

1 Article Type : Original Article

2 Genome Evolution and Host Microbiome Shifts Correspond with Intraspecific Niche Divergence within  
3 Harmful Algal Bloom-Forming *Microcystis aeruginosa*.

4  
5 Short Title: Niche differentiation in *Microcystis aeruginosa*

6  
7 Sara L. Jackrel<sup>1\*</sup>, Jeffrey D. White<sup>2,3</sup>, Jacob T. Evans<sup>1</sup>, Kyle Buffin<sup>1</sup>, Kristen Hayden<sup>1</sup>, Orlando Sarnelle<sup>3</sup>,  
8 Vincent J. Deneff<sup>1</sup>

9  
10 <sup>1</sup>Department of Ecology and Evolutionary Biology, University of Michigan, Ann Arbor, MI 48109 USA

11 <sup>2</sup>Department of Biology, Framingham State University, Framingham, MA 01701, USA

12 <sup>3</sup>Department of Fisheries and Wildlife, Michigan State University, East Lansing, MI 48824, USA

13 \* Corresponding author

14  
15 **Abstract:** Intraspecific niche divergence is an important driver of species range, population abundance,  
16 and impacts on ecosystem functions. Genetic changes are the primary focus when studying intraspecific  
17 divergence; however, the role of ecological interactions, particularly host-microbiome symbioses, is  
18 receiving increased attention. The relative importance of these evolutionary and ecological mechanisms  
19 has seen only limited evaluation. To address this question, we used *Microcystis aeruginosa*, the globally-  
20 distributed cyanobacterium that dominates freshwater harmful algal blooms. These blooms have been  
21 increasing in occurrence and intensity worldwide, causing major economic and ecological damages. We  
22 evaluated 46 isolates of *M. aeruginosa* and their microbiomes, collected from 14 lakes in Michigan, USA  
23 that vary over 20-fold in phosphorus levels, the primary limiting nutrient in freshwater systems. Genomes  
24 of *M. aeruginosa* diverged along this phosphorus gradient in genomic architecture and protein functions.  
25 Fitness in low-phosphorus lakes corresponded with additional shifts within *M. aeruginosa* including  
26 genome-wide reductions in nitrogen use, an expansion of phosphorus assimilation genes, and an  
27 alternative life history strategy of non-clonal colony formation. In addition to host shifts, despite culturing  
28 in common garden conditions, host-microbiomes diverged along the gradient in taxonomy, but converged  
29 in function with evidence of metabolic interdependence between the host and its microbiome. Divergence  
30 corresponded with a physiological tradeoff between fitness in low-phosphorus environments and growth  
31 rate in phosphorus-rich conditions. Co-occurrence of genotypes adapted to different nutrient environments

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1111/MEC.15198](https://doi.org/10.1111/MEC.15198)

This article is protected by copyright. All rights reserved

32 in phosphorus-rich lakes may have critical implications for understanding how *M. aeruginosa* blooms  
33 persist after initial nutrient depletion. Ultimately, we demonstrate that the intertwined effects of genome  
34 evolution, host life history strategy, and ecological interactions between a host and its microbiome  
35 correspond with an intraspecific niche shift with important implications for whole ecosystem function.

36  
37 **Keywords:** genome evolution; adaptation; intraspecific variation; host-microbiome; nutrient limitation;  
38 cyanobacterial harmful algal blooms

39  
40 **Introduction:**

41 Variation within a species has been shown to rival the effects of among-species variation in  
42 regulating community structure (Crutsinger et al., 2006), trophic interactions (Post et al., 2008; Chislock et  
43 al. 2013), and nutrient cycling through ecosystems (Bassar et al., 2010). Therefore, predicting how these  
44 community- and ecosystem-level processes vary over time and space requires a mechanistic understanding  
45 of intraspecific divergence. Variation within a species that leads to niche divergence can occur as the  
46 result of evolutionary change, phenotypic plasticity, and even through altered ecological interactions with  
47 symbiotic partners (Lamichhaney et al., 2015; Lau & Lennon, 2012; Pfennig et al., 2010). As evolutionary  
48 and ecological mechanisms can occur over different time scales and their effects can have varying degrees  
49 of permanence, understanding the relative roles of these mechanisms towards shaping and maintaining  
50 intraspecific variation can elucidate what factors mediate stability of important ecosystem processes.

