

Detecting Urban Emissions Changes and Events with a Near-Real-Time-Capable Inversion System

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Key Points:

- LA CH₄ flux estimates differ by driving meteorology but agree when calibrated for model sensitivity
- Aliso Canyon leak can be detected by inversions using operational meteorology
- Operational meteorology driven inversions significantly detect seasonal emission changes even with only one site

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as doi: [10.1029/2018JD029224](https://doi.org/10.1029/2018JD029224)

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Abstract

In situ observing networks are increasingly being used to study greenhouse gas emissions in urban environments. While the need for sufficiently dense observations has often been discussed, density requirements depend on the question posed and interact with other choices made in the analysis. Focusing on the interaction of network density with varied meteorological information used to drive atmospheric transport, we perform geostatistical inversions of methane flux in the South Coast Air Basin, California in 2015-2016 using transport driven by a locally tuned Weather Research and Forecasting (WRF) configuration as well as by operationally-available meteorological products. We find total-basin flux estimates vary by as much as a factor of two between inversions, but the spread can be greatly reduced by calibrating the estimates to account for modeled sensitivity. Using observations from the full Los Angeles Megacities Carbon Project observing network, inversions driven by low-resolution generic wind fields are robustly sensitive ($p < 0.05$) to seasonal differences in methane flux and to the increase in emissions caused by the 2015 Aliso Canyon natural gas leak. When the number of observing sites is reduced, the basin-wide sensitivity degrades, but flux events can be detected by testing for changes in flux variance, and even a single site can robustly detect basin-wide seasonal flux variations. Overall, an urban monitoring system using an operational methane observing network and off-the-shelf meteorology could detect many seasonal or event-driven changes in near real time – and, if calibrated to a model chosen as a transfer standard, could also quantify absolute emissions.

1 Introduction

Recent years have seen increased efforts to quantify greenhouse gas emissions at or below the scale of individual cities. In complement to process-based inventories [Gurney *et al.*, 2012], aircraft campaigns [Mays *et al.*, 2009; Wecht *et al.*, 2014], and analysis of satellite data [Kort *et al.*, 2012; Ye *et al.*, 2017] among other methods, a common approach has been to deploy a network of sensors within and around a city [McKain *et al.*, 2012; Breon *et al.*, 2014; McKain *et al.*, 2015; Richardson *et al.*, 2016; Shusterman *et al.*, 2016; Pugliese, 2017; Verhulst *et al.*, 2017]. The density and placement of sensors within a network, together with the local meteorology and the spatiotemporal pattern of emissions, determines the extent to which the network is reliably sensitive to emissions over the whole region of interest and within the relevant time scale. Prospective network design

47 studies [e.g., *Kort et al.*, 2013; *Turner et al.*, 2016; *Lopez-Coto et al.*, 2017] have attempted
48 to ensure adequate sensitivity, but the standard of adequacy is necessarily relative to some
49 particular purpose or question.

50 Much urban monitoring work focuses on improving the precision of absolute flux
51 estimates, setting goals such as “to quantify CO₂ and CH₄ emission rates at 1 km² reso-
52 lution with a 10% or better accuracy and precision” [*Davis et al.*, 2017]. Such precision
53 may be a long way off or may not be achievable in every setting; however, a variety of
54 other questions of interest can be answered without precisely constraining the absolute
55 fluxes. For example: what seasonal variations and/or year-over-year trends exist in emis-
56 sions rates, and what fraction of emissions can be attributed to the urban biosphere or to
57 specific anthropogenic source sectors? An operational monitoring system might be able to
58 detect an unusual excursion in the urban flux, and even to suggest a source location, even
59 if the baseline flux is not known accurately.

60 In addition, network density interacts with a host of other factors that also impact
61 the precision and confidence with which the above questions can be answered, including:
62 representation of background concentrations and of the biosphere flux contribution, the
63 statistical method to be used and the choices made in implementing that method (such as
64 the specification of covariance parameters and the choice of a prior), and modeling of me-
65 teorology and of transport processes. This complex web of factors, and their interactions
66 and contributions to the overall uncertainty in modeled posterior fluxes, are only beginning
67 to be understood, especially in the urban setting. In this study, we focus on the meteo-
68 rological driver of transport and how it impacts the inverse results. Future work should
69 consider other factors, including the interaction of data density and driving meteorology
70 with the choice of inversion methodology.

71 Representation of atmospheric transport is believed to be an important source of er-
72 ror in estimating GHG fluxes using atmospheric (in situ or column) observations [*McKain*
73 *et al.*, 2012; *Feng et al.*, 2016]. However, there is no generally-adopted scheme for quanti-
74 fying the effects of transport error. In inversions, some authors simply increase the model-
75 data mismatch covariance across the board to account for transport error [e.g., *Breon et al.*,
76 2014]. *Lin and Gerbig* [2005] proposed using the increase in the variance of modeled con-
77 centrations when the observed error statistics of the wind components are incorporated
78 as additional stochastic variability in the transport model. Recently, *Gourdji et al.* [2018]

79 showed that some of the effects of wind speed error could be mitigated by specifying an
80 additional covariance proportional to the discrepancy in wind speed between model and
81 observations.

82 Along with quantifying transport error, it is difficult to validate transport models or
83 meteorological models in their role as drivers of transport in estimating fluxes for a par-
84 ticular question. On their own, meteorological models can be validated against point ob-
85 servations, most commonly of wind speed and direction and/or mixing depth. Validation
86 of this kind is often used to tune model parameters or to choose a boundary-layer physics
87 scheme or other model configuration [e.g., *Nehrkorn et al.*, 2013; *Feng et al.*, 2016], but
88 does not directly address the fidelity of the transport or the impact on flux estimation.
89 *Deng et al.* [2017] performed a semi-direct evaluation of coupled weather-transport models
90 by comparing the marginal posterior likelihoods of the resulting CO₂ flux estimates. Di-
91 rect validation of transport using controlled release of an inert tracer is also possible [e.g.,
92 *Harrison et al.*, 2012] but rarely included in urban studies.

93 In this study, rather than focus on the optimization of meteorological representation
94 to achieve highest accuracy, highest resolution inversion results, we instead assess whether
95 non-optimized, rapidly available meteorological products can successfully underpin an at-
96 mospheric inversion system. We focus on questions of whether such a system can detect
97 anomalous high emissions events, and whether seasonal flux behaviors can be robustly in-
98 ferred. If a rapidly available meteorological product can successfully underpin such a sys-
99 tem, this indicates near-real time inversions driven by such a product could be conducted
100 and expected to produce statistically useful results in near-real time.

101 To pursue such an approach, we consider Los Angeles as an ideal test case. Cali-
102 fornia has had extensive study and validation of transport models [*Angevine et al.*, 2012,
103 2013; *Zhao et al.*, 2009; *Bagley et al.*, 2017]. A statewide assessment of transport is sum-
104 marized in *Bagley et al.* [2017], and a regional assessment in the greater Los Angeles area
105 in this study indicated little seasonally dependent bias. For Los Angeles specifically, pre-
106 vious work has assessed meteorological representation, determining what could be con-
107 sidered an optimal approach to high-resolution simulations and performing substantive
108 validation [*Feng et al.*, 2016; *Angevine et al.*, 2013].

109 With this meteorological underpinning, *Yadav et al.* [2018] performs inversions in
110 Los Angeles evaluating what can be learned with such an optimized system. In this study,

111 rather than focusing on developing and validating an optimal transport representation, we
112 use the *Yadav et al.* [2018] results as a 'base' case. We compare estimated fluxes from
113 geostatistical inversions driven by this optimized base system with fluxes estimated from
114 geostatistical inversions driven by three broadly-available models or reanalysis products:
115 High-Resolution Rapid Refresh (HRRR), North American Regional Reanalysis (NARR),
116 and the Global Data Assimilation System (GDAS). We evaluate how these different inver-
117 sions perform at determining the absolute flux, detecting both anomalous high emissions
118 events and seasonal flux variance across the basin, and evaluate the role of observation
119 site density in achieving these objectives. Los Angeles provides an opportunistic location
120 for these tests as the large leak from the Aliso Canyon storage facility, which released an
121 estimated 97,100 Mg over four months beginning in October 2015 [*Conley et al.*, 2016],
122 provides what could be considered a tracer release experiment for our purposes. Addition-
123 ally, seasonal variation in methane emissions has been previously observed and reported
124 [*Yadav et al.*, 2018], and also provide a challenge test case for our non-optimized meteorolo-
125 gical drivers.

126 **2 Approach**

127 We perform geostatistical inversions of methane flux between July 1, 2015 and De-
128 cember 31, 2016, using transport driven by each of four meteorological models or reanal-
129 ysis products: WRF, HRRR, NARR, and GDAS. Each product is used to drive the La-
130 grangian transport model STILT [*Lin et al.*, 2003; *Nehrkorn et al.*, 2010] in order to esti-
131 mate the sensitivity of in situ CH₄ mole fraction measurements to emissions fluxes. We
132 estimate fluxes using a geostatistical inversion system based on that developed by *Yadav*
133 *et al.* [2018], with a spatial resolution of 0.03 degrees within the SoCAB and at a tem-
134 poral resolution of four days. The study domain along the coast of Southern California,
135 along with the locations of the observing sites and the Aliso Canyon gas storage facility, is
136 shown in Figure 1.

137 One of the four meteorological drivers we consider, the Weather Research and Fore-
138 casting model (WRF) as configured by *Feng et al.* [2016], has been extensively validated
139 by those authors against observations of wind speed and direction and of PBL height in
140 the Los Angeles area, as well as by comparing forward-modeled CO₂ emissions from the
141 detailed Hestia inventory to in situ and flask mole fraction observations. That validation
142 provided the basis for the WRF runs used in *Yadav et al.* [2018], which are the same ones

143 we use here. The inner WRF domain, which includes the region considered here, has a
144 spatial resolution of 1.3 km and a time step of less than one minute. More details of the
145 WRF setup are given in Supplementary Table S1.

146 To verify that this WRF configuration makes a reasonable base case for a locally-
147 tuned driver of transport, we supplement the existing validation by *Feng et al.* [2016]
148 by directly testing observable meteorological variables in the WRF configuration against
149 those measured at 42 surface observation sites. Agreement is generally good. Across four-
150 day periods between January 2015 and March 2016 (overlapping but not identically with
151 our inversion timeframe), 10 m wind speed bias errors are below 0.5 m/s in 87% of cases,
152 with RMS errors generally in the 1.5 to 2.0 m/s range. Bias errors in 2 m temperature
153 are below 1 K in 92% of cases with RMS errors generally around 1.5 to 2.0 m/s. Despite
154 strong seasonal variation in meteorology in Southern California, we find no discernible
155 seasonality in RMS or bias errors of temperature or wind speed; see Supplemental Figure
156 S1. While future improvements of transport representation are always possible, the combi-
157 nation of past validation and the meteorological comparison presented here establish that
158 it is reasonable for us to treat the WRF system as a representative base case for a locally-
159 tuned driver of transport.

