non-native *Phragmites* harbored distinct root microbial  
12 communities; nor did those communities differ functionally. Therefore, if the trends revealed at  
13 our sites are widespread, it is unlikely that total root microbial communities are driving invasion  
14 by non-native *Phragmites* plants.

15 **Key words:** bacteria; endophytes; fungi; invasive plants; plant-microbial interactions; microbes;  
16 mutualists; oomycetes; pathogens; roots; soil saturation

## 18 **Introduction**

19 The intimate association of plants and their microbiome has significant impacts on plant  
20 performance and thereby may be an important driver of invasion success. Recent studies have  
21 focused on the plant-associated microbiome of invasive plants, because it could provide a  
22 mechanism for their colonization and recruitment, as well as for decreased performance of native  
23 species (Thorpe and Callaway 2006, Salles and Mallon 2014, Kowalski et al. 2015). More  
24 specifically, if invasive plants cultivate a microbial community with stronger mutualistic effects  
25 or less intense pathogenic effects than native plants, then the net effect of their microbiome will  
26 be more positive than for native plants (Reinhart and Callaway 2006).

27 Evidence linking belowground microbial communities to invasive plant success is mixed.  
28 For instance, invasive grasses can have stronger positive or less negative plant-soil feedbacks  
29 than natives in monoculture (Klironomos 2002) and in mixtures (Kulmatiski 2018). However, a  
30 meta-analysis covering many different habitat types from forest to wetland, did not find a  
31 consistent trend supporting feedbacks as a mechanism for invasion success (Suding et al. 2013),

1 suggesting that belowground microbial community effects on plant performance may be taxon-  
2 or habitat-specific. In addition, plant-soil feedback experiments typically do not identify which  
3 microbes are cultivated such that specific microbial mechanisms are unclear. Moreover,  
4 microorganisms are phylogenetically diverse and can affect plant performance in many ways,  
5 ranging from direct pathogenesis to alteration of site nutrient availability. However, studies that  
6 examine plant-associated microbes often target one specific group of taxa (e.g., bacteria or  
7 fungi), thereby ignoring the potential interactions among these diverse organisms that could have  
8 important outcomes for plant performance.

9 One of the most noxious and widespread wetland plant invaders is the common reed,  
10 *Phragmites australis* (Cav.) Trin. ex Steud. (Meyerson et al. 2016). This clonal plant occurs  
11 widely throughout North American wetlands and includes both native (hereafter native  
12 *Phragmites*) and exotic, invasive haplotypes (hereafter non-native *Phragmites*; Saltonstall 2002,  
13 Saltonstall et al. 2004). In many Atlantic Coast and Great Lakes wetlands, non-native  
14 *Phragmites* is much more productive, and in some cases, displaces native *Phragmites* (Martin  
15 and Blossey 2013, Mozdzer et al. 2013, Price et al. 2013). Due to their close phylogenetic  
16 relationship, comparison between the non-native and native lineages of *Phragmites* offers a  
17 unique opportunity to understand how microbial communities may impact invasive success.

18 As a species considered invasive in most of North America, *Phragmites* has been well  
19 studied, including some exploration of select microbial groups. Previous studies have examined  
20 leaf, root, and seed fungal endophytes of non-native *Phragmites* (Clay et al. 2016, Soares et al.  
21 2016, Shearin et al. 2018), characterized and compared soil oomycete communities (Nelson and  
22 Karp 2013, Crocker et al. 2015, Cerri et al. 2017), and compared soil bacterial (Bowen et al.  
23 2017) and archaeal (Yarwood et al. 2016) communities between *Phragmites* lineages. However,  
24 the functional significance of *Phragmites*-associated microbiomes remains unknown. To date, no  
25 comprehensive survey of root microbial communities and their interactions with native and non-  
26 native plant lineages has been performed. A comparison of the microbial communities associated  
27 with the two lineages may lend mechanistic insights into the invasiveness of non-native  
28 *Phragmites*. For example, associating with relatively more mutualists or fewer pathogens, would  
29 give the non-native lineage a performance advantage over the native lineage.

30 Here, we report on an intensive field survey of bacterial, fungal, and oomycete  
31 communities associated with the roots of native and non-native *Phragmites* in sites where both

1 lineages co-occur throughout the state of Michigan, USA. We focus on roots as the main  
2 interface between the plant and the soil environment and where nutritional mutualisms as well as  
3 pathogen attack are common. If performance differences between native and non-native plant  
4 lineages are driven by their root microbial communities, then we hypothesize that plant lineage  
5 would be a stronger predictor of microbial community differences than environmental variables.  
6 Furthermore, mutualist microbes should be more abundant and/or pathogen microbes should be  
7 less abundant in the non-native lineage.

## 8 **Materials and Methods**

### 9 *Study Sites*

10 In August 2015, we sampled roots from native and non-native *Phragmites* individuals at  
11 8 sites distributed across Michigan, USA (Appendix S1: Table S1). We sought sites that had at  
12 least 3 distinct patches of native and non-native *Phragmites* in close proximity to one another  
13 growing under similar environmental conditions (i.e., soil type, hydrology) with no recent history  
14 of invasive plant management (e.g., herbicide, burning). Due to the rarity of co-occurring native  
15 and non-native *Phragmites* populations that met these criteria (non-native is rare and well-  
16 managed in northern Michigan; native is rare in southern Michigan), patch size varied  
17 considerably among sites (1–100 m<sup>2</sup>) and micro-environmental differences among patches within  
18 sites and sample size differences among sites were unavoidable. To account for this variation, we  
19 used environmental characteristics as co-variables in our analyses.

20 At each site, we morphologically identified all *Phragmites* patches as native or non-  
21 native. Size of *Phragmites* clones can be quite variable, wherein patches can consist of a single  
22 clone or multiple clones. While clones can differ morphologically (Křiváčková-Suchá et al.  
23 2007), such differences are not consistent or distinct enough to reliably use for field  
24 identification. To maximize probability that patches represented distinct clones, we considered  
25 patches that were separated by at least 10 m to be separate clones. We classified the degree of  
26 soil saturation as either unsaturated, saturated, or saturated with surface water, recorded height of  
27 water table (if over the surface), and nature of surrounding vegetation. Within each patch, a  
28 ramet near the center of the clone was randomly selected. Roots were collected by cutting with a  
29 serrated knife in a 10 cm diameter circle around the chosen ramet, exhuming subtending roots,  
30 and then placing them in a plastic bag. Leaf samples from the same stem were collected for

1 nutrient analysis. Soil samples (5 cm depth) were collected from the area adjacent to the selected  
2 ramet. All samples were kept on ice until returned from the field.