51 Genomic drivers of niche divergence between closely related organisms has been documented in  
52 several systems (Johnson et al. 2006; Lamichhaney et al., 2015). While selection upon standing genetic  
53 variation within a population can occur relatively rapidly (Barrett & Schluter, 2008), new mutations or  
54 lateral transferred genes rarely improve fitness (Drake et al., 1998; Vox et al., 2015), although examples of  
55 selective sweeps mediated by niche-determining genes acquired by lateral gene transfer have been  
56 documented (Shapiro and Polz, 2014). In contrast to evolutionary change, ecological mechanisms provide  
57 an alternative that may permit organisms to acclimate to divergent environmental conditions more rapidly.  
58 For example, the focal organism may take advantage of beneficial functionalities that pre-exist in other  
59 organisms. As such, microbial symbionts can contribute towards essential functions for the survival of  
60 their host in novel environments, such as those bacteria that facilitated the transition of the marine algae  
61 into freshwater environments (Dittami et al., 2016) and soil microbes that improved plant fitness under  
62 drought stress (Lau & Lennon, 2012).

63 To further our understanding of the mechanisms by which intraspecific niche divergence occurs,  
64 we simultaneously investigated the roles of 1) genomic changes within the host and, 2) shifts in

65 community membership and functionality of the host microbiome. We used the cyanobacterial  
66 phytoplankter *Microcystis aeruginosa* as our model to probe the roles of these two factors in driving  
67 fitness across a nutrient gradient from highly eutrophic, phosphorus-replete lakes to oligotrophic, low-  
68 phosphorus-lakes. This cyanobacterium is a useful model system as we can infer fitness of each of our *M.*  
69 *aeruginosa* isolates in their respective environments by collecting actively growing colonies, which are  
70 frequently composed of  $10^4$  to  $10^5$  cells (Costas et al., 2008). Understanding drivers of *M. aeruginosa*  
71 fitness across a nutrient gradient has major ecological, socioeconomic, and human health implications. As  
72 a dominant, bloom-forming cyanobacterium in nutrient-enriched freshwater systems world-wide, *M.*  
73 *aeruginosa* has produced concentrated levels of hepatotoxic microcystins that have caused mass wildlife  
74 mortalities (Masango et al. 2010) and drinking water crises, such as the 2007 bloom in Lake Taihu (Qin et  
75 al. 2010) and the 2014 bloom in Lake Erie (Steffen et al., 2017). Models predict that blooms of record-  
76 breaking intensity will become increasingly common due to cultural eutrophication and increased  
77 temperatures driven by climate change (Michalak et al., 2013).

78 As a consequence, eutrophic environments have been the focus of genome sequencing effort for  
79 *M. aeruginosa*. Their genomes are noted as having an unusually high percentage of long DNA repeats,  
80 insertions, transposable elements, and lateral gene transfers (Franguel et al., 2008; Kaneko et al., 2007;  
81 Meyer et al., 2017). Sequenced *M. aeruginosa* genomes from eutrophic habitats are highly variable, but  
82 this variation does not show clear biogeographic patterns (Meyer et al., 2017). In contrast, we previously  
83 identified a monophyletic ecotype of *M. aeruginosa* occurring in oligotrophic inland lakes in Michigan  
84 (Berry et al., 2017). Berry et al. (2017) constructed a phylogeny of *M. aeruginosa* isolates from inland  
85 lakes of Michigan using five housekeeping genes but did not investigate any further differences in the  
86 genomes or physiology of *M. aeruginosa* or host-associated microbiomes. Here, using this same dataset,  
87 we investigated the underlying mechanisms that may increase fitness of different ecotypes across a  
88 phosphorus gradient. We also evaluated genomes of heterotrophic bacteria residing in the *M. aeruginosa*  
89 phycosphere, defined as the nutrient-rich microenvironment immediately surrounding a phytoplankton cell  
90 where metabolites between the host and associated-bacteria are most readily exchanged (Seymour et al.  
91 2017). The phycosphere of *M. aeruginosa* has previously been described as harboring a diversity of  $\alpha$ - and  
92  $\beta$ -proteobacteria, as well as bacteroidetes (Cai et al., 2014; Louati et al., 2015).