160 In contrast, the NOAA High Resolution Rapid Refresh model (HRRR) [*Benjamin*
161 *et al.*, 2016] has a resolution of 3 km over the continental United States and uses a WRF
162 physics model assimilating radar data every 15min, but is not optimized for the local en-
163 vironment. HRRR output is available as of mid-2015, albeit with some gaps, most notably
164 in August 2016 when the model was upgraded to Version 2. In addition, some STILT runs
165 driven by HRRR fail before the full prescribed simulation period is complete. We exclude
166 from the HRRR inversions any observations for which the necessary HRRR fields are not
167 available or for which the HRRR-STILT sensitivity calculations cover 12 hours or less due
168 to gaps in STILT-HRRR. The latter condition excludes 4.2% of observations, spanning
169 every month of the study period but especially concentrated (6.9%) in November 2015
170 through March 2016. Although the increased failure rate coincides with the Aliso Canyon
171 gas leak, we judge that it remains low enough to permit evaluation of the HRRR-STILT
172 inversion.

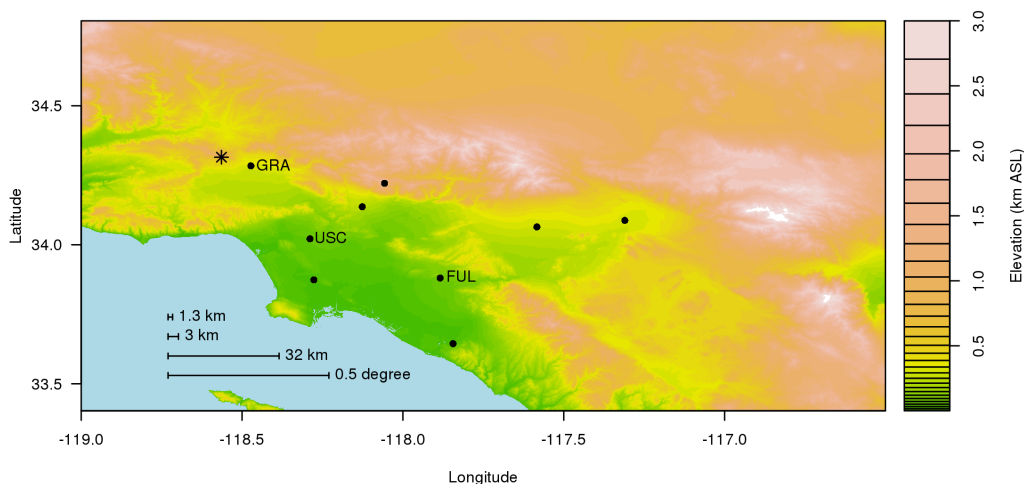
173 The North American Regional Reanalysis (NARR) [*Mesinger et al.*, 2006] and the
174 Global Data Assimilation System (GDAS) are much coarser, with resolutions of 32 km

175 and 0.5 degrees respectively and timesteps of 3 hrs, but cover larger areas (North America
176 and the whole globe). An advantage to using HRRR, NARR, and GDAS is that all are
177 run in a routine operational mode; output can be downloaded from the NOAA READY
178 archive in a format immediately suitable for transport modeling. For low cost, low latency
179 flux estimation in any urban environment, these products are available off-the-shelf.

180 We would not expect coarse products like NARR and GDAS to accurately represent
181 conditions on fine spatial scales within our estimation domain, which spans only about
182 200 km from east to west. The complex topography and sea breeze circulation pattern of
183 the LA basin [Lu and Turco, 1994, 1995] further complicate the environment for transport
184 modeling. Lin et al. [2017] emphasize the failure of transport driven by coarse meteorol-
185 ogy to reproduce the diurnal cycle of CO₂ mole fraction in mountainous terrain. However,
186 several factors may mitigate the effect of poorly resolved topography: while the SOcab
187 domain includes significant elevation changes, most of the observing sites are located in
188 the valley; CH₄ flux generally has a less pronounced diurnal cycle than does CO₂ flux;
189 and, as recommended by Lin et al. [2017] for coarse meteorology, we use only observa-
190 tions taken between 12:00 and 16:00 local time, when the terrain effects are minimized
191 and the representation of vertical mixing is believed to be most reliable.

192 Driven by each meteorological product, STILT simulates the transport of 800 parti-
193 cles 60 hours back in time from each observation. The 60-hour simulation time was cho-
194 sen conservatively to ensure that all recent within-domain influences on the particles are
195 captured. In addition to advection, STILT includes a stochastic component that can simu-
196 late particle motion on spatial and temporal scales shorter than that of the driving meteo-
197 rology, which may help mitigate the effect of using temporally coarse products like NARR
198 and GDAS.

204 Our inversions process data from the surface monitoring network maintained by
205 the LA Megacities Carbon Project, which measures CH₄ mole fractions at nine loca-
206 tions within our domain: Granada Hills (GRA), Mount Wilson Observatory (MWO),
207 Pasadena/Caltech (CIT), downtown LA at the University of Southern California (USC),
208 Compton (COM), CSU Fullerton (FUL), UC Irvine (IRV), Ontario (ONT), and San Bernardino
209 (SBC). Detailed information about each site is given in [Verhulst et al., 2017]. Data avail-
210 ability for each site during the study period is shown in Supplementary Figure S2; an ad-
211 ditional site at Canoga Park (CNP) was not used here because it came online only in Oc-



199 **Figure 1.** Colors: Elevation map of the study domain. Circles: locations of observing sites. The three sites
 200 included in the reduced network are indicated by their three-letter codes. The star in the western part of the
 201 domain indicates the location of the Aliso Canyon facility. Scale bars indicate the grid sizes for the WRF (1.3
 202 km), HRRR (3 km), NARR (32 km), and GDAS (0.5°) meteorological fields, showing the coarse resolution of
 203 the latter fields relative to the domain.

212 tober 2016, at the end of our study period. Background concentrations are estimated as in
 213 *Verhulst et al.* [2017].

214 In order to test the impact of network density, we also perform inversions using a
 215 reduced network and using a single observing site (in addition to the background site).
 216 The single-site inversions use the network's most centrally-located site, at the University
 217 of Southern California in downtown Los Angeles (USC). The USC site was chosen to re-
 218 flect a plausible design for a network consisting of only a single site, which would likely
 219 be designed to be sensitive to as much of the domain as possible at least part of the time.
 220 The reduced-network inversions use the sites at Fullerton (FUL), in the eastern part of the
 221 domain, and at Granada Hills (GRA), in the northwest near the Aliso Canyon facility, in
 222 addition to the USC site. These sites are selected to cover a broad domain in the basin
 223 and because observations are available for these three sites for the vast majority of the
 224 study period. In both the single-site and reduced-network cases, we would expect inver-
 225 sion performance to suffer if sites covering less of the domain were chosen. A complete
 226 description of the observing network is available in *Verhulst et al.* [2017].

227 In all inversions, we employ the geostatistical inversion methodology developed by
228 *Yadav et al.* [2018]. In addition to a model linearly proportional to the distribution of
229 emissions in the CALGEM inventory [*Zhao et al.*, 2009; *Jeong et al.*, 2012], we include
230 a spatially constant model component, since we expect that the inversions using coarse
231 meteorology may be unable to resolve the location of detected fluxes. Note that no input
232 singles out either the location or the time period of the Aliso Canyon natural gas leak. In
233 other words, this inversion makes use of no prior knowledge of the leak. We constrain the
234 methane fluxes to nonnegative values using a bounded version of limited memory BFGS
235 optimization [*Byrd et al.*, 1995], which is well suited to rapidly minimizing functions of
236 many variables and thus facilitates rapid, near-real time calculations. This is different from
237 the Lagrange multiplier approach used in *Yadav et al.* [2018]. Additional subtle differ-
238 ences between the WRF inversion case here and that of *Yadav et al.* [2018] are that we
239 exclude periods in which STILT transport fails using any of our meteorological prod-
240 ucts (as described above), our focused time series is slightly different, and we do not in-
241 clude the Canoga Park site when it comes online late in the time period. These differences
242 are driven by either the motivation to construct a fast, operational system or to ensure we
243 make fair 1:1 comparisons across meteorological products.

244 The nonnegativity constraint on fluxes makes the posterior emissions probability
245 non-Gaussian, which prevents us from calculating posterior uncertainties analytically. Un-
246 certainties can be computed as in *Yadav et al.* [2018] by generating realizations from the
247 posterior covariance distribution. However, each inversion covers only two consecutive
248 four-day periods, the first of which is discarded as a spin-up window. As a result, the
249 posterior uncertainty may not fully account for variation due to changes in the (actual or
250 modeled) sensitivity of the observations to localized surface fluxes. That variation is es-
251 pecially important for our purposes, since we test the detectability of localized flux events
252 and since we use coarse meteorological products in which the footprint of sensitivity may
253 be misplaced even when its magnitude is correct. We therefore rely on the spread of flux
254 estimates across a number of consecutive four-day periods, rather than a calculated uncer-
255 tainty for any given period, as an estimate of variance when testing for flux changes (see
256 section 3.2). For future near-real-time applications, this method has the additional advan-
257 tage of saving the computing time needed to generate the realizations.