### 3 *Sample Preparation*

4 Roots were washed with deionized water to remove all soil particles and litter and placed  
5 into a sample bag. A subset of the fine roots (< 1 mm diameter) was removed and placed into  
6 biopsy cassettes for microscopic analysis, and the remaining roots were stored for molecular  
7 analysis. Biopsy cassettes containing fine roots were submerged in 10% KOH at 20°C for 24–48  
8 h or until roots appeared clear, and then submerged in a mixture of 5% black ink (Sheaffer Pen  
9 and Art Supply, Providence, Rhode Island) in 5% acetic acid at 95°C for 2 mins (modified from  
10 Kosuta et al. 2005). Cassettes were then immediately rinsed with deionized water and a few  
11 drops of acetic acid to remove excess stain. Cassettes containing stained roots were stored at 4°C  
12 until microscopic analysis. Stained roots were cut into 1-cm sections, mounted on microscope  
13 slides and examined at 40× magnification for the presence of fungal hyphae following the line-  
14 intersection method (McGonigle et al. 1990). While we looked for distinct fungal structures, no  
15 vesicles or arbuscules were found. Therefore, counts were restricted to presence/absence of  
16 hyphae only.

17 Fine roots used for molecular analysis were surface sterilized by submerging in 70%  
18 ethanol for 1 min, 0.5% sodium hypochlorite for 3 mins, and 95% ethanol for 30 secs. A subset  
19 of sterilized roots was pressed onto petri plates containing Malt Extract Agar for 30 secs and  
20 incubated at room temperature for 24 h. No growth on the agar was used as an indicator of  
21 successful surface sterilization. After surface sterilization, roots were placed into a sterile whirl  
22 pack bag and lyophilized in preparation for DNA extraction.

23 Soils were passed through a 2-mm sieve and oven dried at 60°C for 48 h. Dried samples  
24 were ground with a mortar and pestle and subsamples from each (0.5 g) were processed in  
25 duplicate in a Leco CNS2000 Analyzer (LECO, St. Joseph, Michigan, USA) to measure carbon  
26 and nitrogen. Soil phosphorus was determined colorimetrically following the Bray P1 extraction  
27 method (Bray and Kurtz 1945).

### 28 *Molecular analysis*

29 To prepare for DNA extractions, approximately 50 mg of freeze-dried fine roots was  
30 ground in a mortar and pestle with approximately 250 mg autoclaved sand and approximately 1  
31 mL liquid N. The finely ground root sample was then further homogenized with the lysis buffer

1 from Macherey-Nagel Nucleospin Plant II DNA extraction kits (Macherey-Nagel, Bethlehem,  
2 Pennsylvania, USA). All genomic DNA extracts were verified by electrophoresis. Extracts were  
3 checked for quality on a NanoDrop UV/Vis spectrophotometer and concentration using a  
4 Quant-iT PicoGreen dsDNA kit (Invitrogen, Carlsbad, Carlsbad, USA).

5 All polymerase chain reactions (PCR) for each taxon of interest (fungi, bacteria,  
6 oomycetes) were performed using subsamples of the same template genomic DNA. Genomic  
7 DNA was diluted to ensure equimolar concentration of template DNA in each PCR reaction. For  
8 fungal amplification, ITS1F and ITS4 forward and reverse primers were used (White et al.  
9 1990). For bacterial amplification, the 16s region was amplified using primers 27F and 519R  
10 (Lane 1991). Oomycete DNA was amplified using the Oom1F and Oom1R (Arcate et al. 2006).  
11 See Appendix S1: Table S2 for all PCR conditions, primer sequences, and master mixes. PCR  
12 products were purified using the Qiagen MinElute PCR kit and quantified using a Quant-iT  
13 PicoGreen dsDNA kit (Invitrogen, Carlsbad, California, USA).

14 Sequencing was performed on a PacBio-RS II system utilizing circular consensus  
15 technology, which can generate 99.5–99.9% sequence accuracy for DNA fragments ranging  
16 from 150 to 500 bp (Travers et al. 2010). Seventeen barcoded samples, pooled in equimolar  
17 concentration, were multiplexed per SMRT chip. Three SMRT chips were sequenced per  
18 microbial group (9 total SMRT chips) at the University of Michigan Sequencing Core.

#### 19 *Bioinformatics analysis*

20 Raw sequence data were processed using mothur v1.39.5 (Schloss et al. 2009).  
21 Operational taxonomic units (OTUs) were clustered at 97% for bacterial sequences. Fungal and  
22 oomycete sequences were binned into phylotypes because ITS does not allow for sequence  
23 alignment over large datasets and mothur requires a sequence alignment to cluster OTUs.  
24 Bacterial taxonomy was determined by comparing representative sequences to the taxa found in  
25 the SILVA database (Quast et al. 2018). Fungal taxonomy was assigned based on the UNITE  
26 database (Nilsson et al. 2013). For oomycete taxonomy, we created a custom oomycete-specific  
27 database from NCBI records. Sequences were compared to this database using the blastn toolkit.  
28 Each microbial group was rarefied according to the sample that yielded the fewest number of  
29 sequences to ensure equal sampling across all samples. Fungi were rarefied to 200 sequences per  
30 sample, bacteria to 459 sequences, and oomycetes to 468 sequences. Sequencing coverage of  
31 fungal and bacterial communities was estimated using Good's coverage estimator (Good 1953).

1 Although rarefaction was carried out at a relatively low number of sequences, Good's coverage  
2 indicated strong sampling effort at each sample (Appendix S1: Fig. S1–S3). Sequences were  
3 uploaded to the NCBI Sequence Read Archive under SRA accession number SRP160913.

#### 4 *Functional Assignment*

5 To evaluate functional potential of fungal OTUs, we used FUNGuild (Nguyen et al.  
6 2015) which parses fungal communities by trophic mode and functional guilds. We analyzed  
7 outputs at the trophic mode level to understand the proportion of the root communities composed  
8 of mutualists (symbiotrophs), pathogens (pathotrophs), and likely commensalists (saprotrophs).  
9 For bacterial functional potential, we used the BugBase (Ward et al. 2017) tool that groups  
10 organisms into functional groups based on KEGG pathways (Ogata et al. 1999) compiled by  
11 PICRUST (Langille et al. 2013). This tool allowed us to view bacterial communities by their  
12 oxygen requirements and potential for stress tolerance. Use of this tool required that we  
13 reclassify our OTUs using the greengenes taxonomic database (Desantis et al. 2006).

#### 14 *Data analysis*

15 We used Principle Component Analysis (PCA) to visualize environmental gradients at  
16 our sample sites and select uncorrelated variables to use in subsequent analysis. We used a two-  
17 phased analysis to account for the fact that environmental variables were not balanced with  
18 respect to site and lineage. First, we tested differences by site and lineage for root colonization,  
19 microbial alpha diversity, and relative sequence abundance using 2-way ANOVA (Type III sum  
20 of squares). Second, we ignored site and assessed the impact of environmental variables (soil  
21 nutrients and saturation) on the response variables, including potential interactions with lineage  
22 using Analysis of Co-Variance (ANCOVA). Permutational multivariate analysis of variance  
23 (PerMANOVA) tested whether plant lineage or site predicted significant microbial community  
24 differences among our samples. Homogeneity of dispersions (PermDISP) tested whether  
25 microbial community samples differed in their degree of dispersion from their centroid.  
26 Additionally, we used Principle Coordinate Analysis (PCoA) with fitting of environmental data  
27 to visualize microbial community differences between lineages and potential environmental  
28 drivers. All statistical analyses were run in the R environment (R Core Team 2016) with the  
29 exception of PerMANOVA and PermDISP, which were conducted in PRIMER-E with  
30 PerMANOVA+ (Clarke and Gorley 2006). All R code, notes, and associated data can be  
31 accessed on GitHub at [https://github.com/wesbick/Root\\_paper](https://github.com/wesbick/Root_paper).