93 To test for evolutionary mechanisms promoting fitness across a lake phosphorus gradient, we  
94 probed for genomic signatures of streamlining in *M. aeruginosa* as a means for improved efficiency in  
95 nutrient-limited environments (Giovannoni et al., 2014). We considered strategies that would improve  
96 assimilation and efficiency of both phosphorus and nitrogen because oligotrophic lakes are typically co-  
97 limited (Sterner, 2008). We also evaluated cyanobacterial genomes for functional shifts across the

98 phosphorus gradient, including gene loss and gain, as well as indicators of positive selection. In addition,  
99 by evaluating population-level genomic heterogeneity within our host genome assemblies, we were able to  
100 infer life history strategies of *M. aeruginosa* colony formation as clonal (by cell division from a single  
101 cell) or non-clonal (by cell adhesion) (Xiao et al., 2018). To test for ecological mechanisms promoting  
102 fitness across the phosphorus gradient, we assessed whether changes in community membership and  
103 functionality of the host microbiome were a function of the phosphorus environment, and examined the  
104 functional roles played by the ubiquitously-associated bacterium, *Phycosocius bacilliformis* (Tanabe et al.,  
105 2015). This bacterium has been previously detected in 35 of 39 blooms of *M. aeruginosa* sampled in Asia  
106 and Africa (Tanabe et al., 2015), and was found in association with each of our 46 isolates of *M.*  
107 *aeruginosa*. Lastly, we evaluated the interdependence between the functionality of the host genome and  
108 that of its microbiome. We demonstrated that intraspecific niche divergence in *M. aeruginosa* corresponds  
109 with a combination of altered life history strategy, genome evolution, and ecological mechanisms. Further,  
110 evolutionary change, such as genome-wide shifts and novel gene acquisitions in the host, interacts with  
111 ecological shifts in the host microbiome that combine to ultimately correspond with improved host fitness.  
112 Additionally, from a methodological perspective, we used our data to demonstrate the need for cautious  
113 interpretation of apparent gene loss and reduced genome size when using metagenome-assembled  
114 genomes.

115  
116

## 117 **Materials and Methods:**

### 118 *Isolate Collection and Maintenance:*

119 We collected colonies of *M. aeruginosa* from 14 inland lakes located throughout southern  
120 Michigan, USA in July 2011, August 2011, and August 2013 (see Table S1 and Fig. S2 for coordinates of  
121 each lake and a map of lake locations). Lakes were selected to span a large nutrient gradient, from  
122 oligotrophic to hyper-eutrophic, as determined by total phosphorus concentration (TP, a widely used index  
123 of lake productivity). Our TP concentrations spanned over an order of magnitude from ~8 – 200  $\mu\text{g/L}$ ,  
124 which encompasses the range of TP documented in over 82% of lakes in the Northeastern United States  
125 (Soranno et al. 2017). While we also measured lake  $\text{NH}_4^+$  and  $\text{NO}_3^-$ , neither nitrogen measure  
126 corresponded with TP (correlations not significant, both  $p > 0.10$ , see Table S1 for TP,  $\text{NH}_4^+$  and  $\text{NO}_3^-$   
127 measures per lake). We collected water from the mixed layer of each lake via two pooled casts of an  
128 integrating tube sampler (12 m x 2.5 cm inside diameter). A subset of the water sample was stored for  
129 measurement of lake total phosphorus using the standard molybdenum-blue colorimetric technique and  
130 long path length spectrophotometry following persulfate digestion of organic matter (Menzel & Corwin,

131 1965; Murphy & Riley, 1962). We used standard thresholds in TP for assigning trophic status with an  
132 oligotrophic-mesotrophic boundary of 10  $\mu\text{g/L}$  and a mesotrophic-eutrophic boundary of 30  $\mu\text{g/L}$  (Wetzel,  
133 2001). To confirm trophic status of these lakes was consistent over time, each lake was sampled at least  
134 three times during multiple years, except for Lake Lansing which was sampled twice. Longer-term data  
135 sets are also reported for Gull Lake, MSU Lake 2, and Little Long Lake (see Table S1 for mean, min, and  
136 max TP, years sampled, and number of observations). To isolate *M. aeruginosa* from water samples, we  
137 used a Leica MS5 dissecting scope at 16X to pipette individual colonies. All colonies isolated were  
138 distinctive in shape, rather than amorphous masses or loose aggregations of cells. To retain only closely-  
139 associated phycosphere bacteria inhabiting the mucilage of the *M. aeruginosa* colony, we washed  
140 individual colonies by pipetting each sequentially through a series of six well plates each containing sterile  
141 0.5x WC-S growth medium. However, we note that this washing step might not have eliminated all free-  
142 living bacteria. We then transferred colonies into 20 mL tubes of sterile 0.5x WC-S growth medium,  
143 which typically has a higher successful establishment rate of ~80% for inland lake *M. aeruginosa*  
144 compared to other mediums such as BG-11 (White et al., 2011; Wilson et al., 2005). We then maintained  
145 all successfully established isolates in 200 mL batch cultures of 0.5x WC-S medium, incubated isolates at  
146 23°C under a 12:12 h light:dark cycle of 80  $\mu\text{mol m}^{-2} \text{s}^{-1}$ , and on a monthly basis, transferred an inoculum  
147 of each culture to fresh, sterile, 0.5x WC-S medium.