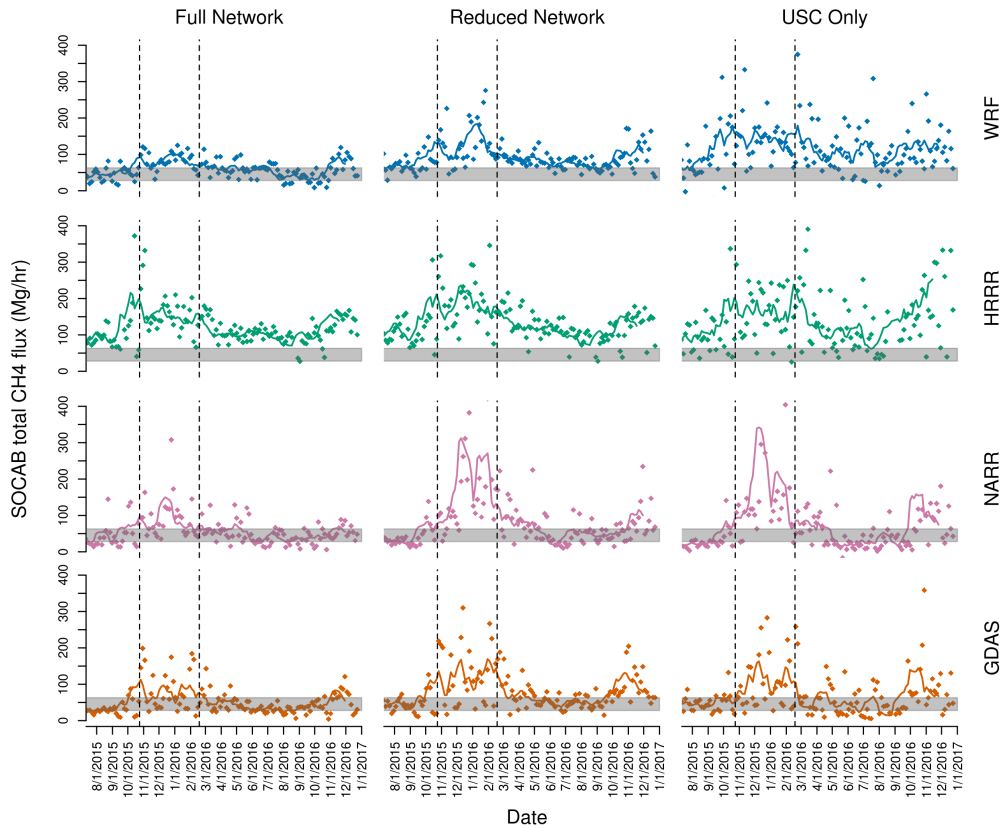
3 Analysis

3.1 Basin Total Flux

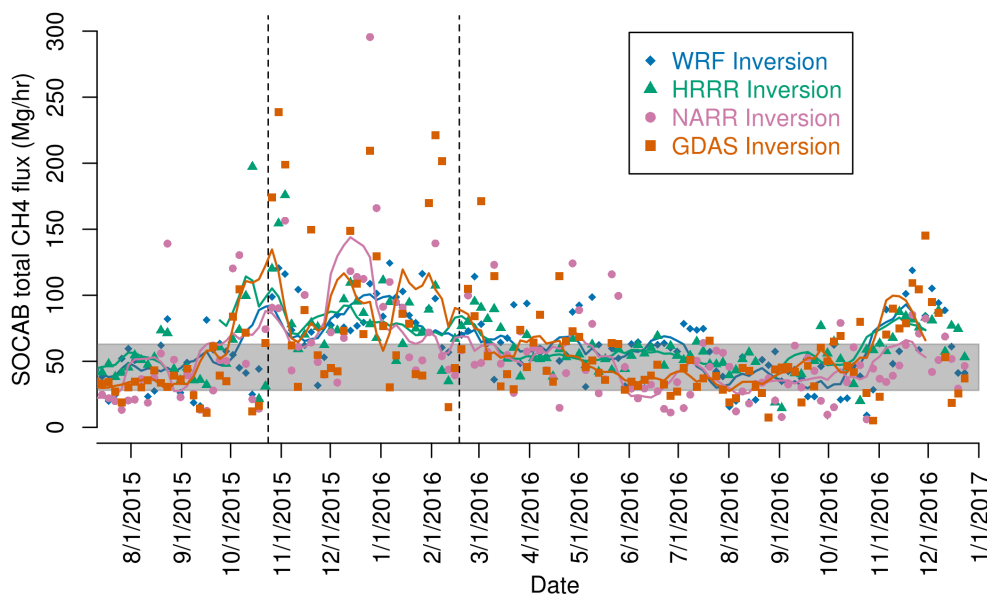
Estimated whole-basin methane fluxes from each of the four inversions are shown in Figure 2. The Aliso Canyon event and seasonal cycle, known features we are using to test operational meteorologies, appear evident in all inversions and we assess this statistically in section 3.2. All inversions show emissions up-ticks prior to the start of the Aliso Canyon event, which could be indicative of the leak beginning before the noted start date, or part of the seasonal increase in emissions. While this study does not attribute this feature, note that it is not explained by the timing of the four-day periods used in the inversion, since the increase begins in periods which do not overlap the reported leak. Considering emissions magnitudes, when the full observing network is included, estimates using transport driven by WRF and NARR average 53 and 47 Mg/hr outside the Aliso Canyon leak period, respectively, in broad agreement with the 35 to 50 Mg/hr range of baseline emissions estimates in other studies [e.g. *Wennberg et al.*, 2012; *Peischl et al.*, 2013; *Wecht et al.*, 2014; *Wong et al.*, 2015]. That our estimates fall at the upper end of that range is not surprising given that much of the previous work relied on observations taken in May-June 2010, not during the peak of the seasonal emissions cycle (see section 3.2). Estimates using HRRR are considerably higher than those using WRF, by about 96% on average over the 18-month study period, and estimates using GDAS are somewhat lower, by about 16% on average.

Much of the difference in estimated flux is explained by the difference in overall mean total sensitivity assigned by each model to the measurement network. We compute the mean total sensitivity H_{mean} for each model over the 18-month period of the study by summing the sensitivity of the nine measurement sites, then taking the mean over spatial flux grid cells and over observation times. In order to make a direct comparison, we exclude (for all models) observations for which HRRR fields are missing or for which HRRR-STILT runs failed; see Section 2. Treating WRF as a transfer standard, we perform an empirical calibration, scaling the posterior fluxes s_j from the NARR, HRRR, and GDAS-driven inversions (j) by the ratios of the sensitivities computed using those models relative to those using WRF:

$$\mathbf{s}_{\text{cal},j} = \frac{H_{\text{mean},j}}{H_{\text{mean,WRF}}} \times \mathbf{s} \quad (1)$$



278 **Figure 2.** Points: estimated total CH₄ flux time series for the South Coast Air Basin (SoCAB), at four-day
 279 time intervals, according to inversions using transport driven by each of four meteorological models and using
 280 the full observing network (9 sites), a reduced network (3 sites), or a single observing site. Curves: 28-day
 281 running means of each time series for visual reference (not used in the analysis). The shaded band indicates
 282 the typical range of estimates in past studies. The dashed vertical lines indicate the start and end dates of the
 283 Aliso Canyon natural gas leak.



304 **Figure 3.** Estimated SoCAB total CH₄ flux time series in inversions using the full observing network after
 305 calibration by scaling the fluxes by the relative total sensitivity assigned to the observing network by each
 306 driver of the transport model. The calibration brings the estimates into close agreement overall. Curves:
 307 28-day running means of each time series for visual reference (not used in the analysis). The shaded band
 308 indicates the typical range of estimates in past studies. The dashed vertical lines indicate the start and end
 309 dates of the Aliso Canyon natural gas leak.

294 After calibration, the mean posterior emissions $s_{\text{cal},j}$ come into much closer alignment
 295 overall. The difference in mean flux over the full 18-month study period relative to the
 296 WRF inversion is reduced to 17% with HRRR and 1% with GDAS and increases mod-
 297 estly to 3% with NARR. The scaled time series are shown in Figure 3. As we look at in-
 298 creasingly shorter time scales, more scatter remains between the calibrated flux estimates.
 299 The mean residual difference between monthly mean fluxes from the WRF inversion and
 300 calibrated estimates over the same periods from the other inversions is about 20% with
 301 HRRR and NARR and about 25% with GDAS. Individual four-day flux estimates after
 302 calibration are moderately well correlated overall, $r = 0.47$ to 0.50 , but often diverge (see
 303 Supplementary Figure S3).

310 If the sensitivity bias could be corrected using direct observations, our results sug-
 311 gest that accurate flux estimates might be possible, at least one monthly and longer time
 312 scales, using more widely available models than is generally assumed. However, several of

the meteorological factors most clearly linked to the sensitivity fail to explain the difference. STILT computes sensitivity to surface fluxes by tracking the amount of time simulated air parcels spend in contact with the surface. The sensitivity H_{ij} of the i th observation to the j th flux region is given by [Lin *et al.*, 2003]

$$H_{ij} = \frac{m_{air}}{\rho_j} \frac{\tau}{z_j}; \quad \tau = \frac{1}{N_i} \sum_{p_i=1}^{N_i} \Delta t_{p_i,j} \quad (2)$$

where z_j is the mixing depth, accounting for the effect of dilution, and τ is the average time spent by the parcels within the bottom one-half of the mixing layer above the flux region. The average is taken over N_i simular parcels released backwards from the i th observation and indexed by p_i . On the basis of these relations, we would expect the intermodel differences in sensitivity to be explained by systematic differences either in the mixing height or in the residence time, i.e., the time for air to travel from the edge of the study domain to the observing site, as driven by the wind speed.

In the STILT runs driven by each model or reanalysis product, we computed the mean time spent in the domain by measured air parcels before encountering an observation site (residence time) as well as the time-averaged mixing depth along the parcel's path. The same filtering was applied as in computing the mean sensitivities. As shown in Table 1, the results do not explain the differences in sensitivity. On average, mixing depths in HRRR are almost the same as those in WRF, and residence times are only modestly shorter – yet the sensitivity is much less. On the contrary, mixing depths in NARR are 80% higher on average than those in WRF, yet the sensitivity is very similar.

Since parcels may be within the horizontal extent of the domain but above the bottom half of the mixing layer (and therefore considered by STILT to be insensitive to surface fluxes), we also computed the fraction of their residence time that measured parcels spent near the surface. As shown in Table 1, this ‘near-surface fraction’ differs from WRF by no more than 13% in any of the other models. The expected combined effect of the mixing depth, residence time, and near-surface fraction is summarized on the fourth line of Table 1, in which we compute the relative sensitivity predicted by those mean variables according to

$$\frac{H_{\text{mean}}}{H_{\text{mean,WRF}}} = \frac{z_{\text{WRF}}}{z} \times \frac{\tau}{\tau_{\text{WRF}}} \times \frac{f}{f_{\text{WRF}}} \quad (\text{predicted}) \quad (3)$$

where f is the near-surface fraction. The resulting prediction fails to capture the actual differences in total mean sensitivity, which are given on the last line of Table 1.

	WRF	HRRR	NARR	GDAS
Mixing Depth (m)	615	612 / 99%	1109 / 180%	573 / 93%
Residence Time (min)	315	278 / 88%	250 / 79%	308 / 98%
Near-Surface Fraction	0.57	0.49 / 87%	0.65 / 115%	0.45 / 80%
Predicted Relative Sensitivity	-	/ 77%	/ 51%	/ 84%
Actual Relative Sensitivity	-	/ 53%	/ 96%	/ 120%

332 **Table 1.** First three rows: mean values of meteorological variables expected to contribute to sensitivity, for
 333 STILT driven by each of four models or reanalysis products. These variables are described in section 3.1, and
 334 percentages are relative to the same variables in WRF. Fourth row: expected ratios of the sensitivity in HRRR,
 335 NARR, and GDAS, relative to that in WRF, given the above variables. Fifth row: actual ratios of the sensitivity
 336 in HRRR, NARR, and GDAS to that in WRF. The actual relative sensitivities are not accurately predicted on
 337 the basis of the mean meteorological variables.

348 Therefore, although basin-wide, 18-month-average sensitivity explains the gross dif-
 349 ferences in estimated flux between the inversions, the basin-wide, 18-month-average dif-
 350 ferences in the relevant underlying meteorological variables do not control the sensitivity
 351 in the same way. In the transport model, the whole basin is not treated as a single region;
 352 rather, Equation 2 applies separately in each 0.03-degree grid cell and for each four-day
 353 period, and the fine-scale interactions between the variables have a substantial effect.

354 An important implication is that our modeled average sensitivities could not be cal-
 355 ibrated to ground truth by debiasing the underlying meteorological variables in a basin-
 356 averaged manner. For example, using lidar observations in Pasadena, California (colocated
 357 with one of the LA Megacities observing sites), *Ware et al.* [2016] showed that NARR
 358 persistently overestimates the mixing depth at that location, by more than a factor of two
 359 on average, and that any local mixing depth bias in WRF was likely much smaller. In-
 360 deed, we can see in Table 1 that mixing depths in NARR are very high on average over
 361 the whole domain. However, if the estimated fluxes in the NARR inversion were scaled to
 362 correct for this bias as suggested by *Ware et al.* [2016], the result would be to introduce a
 363 large positive bias into the fluxes. Of course, wind speed and mixing depth observations
 364 can be used to evaluate and improve meteorological drivers of transport, as was done for

365 the WRF configuration employed here by *Feng et al.* [2016] – but our results show that a
366 mean calibration factor constructed from those observations could not be reliably correct.