## 1 **Results**

### 2 *Environmental Characteristics*

3 A PCA of our environmental variables revealed that soil saturation, soil nitrogen, and soil  
4 phosphorus were uncorrelated. Soil carbon and nitrogen were strongly correlated, so soil carbon  
5 is not reported (Appendix S1: Fig. S4). Soil nitrogen and phosphorus did not differ significantly  
6 by lineage, but patch saturation level was non-random with respect to lineage at our sites ( $X^2 =$   
7 11.99,  $P = 0.005$ ), wherein patches of non-native *Phragmites* were more likely to be unsaturated  
8 and patches of native were more likely to be saturated. Both were equally likely to occur in high  
9 water (Appendix S1: Table S3), suggesting that differences in growth habitat between lineages at  
10 our sites do not represent differential water tolerances, but instead is an artifact of our sampling  
11 locations.

### 12 *Fungal Colonization*

13 Evidence from microscopy revealed that fungal root colonization was significantly  
14 greater in non-native *Phragmites* (Fig. 1a; ANOVA,  $F = 24.57$ ,  $P < 0.001$ ). Fungal colonization  
15 also differed significantly among sites (ANOVA,  $F = 6.52$ ,  $P < 0.001$ ), but site and plant lineage  
16 did not interact (Appendix S1: Table S4). Of all environmental variables measured, only soil  
17 saturation was a significant predictor of fungal root colonization (ANCOVA,  $F = 23.47$ ,  $P <$   
18  $0.001$ , Appendix S1: Table S5). In fact, the magnitude of the differences between root  
19 colonization between native and non-native *Phragmites* depended on the degree of soil saturation  
20 (Fig. 1b), although the interaction between lineage and saturation was not significant (Appendix  
21 S1: Table S5, Fig. S5). Fungal root colonization was most different between native and non-  
22 native lineages in unsaturated conditions and was not different when soil was saturated or  
23 submerged. We observed no structures characteristic of arbuscular mycorrhizal fungi.

### 24 *Alpha Diversity*

25 As one measure of microbial community differences, we analyzed differences in alpha  
26 diversity among sites and between lineages. Bacterial and fungal alpha diversity (richness,  
27 Shannon-Weiner diversity) differed among sites but not by plant lineage (Appendix S1: Table  
28 S6) or with any of the measured environmental variable (Appendix S1: Table S7, Fig. S6).  
29 Oomycete community richness differed modestly between the two lineages, with the non-native  
30 roots hosting a greater number of species and higher diversity than the native roots (Fig. 2;



1 ANOVA,  $F = 4.22$ ,  $P = 0.048$ , Appendix S1: Table S8). Environmental variables did not affect  
2 oomycete alpha diversity (Appendix S1: Table S9, Fig. S7).

### 3 *Community Composition*

4 Host lineage was not a significant predictor of community composition for any microbial  
5 group based on PerMANOVA (Table 1). However, site was a significant predictor of variation  
6 for all three microbial groups and significantly interacted with lineage to shape bacterial  
7 communities (Table 1). PermDISP revealed a marginally significant difference in the dispersion  
8 patterns among sites in bacterial (pseudo- $F = 5.76$ ,  $P = 0.072$ ) and fungal (pseudo- $F = 4.56$ ,  $P =$   
9  $0.064$ ) communities, which indicates that the difference in community detected by  
10 PerMANOVA may be due to heterogeneity of dispersion around the centroids of the sites. Given  
11 the importance of site for community composition, we explored which environmental variables  
12 could account for this spatial variation with PCoA. While soil nitrogen marginally explained  
13 community variation in two of the three microbial groups, degree of saturation more consistently  
14 explained substantial variation in ordination space for all three groups (Fig. 3, Table 2).

### 15 *Taxonomic Analyses*

16 Despite the lack of difference in overall community composition according to the  
17 multivariate analyses, relative abundance of some bacterial phyla on roots did differ by plant  
18 lineage (Fig. 4, Appendix S1: Table S6). Relative sequence abundance of the most common  
19 bacterial phyla was associated with plant host lineage and not site (Fig. 4). While the majority of  
20 all bacterial sequences recovered in both lineages were Proteobacteria, non-native *Phragmites*  
21 roots host relatively more Proteobacteria, but fewer Firmicutes and Bacteroidetes taxa. The  
22 Bacteroidetes were also influenced by site (Appendix S1: Table S6). Major bacterial genera are  
23 listed in Appendix S1: Tables S6–S7 and Fig. S8)

24 Relative sequence abundance of fungi did not differ by site or host lineage at the phylum  
25 level (ANOVA, Appendix S1: Table S4). Over 90% of recovered fungal sequences were from  
26 Ascomycota. Minor phyla included Basidiomycota (5.7%), Zygomycota (0.6%), Glomeromycota  
27 (0.2%), and unclassified fungi (2.7%; Appendix S1: Fig. S9). We expected that Ascomycetes  
28 would make up the majority of sequences as most root endophytes are from this phylum. The  
29 extremely low relative abundance of Glomeromycota, the dominant phylum of AMF, was  
30 consistent with the lack of AMF structures found via microscopy. Relative abundances of  
31 recovered genera differed significantly by site, but not by lineage (Appendix S1: Table S4). The

1 dominant genera recovered in roots were *Gibberella* (19.5%), *Tetracladium* (13.4%),  
2 *Microdochium* (11.0%) and *Stagonospora* (7.6%; Appendix S1: Fig. S10). Environmental  
3 variables did not affect relative sequence abundance at the phylum level, but saturation level was  
4 a strong determinant of relative abundance in all dominant fungal genera (Appendix S1: Table  
5 S5). The majority of our recovered oomycete sequences matched most closely to unclassified  
6 uncultured oomycete strains (59.4%). The dominant classified genus was *Pythium* which made  
7 up 40% of the sequences recovered. *Phytophthora*, *Saprolegnia*, and *Lagena* were recovered in  
8 very low abundance. Oomycete genera did not differ in relative abundance by site or lineage  
9 (Appendix S1: Table S8, Fig. S11), or saturation level (Appendix S1: Table S9, Fig. S7).

#### 10 *Functional classification*

11 Because soil saturation was a strong driver of bacterial community differences, we  
12 focused on respiratory mode as a trait that could differentiate groups based on saturation.  
13 Additionally, differences in respiratory mode could drive rates of nutrient cycling and has the  
14 potential to feedback to plant productivity. We used BugBase to make respiratory mode  
15 determinations based on GreenGenes taxonomy of our 16S sequences. While most OTUs were  
16 from aerobic bacteria (Fig. 5), the relative abundance of anaerobes was marginally determined  
17 by plant lineage, with the native lineage hosting more anaerobes than the non-native. Soil  
18 saturation was also a predictor of relative abundance of facultative anaerobes (ANCOVA,  $F =$   
19  $3.34$ ,  $P = 0.077$ , Appendix S1: Table S7). However, plant lineage and soil saturation did not  
20 significantly interact in determining relative abundance of bacteria based on respiratory mode,  
21 suggesting that plants hosted similar communities with respect to oxygen requirements in all  
22 saturation levels (Appendix S1: Table S7, Fig. S6).