148

#### 149 *16S rRNA gene and Metagenomic Sequencing:*

150 On November 14 -18, 2014, we trapped subsamples of each culture on 0.45  $\mu\text{m}$  nitrocellulose  
151 filters, froze filters immediately, and stored at -80°C until extraction. This pore size allowed smaller free-  
152 living bacteria to pass through the filter. However, we caution that while the washing of colonies at the  
153 start of cultivation and filtration steps should have reduced inclusion of free-living bacteria in sequencing,  
154 these steps would not eliminate them entirely. We then thawed and incubated filters in 100  $\mu\text{L}$  of Qiagen  
155 ATL tissue lysis buffer, 300  $\mu\text{L}$  Qiagen AL lysis buffer, and 30  $\mu\text{L}$  proteinase K for 1 h at 56°C on a  
156 rotisserie at maximum speed. We vortexed cells for 10 minutes to lyse, homogenized lysates with a  
157 Qiashredder column, and purified DNA from the filtrate using a DNeasy Blood and Tissue kit (Qiagen,  
158 Hilden, Germany).

159 We surveyed the phycosphere bacterial community from the extracted DNA of each culture by  
160 generating PCR amplicon of the V4 region of the 16S rRNA gene using 515f/806r primers (Bergmann et  
161 al., 2011). DNA amplicon were sequenced on a 2x250 Illumina MiSeq v2 run at the University of  
162 Michigan Medical School. Data were generated using RTA v1.17.28 and MCS v2.2.0 software. We also  
163 generated metagenomic data of the *M. aeruginosa* host and associated phycosphere bacteria on an

164 Illumina HiSeq 100 cycle 2 x 100 nt PE sequencing run at the University of Michigan Sequencing Core.  
165 Libraries were generated with a 500 nt insert size using an automated Apollo 324 library preparation  
166 system (Wafergen Biosystems, Fremont, CA). We aimed to obtain approximately equal coverage of *M.*  
167 *aeruginosa* across all metagenome samples by adjusting the proportions of each individual library to the  
168 pooled-libraries sample based on the relative abundance estimates of *M. aeruginosa* from our 16S  
169 amplicon data. Raw sequencing data files are available under SRA accession number PRJNA351875.

170  
171 *Sequencing Analyses:*

172 We analyzed 16S rRNA gene survey data, including quality control of raw reads, read alignment,  
173 taxonomy assignment, and OTU clustering at 97% sequence similarity, using the mothur v1.34.3 standard  
174 operating procedure (accessed March 13, 2016 at ) (Schloss et al., 2009, Kozich et al. 2013). We  
175 completed taxonomy assignment of sequences using the TaxAss pipeline, which classifies sequences to a  
176 smaller database of freshwater taxa (Newton et al., 2011) and the larger SILVA database (Quast et al.,  
177 2012; Wang et al., 2007). Based on methods recommended by McMurdie and Homes (2014), read depth  
178 was normalized to depth of the smallest sample (n = 10,090 reads, see Table S2 for original and scaled  
179 read depths for each sample) using custom scripts that can be found here: . Raw metagenome reads were  
180 trimmed of adapters using Scythe and quality trimmed using Sickle with default parameters (Joshi & Fass,  
181 2011). Sequence quality was assessed before and after quality filtering using FastQC. We ran these quality  
182 control steps with a composite bash script that can be found here: [https://github.com/Geo-](https://github.com/Geomics/scripts/blob/master/wrappers/Assembly/qc.sh)  
183 [omics/scripts/blob/master/wrappers/Assembly/qc.sh](https://github.com/Geomics/scripts/blob/master/wrappers/Assembly/qc.sh). Sequencing reads were then assembled using idba-ud  
184 with the following parameters (--mink 50, --maxk 92, --step 4 or 6, --min\_contig 500) (Anantharaman et  
185 al., 2014). Metagenomic assemblies were first visualized with ESOM (Dick et al., 2009). We identified  
186 bins for both the target organism, *M. aeruginosa*, and an abundant phycosphere bacterium, *Phycosocius*  
187 *bacilliformis*. We used the default protocol for ESOM (Emergent Self-Organizing Maps), which is a  
188 binning approach that takes advantage of taxon-specific genomic signatures that arise due to genome-  
189 specific biases in codon usage. All sequences from our 46 metagenomic assemblies were trimmed into  
190 sequences of 10 kb in length and then imported into ESOM for assessment of tetranucleotide frequency of  
191 each contig. Each contig was plotted as a dot on the map using an unsupervised clustering algorithm to  
192 minimize distances between contigs sharing similar tetranucleotide frequency. Clusters of sequences were  
193 then manually selected and extracted as a bin. We identified high quality *P. bacilliformis* genomes using  
194 drep (Olm et al., 2017), compared similarity among these genomes using compareM (), and then retained  
195 only those genotypes at least 0.70% divergent from all other genotypes for further analysis. We uploaded  
196 1) complete metagenome assemblies, 2) isolated *M. aeruginosa* bins, and 3) *Phycosocius bacilliformis*