367 We might expect that the mean meteorological variables would better predict the to-
368 tal sensitivity over shorter time periods, since correlations between the variables might be
369 less important. However, we find that this is not the case on monthly timescales (see Sup-
370plementary Table S2), nor do calibration factors constructed from monthly average total
371sensitivities perform as well as the calibration factors calculated over the full 18-month
372study period. Calibration factors computed seasonally do somewhat better, but in most
373cases, seasonal mean fluxes come into closer agreement after applying the full 18-month
374calibration than after applying seasonal calibration. Overall, the calibration method seems
375to be most effective when applied over a year or more.

376 One alternative to computing calibration factors from meteorological observations
377could be to run a trusted custom model for a limited period, compute a calibration using
378the mean sensitivity for that period, then continue estimating fluxes using an operational
379product. Though the time period of our study is too limited for a conclusive demonstra-
380tion, our experience suggests that this approach could be successful. We computed cal-
381ibration factors for each of HRRR, NARR, and GDAS based on the first twelve months
382of the study period, then applied those factors to the flux estimates for the last six months,
383July-December 2016. That calibration reduced the difference in mean flux between HRRR-
384and WRF-driven inversions from 103% to 2% and between GDAS- and WRF-driven in-
385versions from 15% to 6%, though it increased the difference between NARR- and WRF-
386driven inversions modestly, from 16% to 22%.

387 **3.2 Anomaly and Trend Detection**

388 We evaluate the ability of each inversion system to detect changes in the total basin
389flux, both seasonally and due to an unusual event or change. We test significance using
390Welch's unequal-variances t-test, which has similar power to a standard t-test and is ap-
391propriate whether or not the samples to be compared have the same variance. The signifi-
392cances (p-values) for all the tests described in this section are given in Table 2.

393 In all of the inversions using the full observing network, we observe a seasonal
394trend in CH₄ emissions. Emissions in November-December 2016 are estimated to be 38%
395(NARR inversion) to 83% (GDAS inversion) higher than those in July-August. These pe-

396 riods were selected so as not to overlap the timeframe of the Aliso Canyon leak, in order
397 to isolate the ‘normal’ seasonal difference. The estimated difference is significant at the
398 95% level or better in all four inversions. The consistent detection and timing of the sea-
399 sonal change, regardless of the meteorology used to drive transport, reinforce its status as
400 a robust and substantial feature of Los Angeles methane emissions.

401 We also test the detectability of the increase in flux during the Aliso Canyon leak
402 period. To remove the impact of the seasonal dependence, we compare the period October
403 24 through December 27, 2015 to the corresponding period in 2016 (in an operational
404 setting, the comparison would generally be to previous years). The difference is significant
405 at the 95% level in Welch’s t-test in the HRRR, NARR, and GDAS inversions but much
406 less significant ($p=0.17$) in the WRF inversion. Note that this test of event detectability is
407 distinct from quantifying the rate of a known leak as in *Yadav et al.* [2018].

408 Our ability to observe the Aliso Canyon gas leak using the LA Megacities observ-
409 ing network is limited by its position near the edge of the inversion domain, such that its
410 emissions are observable only intermittently. However, as is apparent in Figure 2, this in-
411 termittency can result in an increase in the variance of the retrieved fluxes, which may be
412 significant even, or indeed especially, when the change in mean is not. In fact, in an F-test
413 for difference of variance comparing October-December 2015 to 2016 as above, the in-
414 crease in retrieved flux variance during the Aliso Canyon period is nearly as significant or
415 more significant than the change in mean flux in the inversions driven by HRRR, NARR,
416 and GDAS. The increase in variance is not significant ($p=0.32$) in the inversion driven by
417 WRF, which shows the least variability relative to the estimated flux values. These results
418 highlight the complimentary value of the two approaches, particularly for less-optimized
419 meteorology.

426 That the inversion driven by WRF does not significantly detect the Aliso Canyon
427 event using our tests may be surprising. One plausible explanation is that, during the leak
428 period, the WRF inversion produces consistent but only moderately elevated flux esti-
429 mates. This moderate increase is not sufficient to distinguish itself from the corresponding
430 increase in late 2016. By contrast, the other inversions produce exceptionally high esti-
431 mates for some four-day periods. Even though estimates for other periods are not elevated,
432 the average increase is sufficient for detection.

a.) Seasonal Difference, Welch's t-test

	WRF	HRRR	NARR	GDAS
Full Network	0.047*	<0.001*	0.048*	0.012*
Reduced Network	0.024*	<0.001*	0.075	<0.001*
USC Site Only	0.53	0.0012*	0.025*	0.015*

b.) Aliso Canyon Period, Welch's t-test

	WRF	HRRR	NARR	GDAS
Full Network	0.17	0.025*	0.016*	0.039*
Reduced Network	0.63	0.004*	0.051	0.30
USC Site Only	0.15	0.39	0.24	0.89

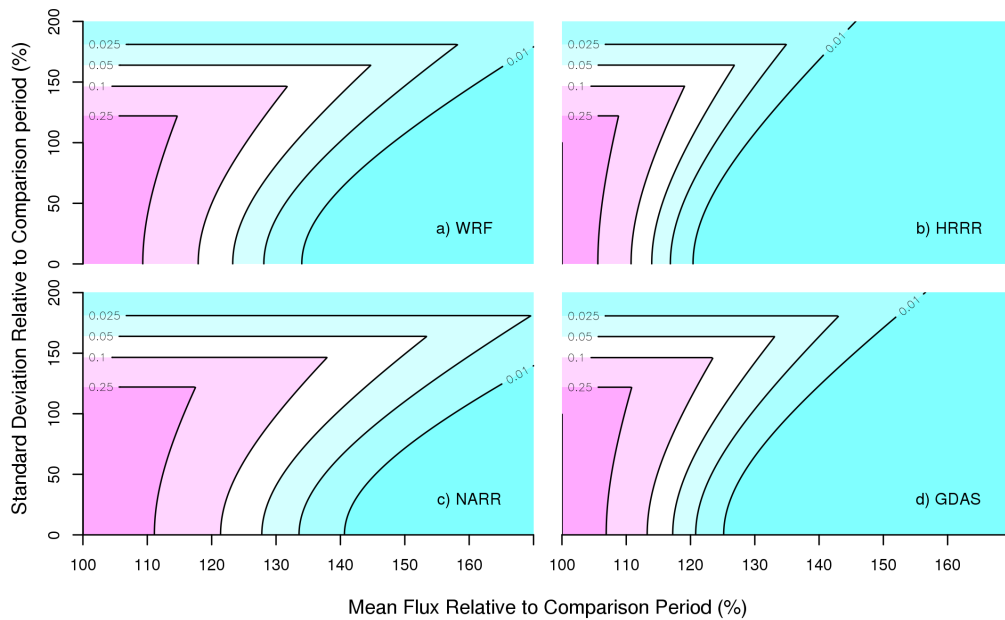
c.) Aliso Canyon Period, F-test for Difference of Variance

	WRF	HRRR	NARR	GDAS
Full Network	0.32	<0.001*	<0.001*	0.044*
Reduced Network	0.60	0.056	0.016*	0.021*
USC Site Only	0.45	0.21	0.36	0.82

420 **Table 2.** Summary of p-values of two-sided tests for changes in mean emissions (a and b) or variance of
 421 emissions (c), comparing summer to winter of 2016 (a) or the first 64 days of the Aliso Canyon gas leak in
 422 2015 to the equivalent period in 2016 (b and c). Tests significant at the 95% level are indicated with an aster-
 423 isk. Seasonal flux differences are detected in most cases even with reduced observations; the Aliso Canyon
 424 leak is detected with the full network in the non-WRF inversions and with the reduced network in some cases
 425 using the test of difference of variance.

433 The difference in variability between the WRF inversion and the others may be due
434 to the assignment of covariance parameters according to Restricted Maximum Likelihood
435 (RML) analysis. Rather than assign prior uncertainties by expert judgment, RML finds the
436 combination of covariances that make the actual observations most likely, given the sen-
437 sitivity footprints computed by the transport [Michalak *et al.*, 2004]. The variances of the
438 observations and the spatial pattern of prior covariance are therefore intermediate statisti-
439 cal quantities which are calculated during the course of the inversion. In our WRF-driven
440 inversion, RML assigns most of the prior covariance to the spatially constant pattern. The
441 result is that the cost of attributing an observed excess mole fraction to a flux is mostly
442 insensitive to the spatial distribution of the observation's sensitivity footprint. In the other
443 inversions, although the magnitude of prior covariance is similar *on average*, RML assigns
444 more weight to the spatial pattern proportional to the CALGEM inventory, so the penalty
445 for assigning an excess flux is more spatially variable. This would tend to make the in-
446 version more sensitive to the modeled wind direction, which may not be accurate. If the
447 footprint of a high observed mole fraction falls over a source known to CALGEM, the
448 flux estimate can be increased a great deal at little cost; but if the footprint falls over an
449 area without sources in CALGEM, increasing the flux estimate is costly.

450 In general, the threshold for a flux event to be detectable by a given observing and
451 inversion system depends not only on the magnitude of the event but also on its duration
452 and variance. It also depends on the event's timing, because the mean flux and variance
453 during the reference period used for comparison will vary according to the seasonal cy-
454 cle. By way of an example, for a hypothetical event persisting at least from September 4
455 to October 26, 2017 (and compared to the corresponding period in 2016), we compute the
456 sensitivity according to the better of Welch's t-test and the F-test for difference of vari-
457 ance for a range of estimated flux increases and variances. The results are shown in Fig-
458 ure 4 for the inversions driven by each of the four meteorological products. In this exam-
459 ple, a flux increase estimated at 30-40% above the baseline by an inversion using WRF
460 or NARR would be detected as significant if the variance were approximately unchanged.
461 The same is true for an increase estimated at 20-30% by the inversion using GDAS or es-
462 timated at about 20% by the inversion using HRRR. Note, however, that the same thresh-
463 olds do not persist at other times and that the threshold for the actual flux increase due to
464 an event may be higher if the event is not consistently upwind of the observing sites.



465 **Figure 4.** Sensitivity (p-values) of inversions using each meteorological driver to hypothetical flux events
 466 occurring between September 4 and October 26, 2017, as a function of the change in mean flux and vari-
 467 ance relative to the same period in 2016. The inversions shown here use the full observing network (9 sites).
 468 Changes in mean flux are less significant when accompanied by high variance, but sufficiently large variance
 469 increases are themselves significant in an F-test.