23 Using functional assessment from FUNguild, we compared the relative sequence  
24 abundance of the dominant trophic modes of fungi inhabiting *Phragmites* roots (Fig. 6). Most  
25 taxa were categorized as pathotrophs or saprotrophs, with no difference in the relative abundance  
26 of either group between native and non-native lineages (Appendix S1: Table S4). Symbiotrophs  
27 (mutualists) and pathotroph-symbiotrophs (a hybrid group consisting mostly of Dark Septate  
28 Endophytes) also did not differ by site or lineage. Pathotrophs (ANCOVA,  $F = 9.13$ ,  $P = 0.005$ )  
29 and Saprotrophs (ANCOVA,  $F = 14.63$ ,  $P < 0.001$ ) did differ based on soil saturation (Appendix  
30 S1: Table S5, Fig. S5), perhaps reflecting the oxygen demands of these functional groups under

1 various saturation levels. In the fungal communities, there was no evidence for higher mutualist  
2 or lower pathogen load on the non-native *Phragmites*.

### 3 **Discussion**

4 If performance differences between native and non-native plant lineages (e.g., growth  
5 rate, productivity, etc.) are primarily driven by their root microbial communities, then we would  
6 expect that (1) bacterial, fungal, and oomycete communities will be differentially assembled,  
7 comprised of different taxa abundances, OTU abundances, or colonization rates in native and  
8 non-native *Phragmites* roots and (2) non-native *Phragmites* roots will associate with relatively  
9 fewer microbial taxa capable of pathogenesis and relatively more microbial taxa capable of  
10 mutualistic interactions than native *Phragmites* roots. Despite a thorough examination of the root  
11 microbial community encompassing three of the major microbial groups (bacteria, fungi, and  
12 oomycetes), we could find no evidence to support the hypothesis that performance differences  
13 between native and non-native *Phragmites* are driven by their root microbial communities alone.  
14 With the exception of relative abundance of a few bacterial genera and the extent of fungal  
15 colonization, plant lineage did not predict root microbial community structure or composition in  
16 our study. Instead, site, and specifically soil saturation level, was a stronger predictor of the  
17 identity of microorganisms that endophytically colonized roots. Additionally, there was no  
18 evidence for functional differences in the root microbial communities that could explain  
19 performance advantages observed in the non-native lineage. Consequently, it is unlikely that the  
20 invasive capacity of non-native *Phragmites* arises due to a greater abundance of mutualists or a  
21 lower abundance of pathogens in roots.

22 For bacterial root endophytes, soil saturation level was a strong predictor of community  
23 composition wherein we observed different bacterial communities under saturated and  
24 unsaturated conditions. However, soil saturation only modestly affected the relative abundance  
25 of bacterial groups by respiratory requirement or taxonomy. *Pseudomonas*, the most common  
26 genus recovered, were slightly less abundant in high water than in unsaturated patches  
27 (Appendix S1: Fig. S6). Most of the sequences recovered were from aerobic taxa, and the  
28 abundance of aerobes did not differ with saturation. Facultative anaerobes were affected by soil  
29 saturation level, but they made up a much smaller proportion of bacterial sequences (Appendix  
30 S1: Table S7, Fig. S6). Given that ventilation efficiency differs dramatically between native and  
31 non-native *Phragmites* (Tulbure et al. 2012), one might expect to see differences in microbial

1 communities based on respiratory mode. However, the discrepancies in ventilation did not  
2 appear to affect endophytic colonization. It may be that, while rhizosphere oxygen  
3 concentrations likely depend on ventilation and differ in the two lineages, internal root tissues  
4 can remain sufficiently aerobic due to the existence of aerenchyma, despite the differences in  
5 surrounding soil. Bacterial endophytes mainly colonize roots from the rhizosphere (Hardoim et  
6 al. 2008) or plant litter (Ryan et al. 2008). Thus, while there may be differences in the  
7 rhizosphere bacterial communities between the two lineages (Bowen et al. 2017) and those  
8 differences may be in part influenced by ventilation, the broadly habitable, aerobic internal root  
9 may select for aerobes among the community of bacteria in the rhizosphere.

10 For fungal endophytes, soil saturation was also important, influencing colonization rates,  
11 community composition, alpha diversity, as well as the relative abundance of many common  
12 genera (Appendix S1: Table S5). As with bacteria, the soil environment may act as a barrier for  
13 colonization. High water levels creating anaerobic soil conditions would make it difficult for  
14 germinating spores of AMF or other aerobic filamentous fungi to reach the root surface. While  
15 fungal colonization was higher in non-native *Phragmites* roots overall, colonization rate did  
16 differ by soil saturation with the largest differences found in unsaturated conditions (Fig. 1).  
17 Fungal root colonization in native and non-native *Phragmites* lineages has been investigated  
18 previously (Holdredge et al. 2010) with no consistent or statistical difference appearing between  
19 lineages. That study was conducted in a coastal tidal marsh that likely experiences a high degree  
20 of saturation, which may have affected the colonization rates similarly to those in our study. No  
21 arbuscules or vesicles were identified in any samples and the relative abundance of  
22 Glomeromycota taxa in our samples was extremely low. Although we note the deficiencies of  
23 ITS primers in detecting Glomeromycota (Stockinger et al. 2010), given the low levels of AMF  
24 sequences recovered in our samples and the results of the FUNGuild analysis, the fungal hyphae  
25 observed in our microscopy analysis were likely endophytes or root pathogens rather than AMF  
26 mutualists.

27 The only consistent microbial community difference between native and non-native  
28 *Phragmites* was found in oomycete diversity, with the non-native *Phragmites* roots hosting more  
29 diverse oomycete pathogens. While pathogen diversity is not the same as pathogen load, we  
30 might expect that the trend would be in the opposite direction to agree with classical enemy  
31 release dynamics. Additionally, we did not observe any difference in overall communities of

1 oomycetes between plant lineages that would suggest native and non-native *Phragmites* roots are  
2 differentially colonized by oomycete pathogens. Nelson and Karp (2013) compared oomycete  
3 communities between native and non-native *Phragmites* and found no overall difference in  
4 oomycete richness between plant lineages but discovered higher richness in the native lineage at  
5 two of their four sites. Although our results are thus inconsistent with previous findings (Nelson  
6 and Karp 2013), they are consistent with the suggestion by Allen et al. (2018) that non-native  
7 *Phragmites* attracts more generalist soil pathogens than the native *Phragmites*. Importantly,  
8 richness reveals very little about pathogen load or pressure and it is still possible that the two  
9 lineages are experiencing different levels of pathogen pressure given differences in pathogen  
10 virulence on each plant lineage or surrounding plants (Crocker et al. 2015, Allen et al. 2018).