3.3 Network Density

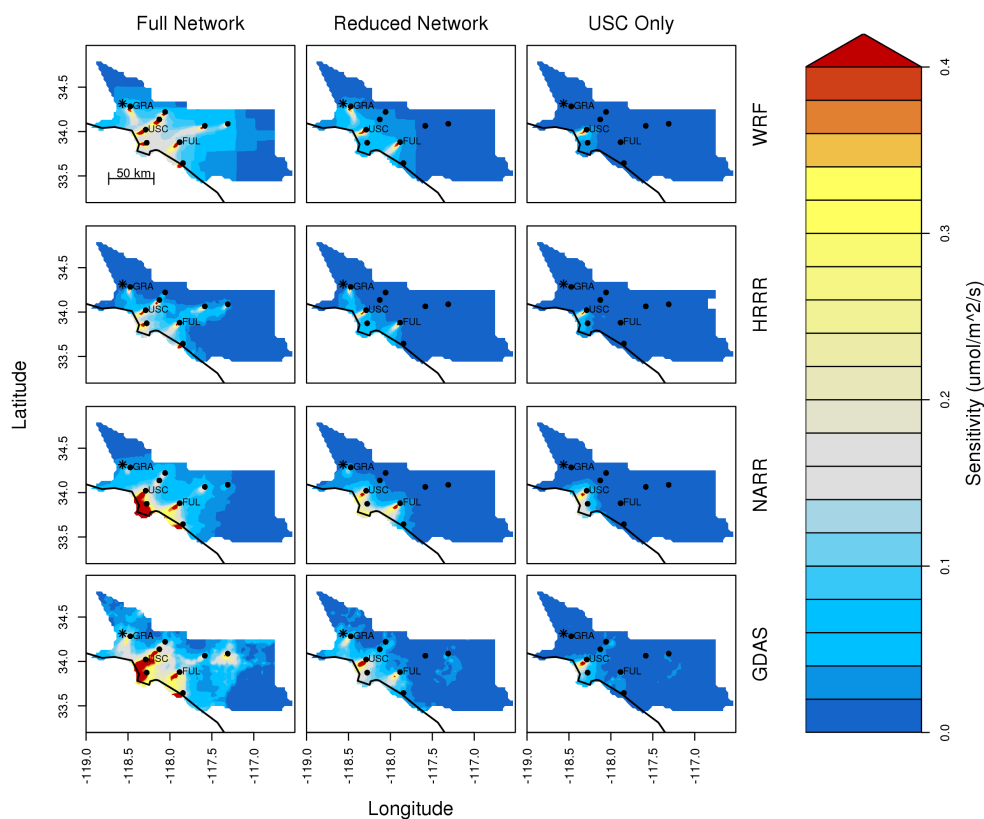
As the number of observing sites is reduced, the methane flux retrievals generally become noisier, exhibiting greater variance even in the absence of any known flux event. In almost all cases, robustly detecting the Aliso Canyon leak event is more difficult with only three observing sites than with the full network. However, the HRRR-driven inversion remains sensitive to the change in mean flux ($p=0.004$) and the NARR- and GDAS-driven inversions remain sensitive to the increase in variance ($p=0.016$ and $p=0.021$, respectively).

With only a single observing location, none of our inversions can detect a significant change either in the mean or in the variance of the fluxes during the Aliso Canyon leak. The USC site alone can constrain only a small part of the study domain, and even that part only inconsistently. Figure 5 illustrates the decrease in measurement constraint when the number of the number of observing sites is reduced.

By contrast, even a single measurement location is sufficient in most of our inversions (excepting that using WRF) to observe the seasonal cycle. Broad and consistent sensitivity may be less critical for this purpose than for detecting a point source event because the seasonal difference is likely to be widely distributed throughout the domain. Although our study period is too short to observe it, we might expect the same to apply to year-over-year secular changes.

4 Conclusions

Our results suggest that the ability of an in situ observing network to detect changes in emissions may be less sensitive to the choice of transport driver than are estimates of the absolute total flux. Much of the difference in absolute flux estimates between inversions driven by divergent meteorology seems to be attributable to biases in long-term sensitivity, which can be calibrated by comparison to a trusted model chosen as a transfer standard. Debiasing with weather observations (e.g. scaling results by observed bias in mixing depth) would not be successful as the sensitivity bias is not predicted by the mean values of the relevant underlying meteorological variables. However, an accurate total estimate is not a prerequisite for observing changes, including seasonally or in the case of leaks or other large anomalies. Although our study period is not long enough to directly observe, trends over the course of years could likely be characterized in the same way. We



483 **Figure 5.** Heat map: sensitivity of the full observing network (9 sites), a reduced network (3 sites), and the
 484 USC site alone to fluxes within the SoCAB during the first four days of the Aliso Canyon natural gas leak, Oc-
 485 tober 24-27, 2015, as computed by STILT driven by each of four meteorological products. Circles: locations
 486 of observing sites. The three sites included in the reduced network are indicated by their three-letter codes.
 487 The star near the western edge of the domain indicates the location of the Aliso Canyon facility. The breadth
 488 and magnitude of sensitivity degrade as measuring locations are removed.

507 find that even with only a single observing site, seasonal flux changes emerge as robustly
508 detectable with operational meteorology supporting an inversion, suggesting sparse urban
509 networks can potentially provide valuable, rapid information.

510 The ability of a surface network to detect flux changes contributes to the function-
511 ing of a ‘tiered’ observing system [Duren and Miller, 2012] for megacities carbon emis-
512 sions, which includes continuous monitoring at the urban scale, targeted deployments to
513 characterize significant individual sources, and regional or boundary condition data from
514 aircraft and satellites, as well as bottom-up inventories. A flux inversion system run oper-
515 ationally could provide the first notice of events worthy of more detailed investigation by
516 other methods. The more quickly these events can be identified, the better opportunity we
517 will have to quantify and characterize them as well as to inform stakeholders.

518 So far, the ability to usefully detect emissions events using urban concentration mea-
519 surements has been limited by the long time delay, typically measured in years, between
520 collecting initial data and producing a flux estimate. (An exception was the near-real-time
521 monitoring performed by *Lauvaux et al.* [2013] in Davos, Switzerland in 2011-2012.) One
522 major source of latency is the time, expense, and computational resources involved in me-
523 teorological modeling for transport. Others have begun demonstrating forward model sim-
524 ulations using operational meteorology *Pugliese* [2017]. We now have demonstrated that
525 at least some operational monitoring goals utilizing atmospheric inversions can be met us-
526 ing a variety of meteorological products, including several that are made available on a
527 routine basis and nearly in real time. Output from HRRR is posted on the NOAA READY
528 archive each day, covering the previous day. Continuous archival of GDAS has recently
529 been supplanted by Global Forecast System (GFS) short-term forecasts, which are initial-
530 ized with GDAS but have twice the resolution both in space (0.25 degrees) and in time
531 (3 hours). GFS zero-hour forecasts are finalized the same day, and since GFS covers the
532 whole globe, they can be retrieved for the vicinity of any major city or other area of inter-
533 est. Our work shows that the coarse spatial resolution of these products does not necessar-
534 ily limit their utility in an urban setting.

535 Once the meteorological fields are ready, the remaining computational requirements
536 can be modest. For this study, calculating influence footprints with STILT using HRRR
537 meteorology took about fifteen minutes for each observation on a 2.2 GHz CPU with 128
538 GB of RAM. In total, running footprints for up to 16 observations in parallel, the foot-

539 prints for a single inversion covering two consecutive four-day periods took about 5.5
540 hours to calculate. In an operational mode, each day's footprints could be run the next
541 day, taking less than one hour. The geostatistical inversions themselves each took only
542 about two minutes, although that time would be longer if we computed posterior covari-
543 ances as in *Yadav et al.* [2018] or, especially, if we allowed off-diagonal terms in the prior
544 covariances.

545 This suggests that the remaining obstacle for an operational near-real time inversion
546 system lies not in latency of meteorological drivers, flux priors, or inversion calculation,
547 but instead on the rapid collection of QA/QC'd network observations, and in cases where
548 global models are used for background concentrations, the latency of those global model
549 runs. Given that this work suggests fluxes can be estimated rapidly once concentration
550 data is collected and quality-controlled, accelerating this step could see a near-real time
551 system actually implemented.

552 **Acknowledgments**

553 The authors thank Thomas Nehrkorn for assistance with observational validation of WRF
554 meteorology. Support was provided by the NIST Greenhouse Gas and Climate Science
555 Measurements Program, including under grant 70NANB17H176, by the NOAA Atmo-
556 spheric Chemistry, Carbon Cycle, and Climate Program, and by NASA under grant NNN12AA01C.
557 Portions of this work were performed at the Jet Propulsion Laboratory, California Institute
558 of Technology, under contract with NASA.

559 Mole fraction data used in this study are provided in a supplementary spreadsheet.
560 Up to date data from LA basin including other time periods can be found at the Megac-
561 ities Carbon Project web portal at <https://megacities.jpl.nasa.gov/portal/>. Meteorological
562 fields from HRRR, NARR, and GDAS are available on the NOAA READY archive at
563 <https://www.ready.noaa.gov/archives.php>.

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Figure 1.

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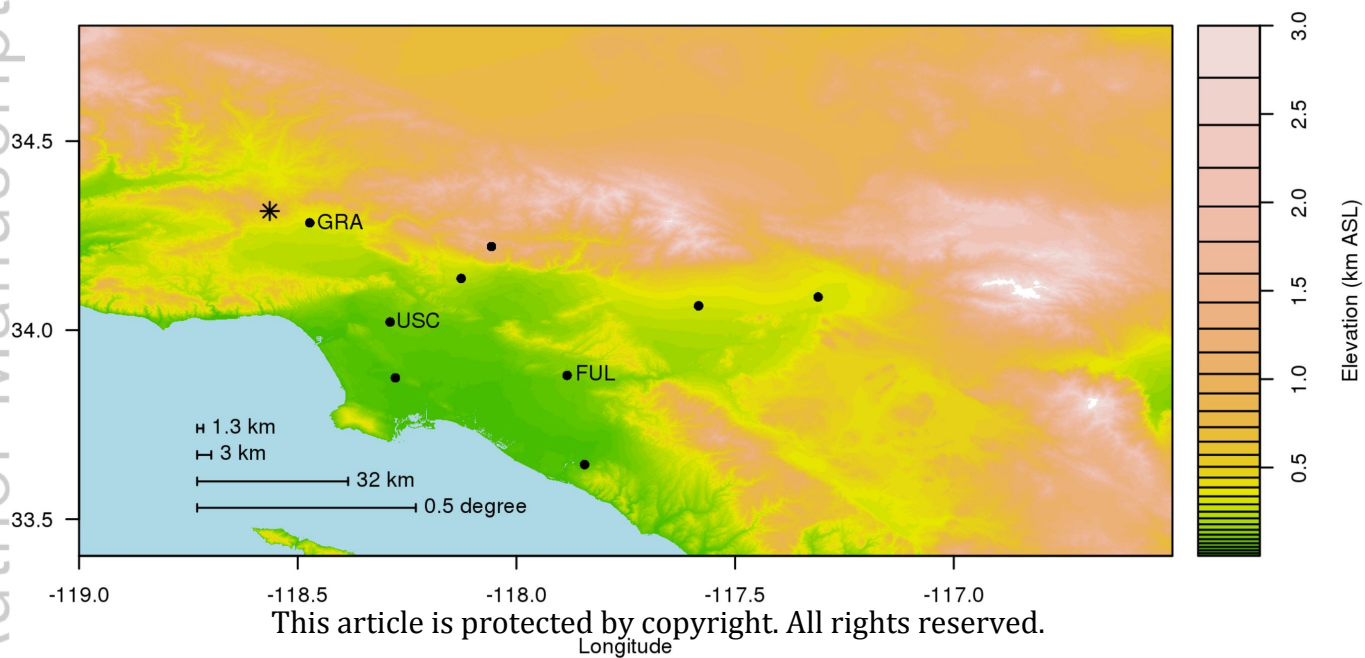


Figure 2.