11 Functional potential of endophytic bacteria is difficult to determine (Hardoim et al.  
12 2008). While others have found specific endophytic bacteria that provide stress tolerance to  
13 invasive *Phragmites* (White et al. 2018), functional determinations often come from culture  
14 studies that can measure plant growth response to individual inoculations. Without using an  
15 experimental approach, functional potential can be approximated by looking at specific genes  
16 (Hardoim et al. 2008), but that requires mapped genomes of OTUs found in the study, which can  
17 be difficult to obtain. We instead relied on community and respiratory mode differences between  
18 lineages to infer functional differences. We found no such differences between the native and  
19 non-native lineages within the roots.

20 While our study found no differences in root microbial communities of native and non-  
21 native *Phragmites*, the complete role of microbes in *Phragmites*' invasiveness is likely much  
22 more complex. For example, important microbial community differences driving plant  
23 performance differences could exist elsewhere in the phyllosphere or in the rhizosphere. In fact,  
24 Bowen et al. (2017) found strong rhizosphere bacteria differences between *Phragmites* lineages,  
25 and rhizosphere fungal and oomycete community differences have not yet been studied on a  
26 large scale. Additionally, the plant response to microbes could be lineage specific. We know that  
27 some oomycetes have lineage-specific virulence in *Phragmites* (Crocker et al. 2015), and that the  
28 roles of many endophytic fungi and bacteria are plant and genotype specific (Schulz and Boyle  
29 2005, Hardoim et al. 2008). Therefore, plant responses to microbial communities as well as  
30 community differences are important to definitively understand if microbial communities are  
31 driving invasiveness in *Phragmites*. Finally, we recognize the limitations of this study in that it

1 relies heavily on publicly available databases of microbial taxonomy (SILVA, NCBI, UNITE)  
2 and function (FUNguild, BugBase). While these databases are improving, analyses derived from  
3 them are not definitive, but should provide useful trends in functional potential (Nguyen et al.  
4 2015).

## 5 **Conclusions**

6 We found little evidence of differences in the *Phragmites* root microbiome between  
7 lineages at our sites. There was also no evidence that invasive *Phragmites* associated with  
8 relatively more mutualist or relatively fewer pathogen microbes than the native *Phragmites*.  
9 Thus, it is unlikely that root microbial communities are driving *Phragmites* invasions at these  
10 sites. Future studies should explore microbial communities in the rhizosphere or elsewhere in the  
11 phyllosphere, as well as differential response to microbes in native and non-native *Phragmites* as  
12 potential mechanisms for invasiveness.

## 13 **Acknowledgments**

14 Financial support was provided by the U.S. Geological Survey, the Great Lakes  
15 Restoration Initiative, and the University of Michigan's 2014-2106 EEB Block Grant Awards.  
16 We thank the landowners and organizations who allowed collections on their properties:  
17 Michigan Nature Association, Michigan State Parks, Three Shores CISMA, and D.E. Wee. We  
18 thank R. Upchurch, M. Dworz, G. Hsu, A. Pangallo, M. Nightengale, S. Salley, J. Vance, and K.  
19 Romanowicz for lab assistance. Thanks to T. James and I. Ibanez for providing helpful feedback  
20 and reviews. Thanks also to M. Byappanahalli for providing an additional review of this  
21 manuscript. Any use of trade, firm, or product names is for descriptive purposes only and does  
22 not imply endorsement by the U. S. Government.

23

## 24 **Literature Cited**

- 25 Allen, W. J., L. A. Meyerson, A. J. Flick, and J. T. Cronin. 2018. Intraspecific variation in  
26 indirect plant-soil feedbacks influences a wetland plant invasion. *Ecology*.  
27 DOI:10.1002/ecy.2344.
- 28 Arcate, J. M., M. A. Karp, and E. B. Nelson. 2006. Diversity of Peronosporomycete (Oomycete)  
29 Communities Associated with the Rhizosphere of Different Plant Species. *Microbial*  
30 *Ecology* 51:36–50.
- 31 Bowen, J. L., P. J. Kearns, J. E. K. Byrnes, S. Wigginton, W. J. Allen, M. Greenwood, K. Tran,

- 1 J. Yu, J. T. Cronin, and L. A. Meyerson. 2017. Lineage overwhelms environmental  
2 conditions in determining rhizosphere bacterial community structure in a cosmopolitan  
3 invasive plant. *Nature Communications* 8. DOI:10.1038/s41467-017-00626-0.
- 4 Bray, R. A., and L. T. Kurtz. 1945. Determination of total, organic and available forms of  
5 phosphate in soils. *Soil Science* 59:39–45.
- 6 Cerri, M., R. Sapkota, A. Coppi, V. Ferri, B. Foggi, D. Gigante, L. Lastrucci, R. Selvaggi, R.  
7 Venanzoni, M. Nicolaisen, F. Ferranti, and L. Reale. 2017. Oomycete Communities  
8 Associated with Reed Die-Back Syndrome. *Frontiers in Plant Science* 8:1–11.
- 9 Clarke, K. R., and R. N. Gorley. 2006. PRIMER v6: User Manual/Tutorial. PRIMER-E.  
10 Plymouth.
- 11 Clay, K., Z. R. Shearin, K. A. Bourke, W. A. Bickford, and K. P. Kowalski. 2016. Diversity of  
12 fungal endophytes in non-native *Phragmites australis* in the Great Lakes. *Biological*  
13 *Invasions* 18:2703–2716.
- 14 Crocker, E. V., M. A. Karp, and E. B. Nelson. 2015. Virulence of oomycete pathogens from  
15 *Phragmites australis* -invaded and noninvaded soils to seedlings of wetland plant species.  
16 *Ecology and Evolution* 5:2127–2139.
- 17 Desantis, T. Z., P. Hugenholtz, N. Larsen, M. Rojas, E. L. Brodie, K. Keller, T. Huber, D.  
18 Dalevi, P. Hu, and G. L. Andersen. 2006. Greengenes, a Chimera-Checked 16S rRNA Gene  
19 Database and Workbench Compatible with ARB. *Applied Environmental Microbiology*  
20 72:5069–5072.
- 21 Good, I. J. 1953. The Population Frequencies of Species and the Estimation of Population  
22 Parameters. *Biometrika* 40:237–264.
- 23 Hardoim, P. R., L. S. Van Overbeek, and J. D. Van Elsas. 2008. Properties of bacterial  
24 endophytes and their proposed role in plant growth. *Trends in Microbiology* 16:463–471.
- 25 Holdredge, C., M. D. Bertness, E. Von Wettberg, and B. R. Silliman. 2010. Nutrient enrichment  
26 enhances hidden differences in phenotype to drive a cryptic plant invasion. *Oikos*  
27 119:1776–1784.
- 28 Klironomos, J. N. 2002. Feedback with soil biota contributes to plant rarity and invasiveness in  
29 communities. *Nature* 417:1096–1099.
- 30 Kosuta, S., T. Winzer, and M. Parniske. 2005. Arbuscular mycorrhiza. Pages 87–95 in A. J.  
31 Marquez, editor. *Lotus Japonicus Handbook*. Springer Netherlands.