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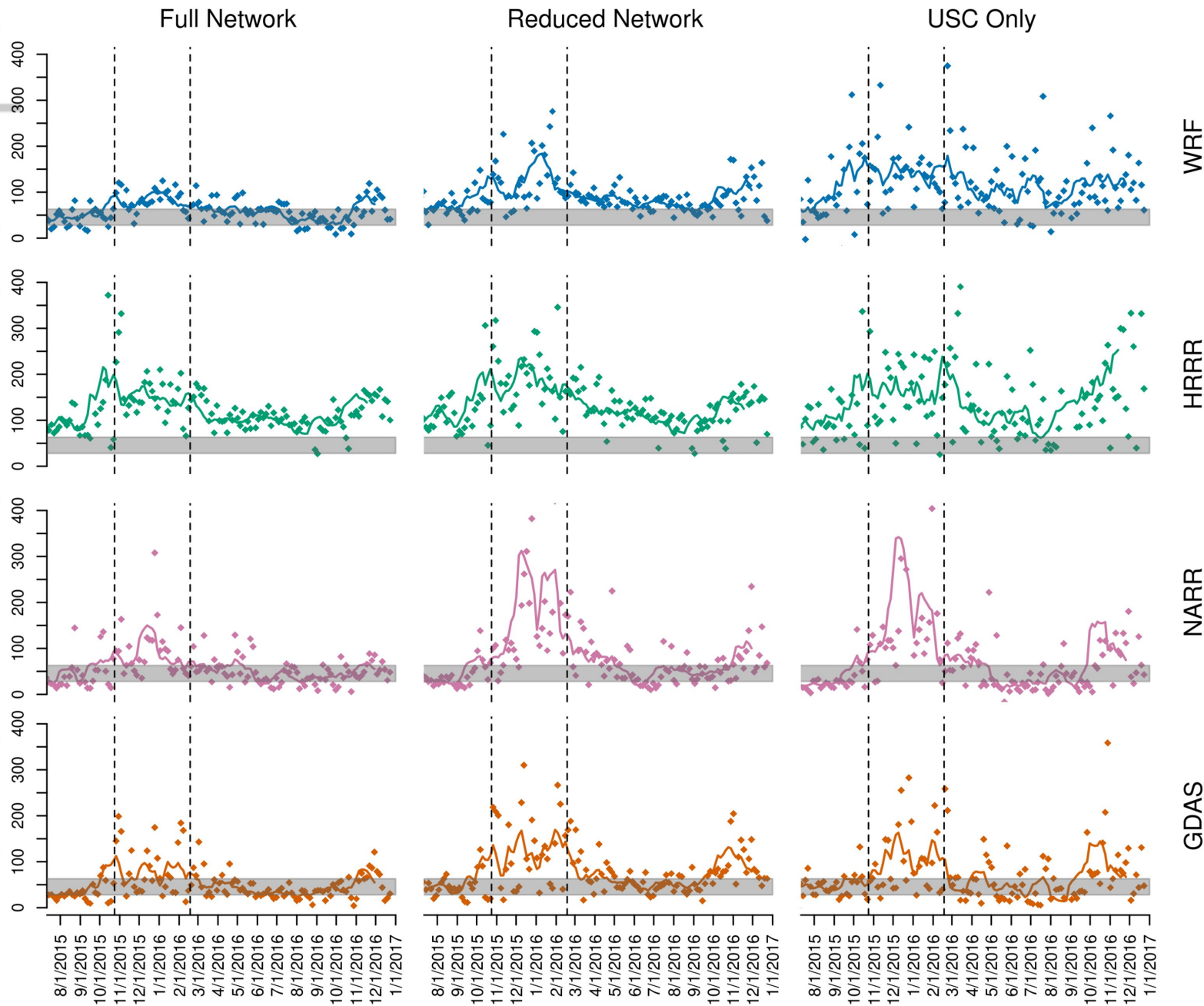


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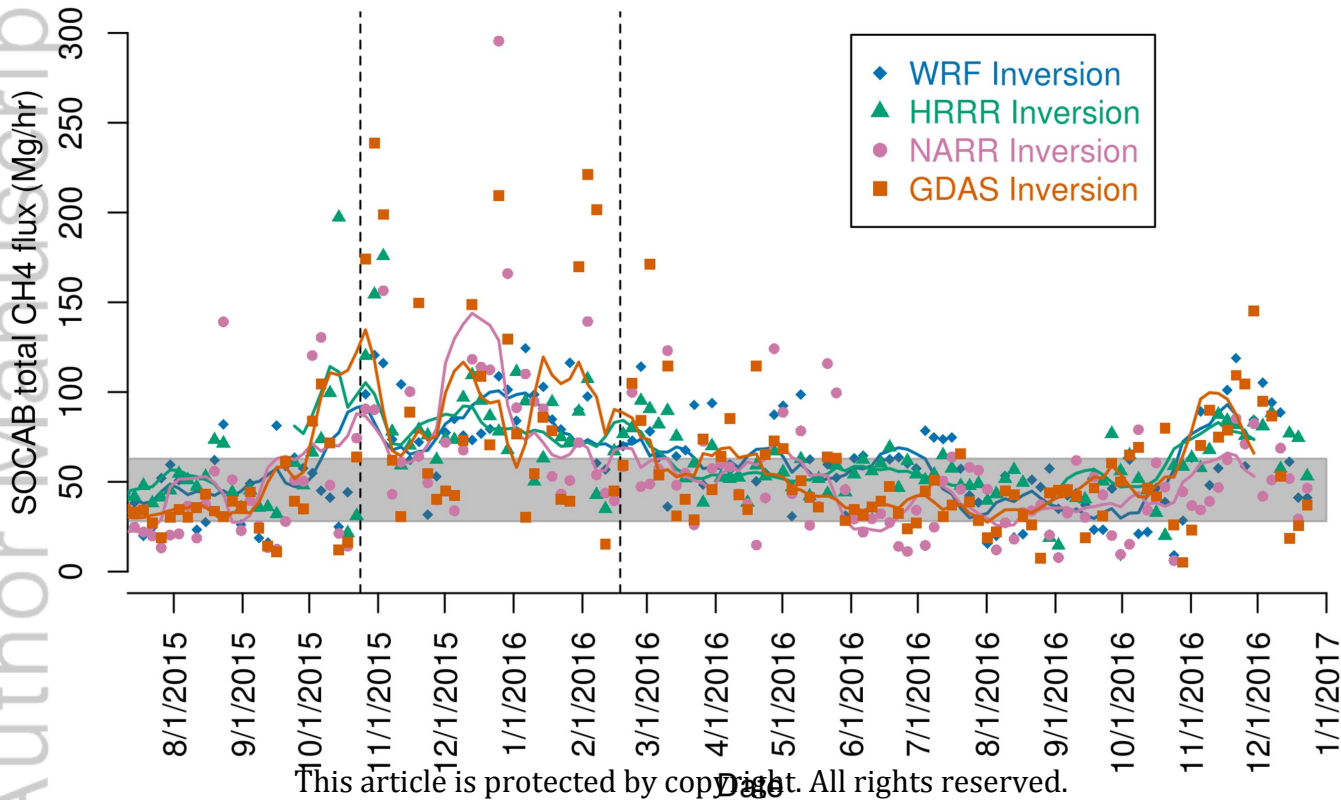