- 1 Kowalski, K. P., C. Bacon, W. Bickford, H. Braun, K. Clay, M. Leduc-Lapierre, E. Lillard, M.  
2 K. McCormick, E. Nelson, M. Torres, J. White, and D. a. Wilcox. 2015. Advancing the  
3 science of microbial symbiosis to support invasive species management: a case study on  
4 *Phragmites* in the Great Lakes. *Frontiers in Microbiology* 6:95:1–14.
- 5 Křiváčková-Suchá, O., P. Vávřová, H. Čížková, V. Čurn, and B. Kubátová. 2007. Phenotypic  
6 and genotypic variation of *Phragmites australis*: A comparative study of clones originating  
7 from two populations of different age. *Aquatic Botany* 86:361–368.
- 8 Kulmatiski, A. 2018. Community level plant–soil feedbacks explain landscape distribution of  
9 native and non-native plants. *Ecology and Evolution* 8:2041–2049.
- 10 Lane, D. 1991. 16S/23S rRNA sequencing. Pages 115–175 in E. Stackebrandt and M.  
11 Goodfellow, editors. *Nucleic acid techniques in bacterial systematics*. Wiley, New York,  
12 New York, USA.
- 13 Langille, M. G. I., J. Zaneveld, J. G. Caporaso, D. McDonald, D. Knights, J. A. Reyes, J. C.  
14 Clemente, D. E. Burkepille, R. L. V. Thurber, R. Knight, R. G. Beiko, and C. Huttenhower.  
15 2013. Predictive functional profiling of microbial communities using 16S rRNA marker  
16 gene sequences. *Nature Biotechnology* 31:814–821.
- 17 Martin, L. J., and B. Blossey. 2013. The Runaway Weed: Costs and Failures of *Phragmites*  
18 *australis* Management in the USA. *Estuaries and Coasts* 36:626–632.
- 19 McGonigle, T. P., M. H. Miller, D. G. Evans, G. L. Fairchild, and J. a. Swan. 1990. A new  
20 method which gives an objective measure of colonization of roots by vesicular-arbuscular  
21 mycorrhizal fungi. *New Phytologist* 115:495–501.
- 22 Meyerson, L. A., J. T. Cronin, and P. Pyšek. 2016. *Phragmites australis* as a model organism for  
23 studying plant invasions. *Biological Invasions* 18:2421–2431.
- 24 Mozdzer, T. J., J. Brisson, and E. L. G. Hazelton. 2013. Physiological ecology and functional  
25 traits of North American native and Eurasian introduced *Phragmites australis* lineages.  
26 *AoB Plants* 5. DOI:10.1093/aobpla/plt048.
- 27 Nelson, E. B., and M. A. Karp. 2013. Soil pathogen communities associated with native and non-  
28 native *Phragmites australis* populations in freshwater wetlands. *Ecology and Evolution*  
29 3:5254–5267.
- 30 Nguyen, N. H., Z. Song, S. T. Bates, S. Branco, L. Tedersoo, J. Menke, J. S. Schilling, and P. G.  
31 Kennedy. 2015. FUNGuild: An open annotation tool for parsing fungal community datasets



1 by ecological guild. *Fungal Ecology* 20:241–248.

2 Nilsson, R. H., A. F. S. Taylor, S. T. Bates, D. Thomas, J. Bengtsson-palme, T. M. Callaghan, B.  
3 Douglas, G. W. Griffith, R. L. Ucking, A. V. E. Suija, D. L. E. E. Taylor, and M. Teresa.  
4 2013. Towards a unified paradigm for sequence-based identification of fungi. *Molecular*  
5 *Ecology* 22:5271–5277.

6 Ogata, H., S. Goto, K. Sato, W. Fujibuchi, and H. Bono. 1999. KEGG : Kyoto Encyclopedia of  
7 Genes and Genomes 27:29–34.

8 Price, A. L., J. B. Fant, and D. J. Larkin. 2013. Ecology of Native vs. Introduced *Phragmites*  
9 *australis* (Common Reed) in Chicago-Area Wetlands. *Wetlands* 34:369–377.

10 Quast, C., E. Pruesse, P. Yilmaz, J. Gerken, T. Schweer, F. O. Glo, and P. Yarza. 2018. The  
11 SILVA ribosomal RNA gene database project: improved data processing and web-based  
12 tools. *Nucleic Acids Research* 41:590–596.

13 R-Core-Team. 2016. R: A language and environment for statistical computing. R Foundation for  
14 Statistical Computing, Vienna, Austria. <https://www.r-project.org/DOI>:

15 Reinhart, K. O., and R. M. Callaway. 2006. Soil biota and invasive plants. *New Phytologist*  
16 170:445–457.

17 Ryan, R. P., K. Germaine, A. Franks, D. J. Ryan, and D. N. Dowling. 2008. Bacterial  
18 endophytes: recent developments and applications. *FEMS Microbiology Letters* 278:1–9.

19 Salles, J. F., and C. A. Mallon. 2014. Invasive plant species set up their own niche. *New*  
20 *Phytologist* 204:435–437.

21 Saltonstall, K. 2002. Cryptic invasion by a non-native genotype of the common reed, *Phragmites*  
22 *australis*, into North America. *Proceedings of the National Academy of Sciences of the*  
23 *United States of America* 99:2445–9.

24 Saltonstall, K., P. M. Peterson, and R. J. Soreng. 2004. Recognition of *Phragmites australis*  
25 subsp. *americanus* (Poaceae: Arundinoideae) in North America: Evidence from  
26 morphological and genetic analyses. *SIDA, Contributions to Botany* 21:683–692.

27 Schloss, P. D., S. L. Westcott, T. Ryabin, J. R. Hall, M. Hartmann, E. B. Hollister, R. a.  
28 Lesniewski, B. B. Oakley, D. H. Parks, C. J. Robinson, J. W. Sahl, B. Stres, G. G.  
29 Thallinger, D. J. Van Horn, and C. F. Weber. 2009. Introducing mothur: Open-source,  
30 platform-independent, community-supported software for describing and comparing  
31 microbial communities. *Applied and Environmental Microbiology* 75:7537–7541.

- 1 Schulz, B., and C. Boyle. 2005. The endophytic continuum. *Mycological research* 109:661–686.
- 2 Shearin, Z. R. C., M. Filipek, R. Desai, W. A. Bickford, K. P. Kowalski, and K. Clay. 2018.
- 3 Fungal endophytes from seeds of invasive, non-native *Phragmites australis* and their
- 4 potential role in germination and seedling growth. *Plant and Soil* 422:183–194.
- 5 Soares, M. A., H. Li, K. P. Kowalski, M. S. Bergen, M. S. Torres, and J. F. White. 2016.
- 6 Evaluation of the functional roles of fungal endophytes of *Phragmites australis* from high
- 7 saline and low saline habitats. *Biological Invasions* 18:2689–2702.
- 8 Stockinger, H., M. Kruger, and A. Schubler. 2010. DNA barcoding of arbuscular mycorrhizal
- 9 fungi. *New Phytologist* 187:461–474.
- 10 Suding, K. N., W. Stanley Harpole, T. Fukami, A. Kulmatiski, A. S. Macdougall, C. Stein, and
- 11 W. H. van der Putten. 2013. Consequences of plant-soil feedbacks in invasion. *Journal of*
- 12 *Ecology* 101:298–308.
- 13 Thorpe, A. S., and R. M. Callaway. 2006. Interactions between invasive plants and soil
- 14 ecosystems: positive feedbacks and their potential to persist. Pages 323–341 in M. W.
- 15 Cadotte, S. M. McMahon, and T. Fukami, editors. *Conceptual ecology and invasion*
- 16 *biology: reciprocal approaches to nature*. Springer.
- 17 Travers, K. J., C. Chin, D. R. Rank, J. S. Eid, and S. W. Turner. 2010. A flexible and efficient
- 18 template format for circular consensus sequencing and SNP detection. *Nucleic Acids*
- 19 *Research* 38. DOI:10.1093/nar/gkq543.
- 20 Tulbure, M. G., D. M. Ghioca-Robrecht, C. a. Johnston, and D. F. Whigham. 2012. Inventory
- 21 and Ventilation Efficiency of Nonnative and Native *Phragmites australis* (Common Reed)
- 22 in Tidal Wetlands of the Chesapeake Bay. *Estuaries and Coasts* 35:1353–1359.
- 23 Ward, T., J. Larson, J. Meulemans, B. Hillmann, J. Lynch, D. Sidiropoulos, J. R. Spear, G.
- 24 Caporaso, R. Blehman, R. Knight, R. Fink, and D. Knights. 2017. BugBase predicts
- 25 organism-level microbiome phenotypes. *bioRxiv preprint*:1–19.
- 26 White, J. F., K. I. Kingsley, K. P. Kowalski, I. Irizarry, A. Micci, M. A. Soares, and M. S.
- 27 Bergen. 2018. Disease protection and allelopathic interactions of seed-transmitted
- 28 endophytic pseudomonads of invasive reed grass (*Phragmites australis*). *Plant and Soil*
- 29 422:195–208.
- 30 White, T. J., T. Bruns, S. Lee, and J. Taylor. 1990. Amplification and direct sequencing of fungal
- 31 ribosomal RNA genes for phylogenetics. Pages 315–322 in M. Innis, D. Gelfand, J.

1 Sninsky, and T. White, editors. PCR protocols: a guide to methods and applications.  
2 Academic Press, Cambridge, Massachusetts, USA.  
3 Yarwood, S. A., A. H. Baldwin, M. Gonzalez Mateu, and J. S. Buyer. 2016. Archaeal  
4 rhizosphere communities differ between the native and invasive lineages of the wetland  
5 plant *Phragmites australis* (common reed) in a Chesapeake Bay subestuary. Biological  
6 Invasions 18:2717–2728.

Author Manuscript

TABLES

**Table 1.** Community differences by site and lineage in all microbial groups via PerMANOVA.

Predictor variable	Fungi		Bacteria		Oomycetes	
	Pseudo - F	<i>P</i>	Pseudo - F	<i>P</i>	Pseudo - F	<i>P</i>
Lineage	1.06	0.438	1.24	0.298	1.10	0.389
Site	<b>1.68</b>	<b>&lt;0.001</b>	<b>2.43</b>	<b>&lt;0.001</b>	<b>1.24</b>	<b>0.088</b>
Lineage × site	0.96	0.616	<b>1.14</b>	<b>0.050</b>	1.22	0.109

Author Manuscript

**Table 2.** Environmental fit of Soil Parameters to PCoA ordinations of microbial community compositions.

Predictor variable	Fungi		Bacteria		Oomycetes	
	$r^2$	$P$	$r^2$	$P$	$r^2$	$P$
Soil N	0.096	0.179	<b>0.309</b>	<b>0.004</b>	<b>0.135</b>	<b>0.097</b>
Soil P	0.067	0.334	0.049	0.338	0.100	0.193
Saturation	<b>0.399</b>	<b>0.001</b>	<b>0.492</b>	<b>0.001</b>	<b>0.166</b>	<b>0.046</b>

## FIGURE LEGENDS

**Fig. 1.** (a) Percent fungal colonization of roots of Native and Non-native *Phragmites*. Each point represents the average colonization at a site in the non-native patches relative vs. the native patches. Error bars indicate standard error. Diagonal line represents equal colonization among plant lineages. Colors represent sites (blue = BL, red = CB, green = CH, coral = CM, black = CR, brown = PLB, gray = Rt2, purple = SB). (b) Effects of soil saturation on fungal colonization. Black lines indicate the median; bars indicate the full spread of data.

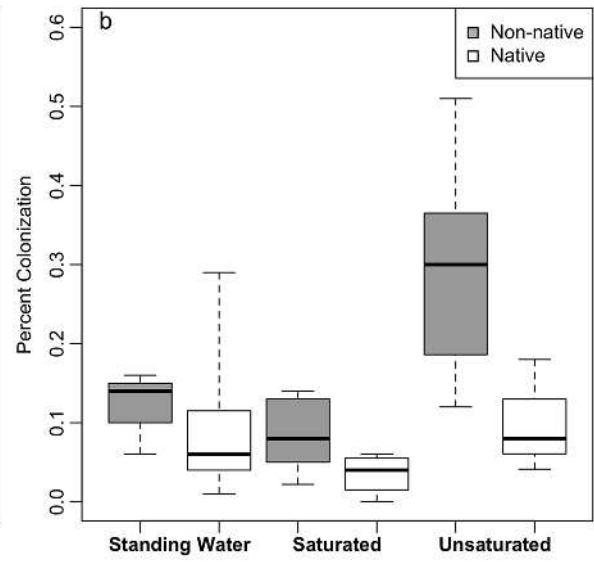
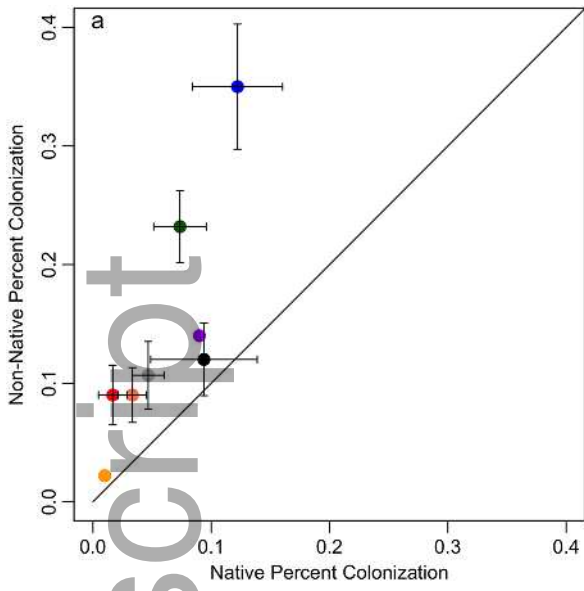
**Fig. 2.** Oomycete Chao Richness and Observed OTUs on Native and Non-Native *Phragmites* Roots. Bars indicate standard error. Significance calculated using ANOVA with Type III Sum of Squares.