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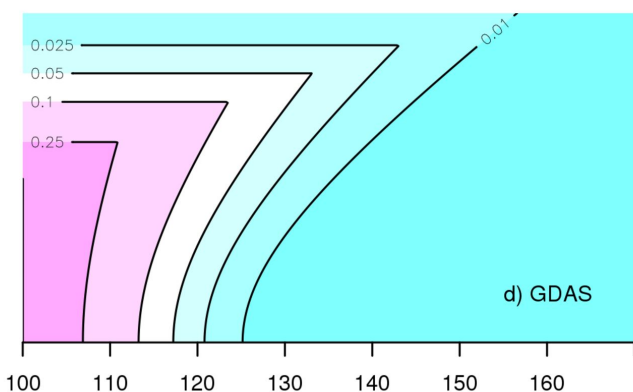
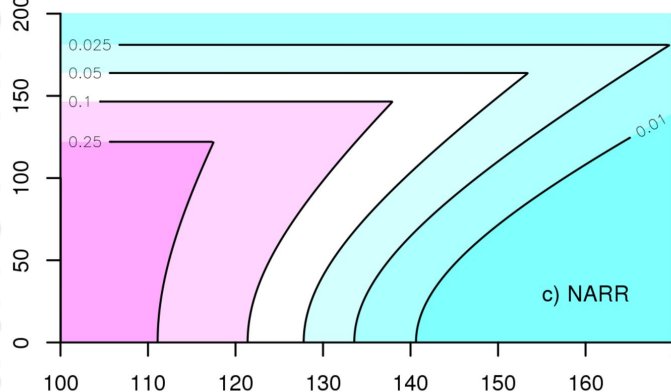
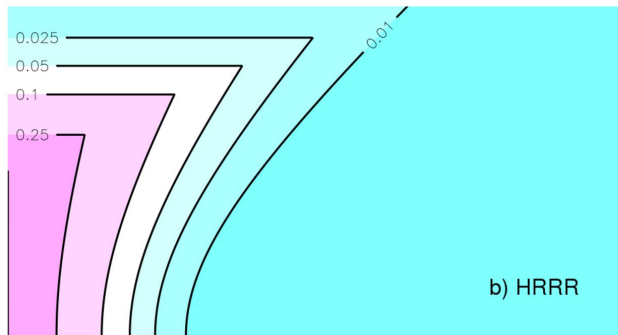
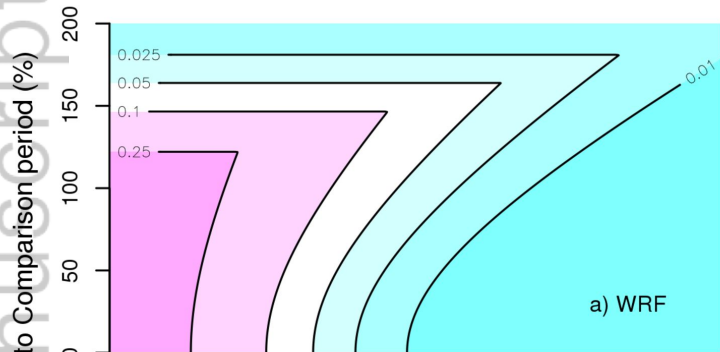
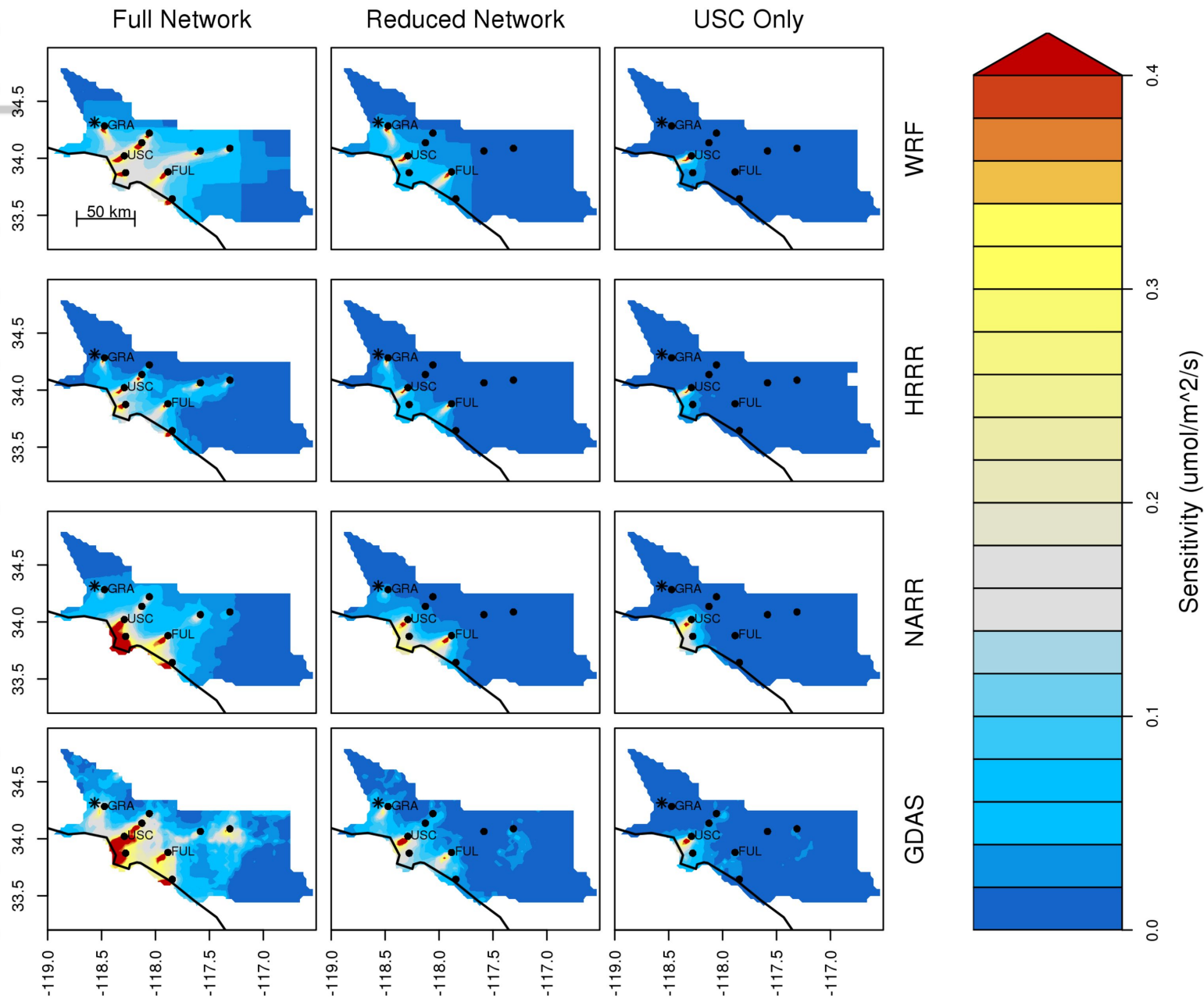
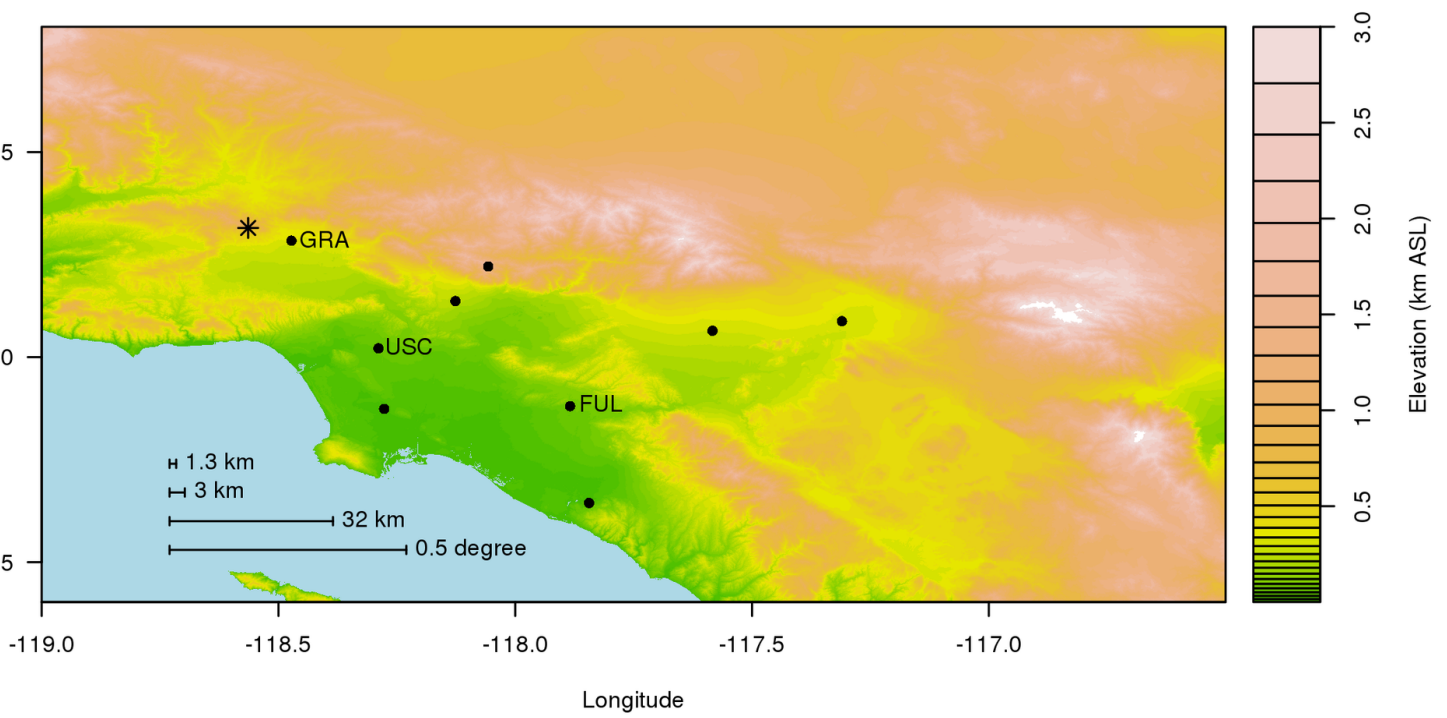


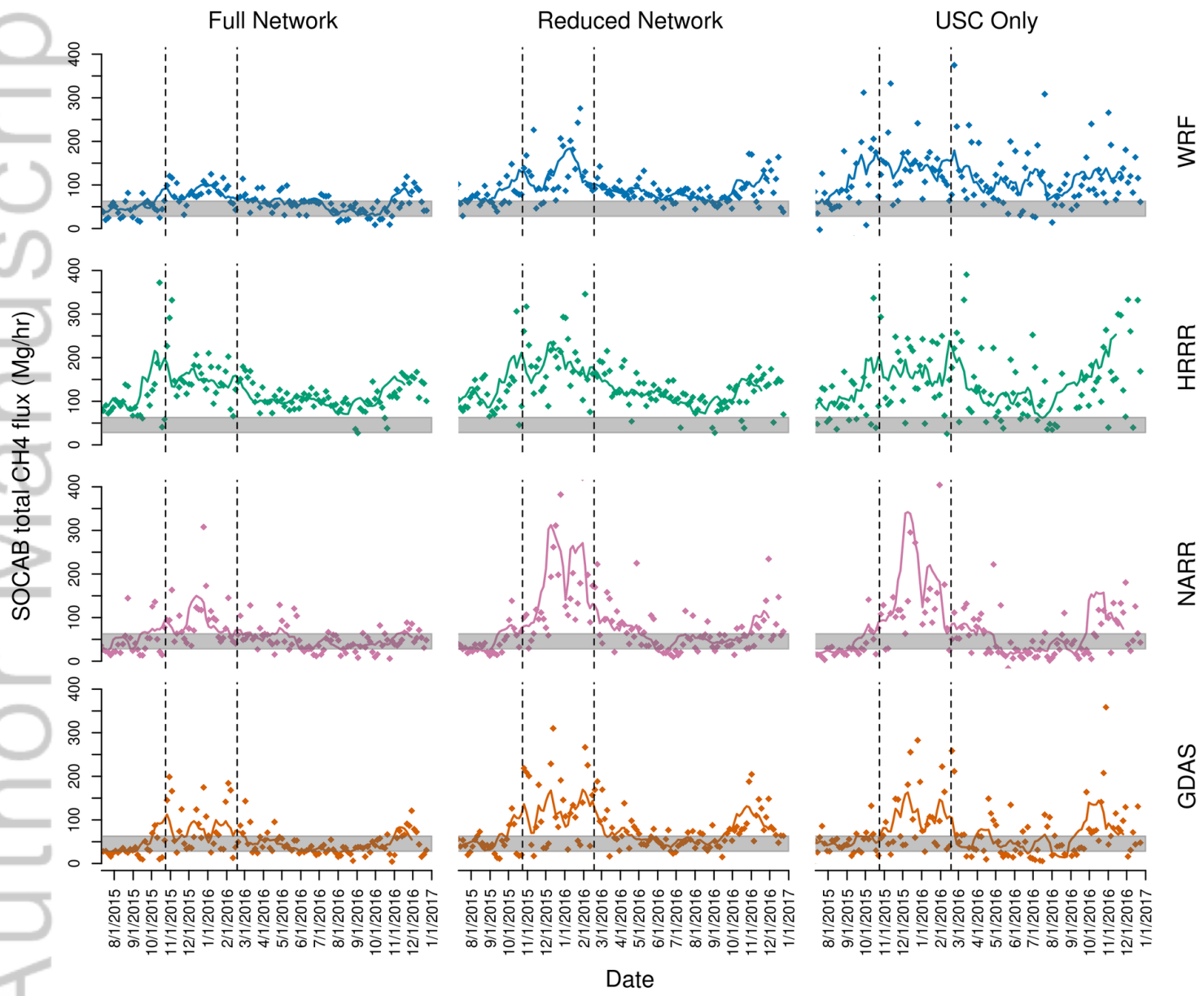
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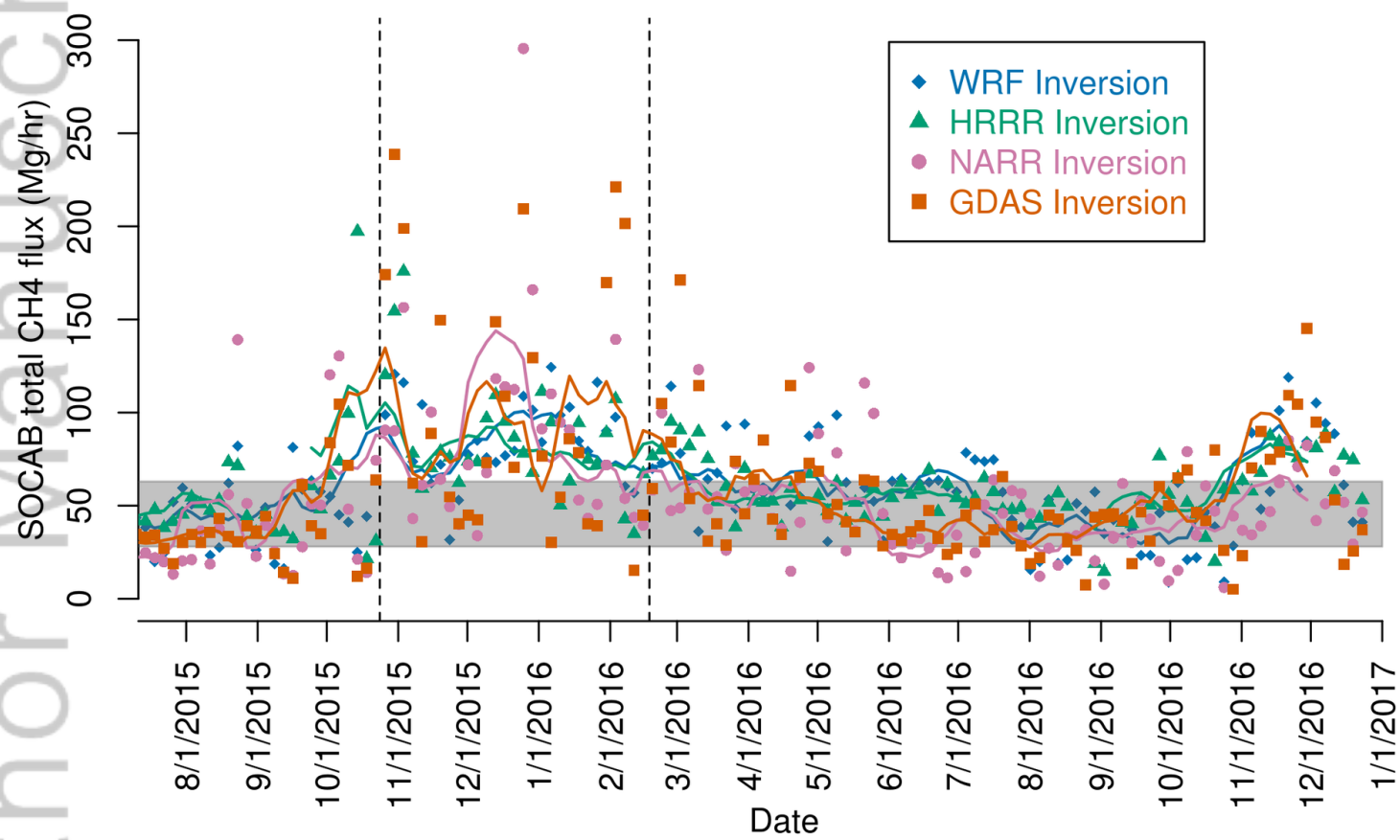




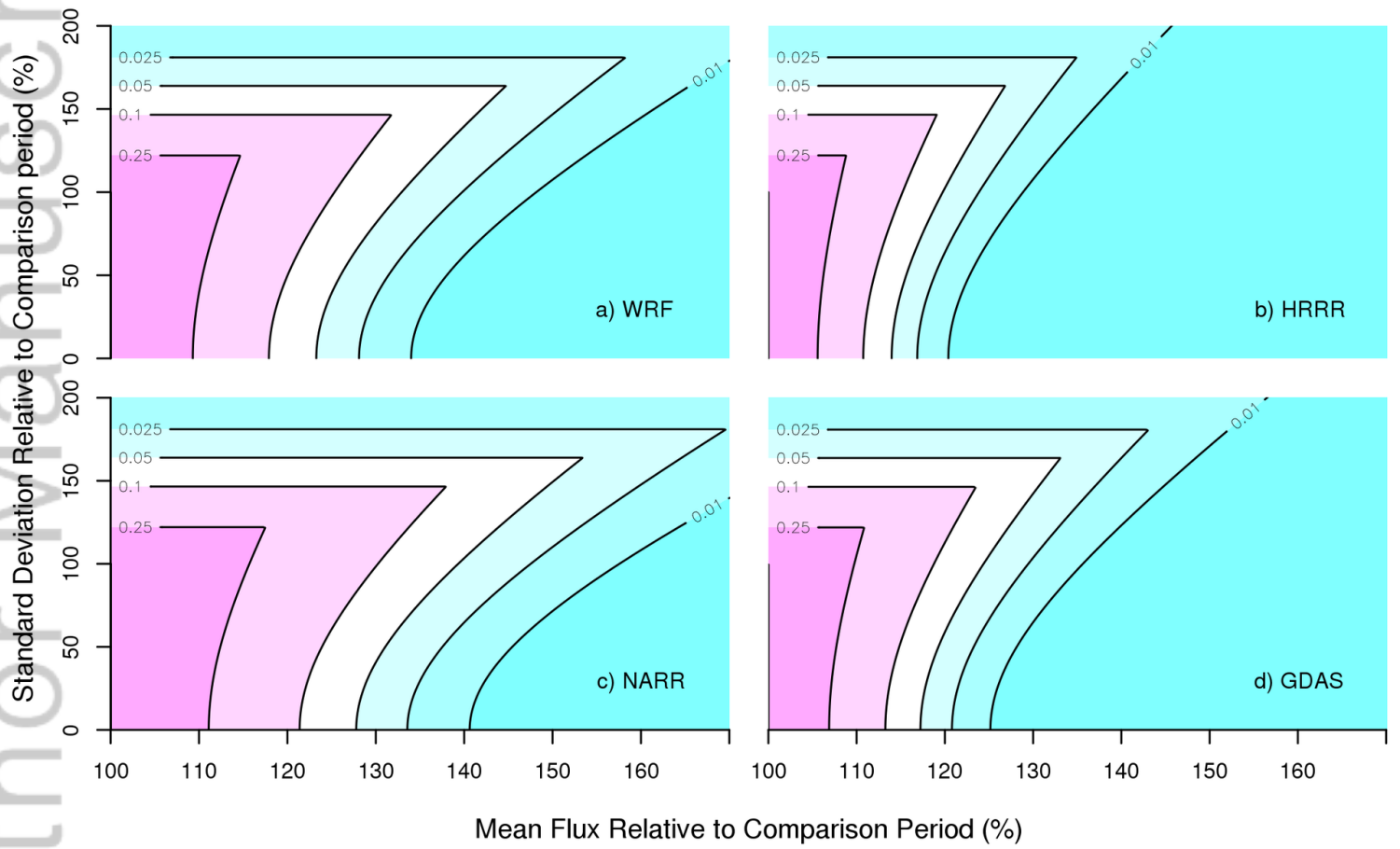
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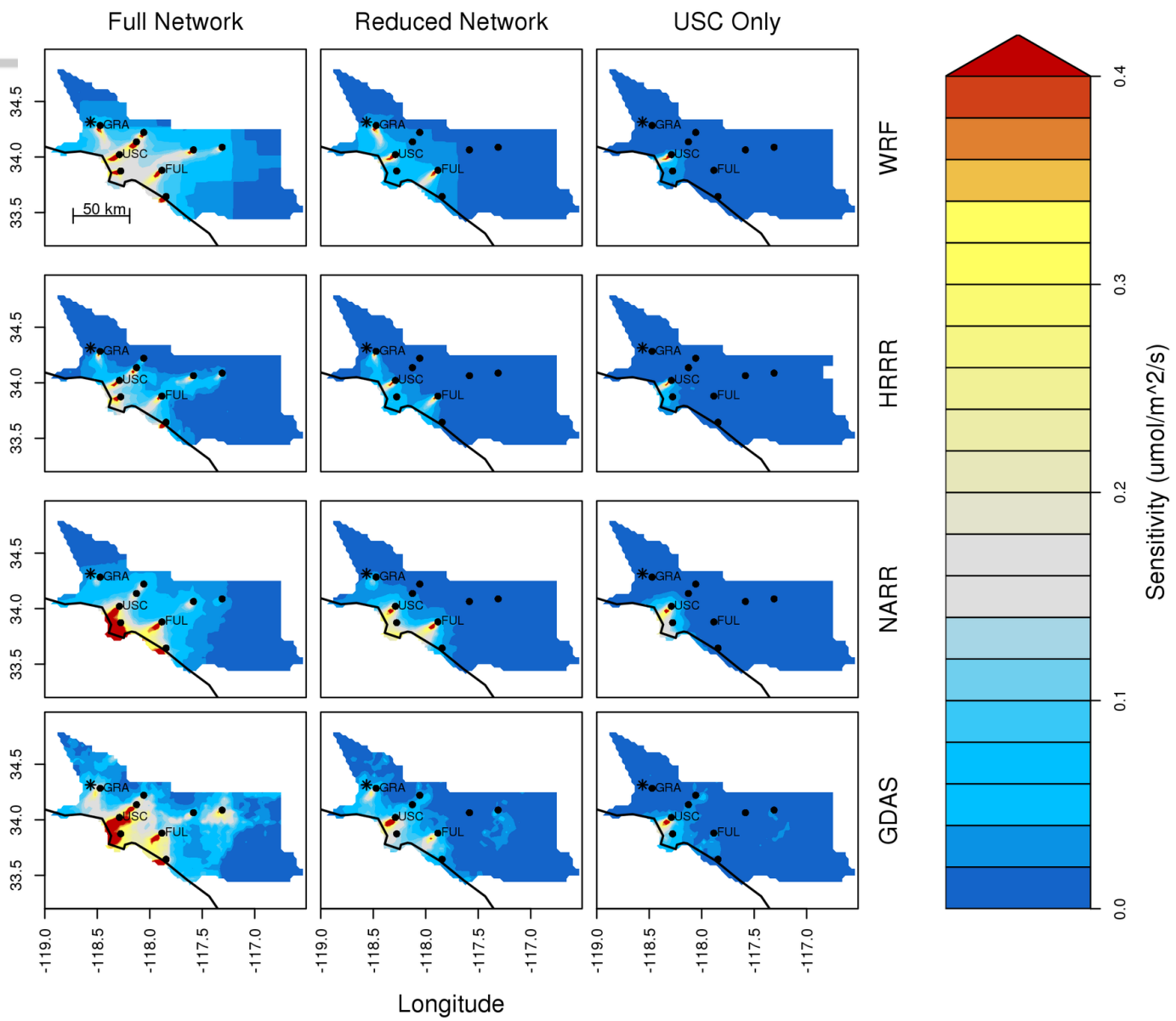
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