**Fig. 3.** Principle coordinate analyses (PCoA) of (a) bacterial, (b) fungal, and (c) oomycete communities. Points represent the microbial community within a single root sample. Vectors displayed are environmental fit of variables (Soil N, Soil P, and Soil Saturation) to ordination space significant at the  $\alpha = 0.10$  level (See Table 2 for *P*-values).

**Fig. 4.** Relative sequence abundance of major bacterial phyla in native and non-native samples. All samples were rarified to a consistent number of sequences (459) making relative abundances comparable. Error bars represent standard error. Significance calculated using ANOVA with Type III Sum of Squares.

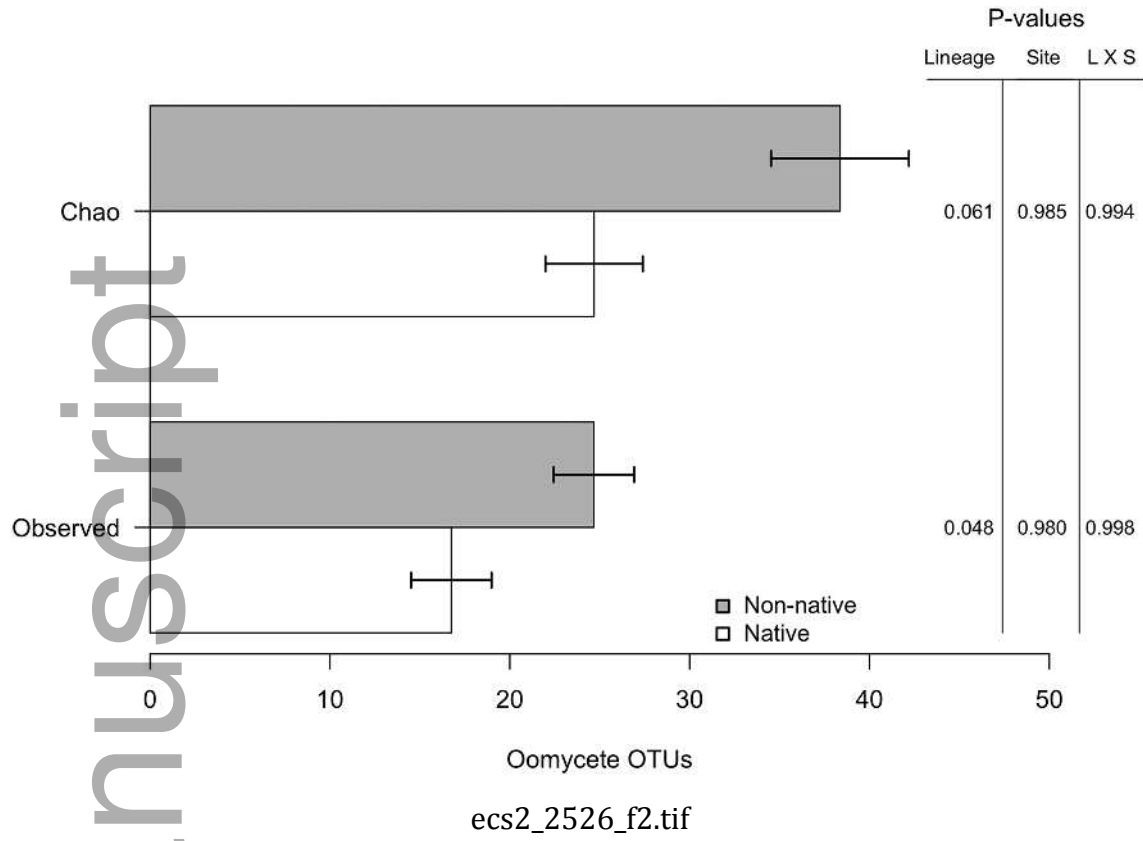
**Fig. 5.** Relative abundance of bacterial classes based on oxygen requirements compared by plant lineage. Oxygen-use classes determined using BugBase. Error bars indicate standard error of the mean. Significance calculated using ANOVA with Type III Sum of Squares.

**Fig. 6.** Relative abundance of trophic modes based on FUNguild determinations. Error bars represent standard error of the mean. Significance calculated using ANOVA with Type III Sum of Squares.

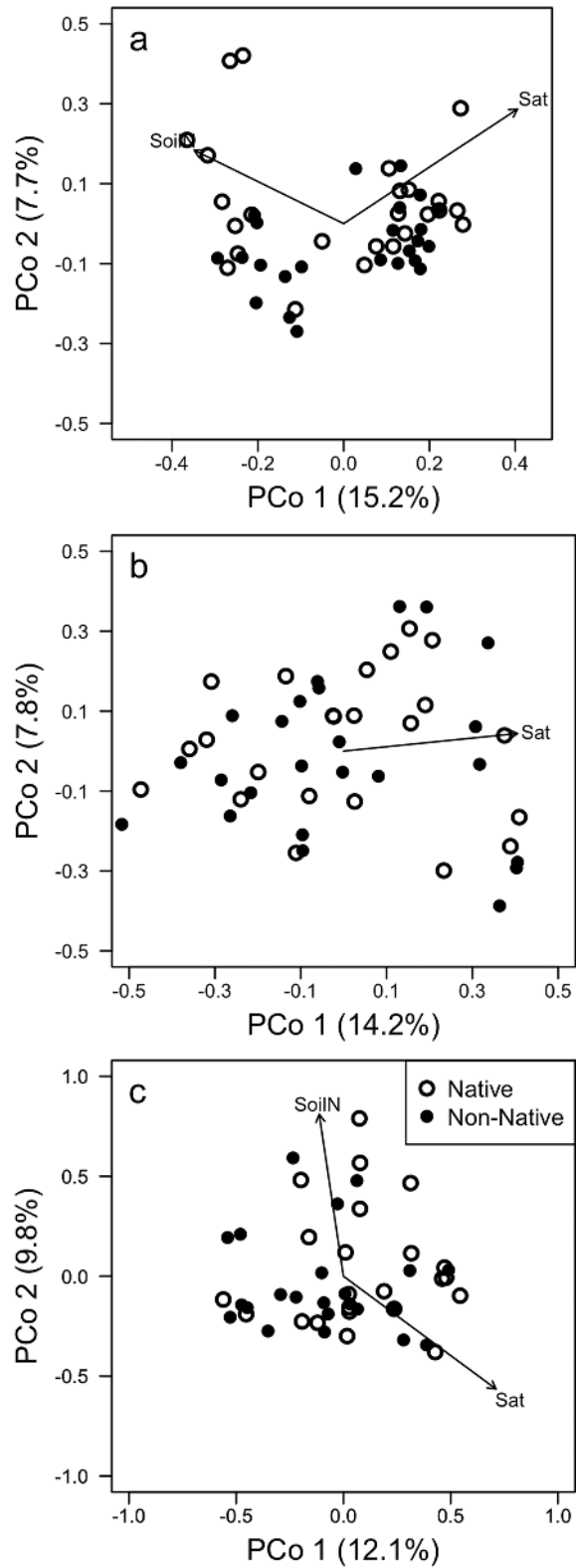


ecs2\_2526\_f1.tif

Author Manuscript







ecs2\_2526\_f3.tif