197 bins, all with a 4kb contig length cutoff, into the Joint Genome Institute Integrated Microbial Genomes  
198 database. We passed these sequencing data through standard analysis pipelines for assigning protein  
199 families (Huntemann et al., 2015). Analyses of these *M. aeruginosa* bins indicated our contig length cutoff  
200 was likely too stringent, because core photosynthesis genes, among others, often resided on contigs of  
201 approximately 3kb in length. We therefore repeated binning of *M. aeruginosa* at a lower 2 kb length  
202 threshold using VizBin (Laczny et al., 2015). Coverage of sequences within this bin was then calculated  
203 using bwa (Li & Durbin, 2009). Histograms of contig frequency versus coverage were then used to  
204 visualize coverage distributions for each sample. We then discarded all contigs below the main coverage  
205 distribution, which varied depending on sample.

206 *Microcystis aeruginosa* genomes at this 2 kb length cutoff were then re-annotated using our  
207 custom pipeline. We quantified single nucleotide polymorphisms, inserts, and deletions within each  
208 genome using samtools (Li et al., 2009). We also identified paralogs as any reciprocal hits within a  
209 genome below an e-value of 0.10 using the diamond protein alignment software (Buchfink et al., 2014).  
210 We identified sigma factors as genes assigned to any of 27 different protein families that contained the  
211 keyword 'sigma' within either the pfam name or pfam summary. We report sigma factors as a percentage  
212 of all genes in the genome. Given the tendency for some genes to occur on even smaller contigs,  
213 especially in the low nutrient group of genomes, we added all called genes on scaffolds below 2 kb that  
214 were assigned as *M. aeruginosa* using the USEARCH based Phylogenetic Distribution tool in the JGI  
215 IMG Standard Operating Procedure (Huntemann et al., 2015). We used these final bins with contigs of all  
216 lengths for determining GC content, genome size, percent completeness and percent contamination using  
217 checkM (Parks et al., 2015). We determined which genes are considered 'core' to the cyanobacterial  
218 phylum using checkM, which estimates genome completeness by referencing sets of marker genes that are  
219 specific to the inferred lineage of a genome within a reference genome phylogeny. We also used these  
220 final bins with contigs of all lengths for determining pairwise average nucleotide identity (ANI) between  
221 genomes using PYANI (Pritchard et al. 2016), where the boundary for prokaryotic species is generally  
222 accepted as ~ 95 – 96% ANI (Richter and Rossello-Mora, 2009). We used our custom pipeline to assign  
223 protein families to each gene residing on these < 2 kb contigs and added these protein family assignments  
224 to our primary annotation of all *M. aeruginosa* 2 kb+ bins. We used this approach to be conservative when  
225 determining if any gene functions were completely absent from a genome. However, because it becomes  
226 more challenging to call genes on such short fragments, we report all gene percentage data, including %  
227 coding DNA, % sigma factors, and % paralogs, using the 2 kb+ bins. However, we provide a comparison  
228 of all genome metrics with and without scaffolds below 2 kb in supplementary materials.

229 We constructed a multilocus sequence typing phylogeny using our 46 isolates collected from  
230 inland lakes in Michigan, as well as additional *M. aeruginosa* genotypes that had been collected from  
231 multiple locations across six continents. In addition to our 46 isolates first reported in Berry et al. (2017),  
232 genomes obtained from NCBI included 12 genomes referenced in Humbert et al. (2013) (), 8 genomes  
233 referenced in Meyer et al., (2017) and a *Synechococcus* outgroup (CP000097.1). We used gene sequences  
234 from five housekeeping genes (*pgi*, *gltX*, *ftsZ*, *glnA*, *gyrB*) that were obtained from *M. aeruginosa* strain  
235 NIES483. As in our previous analysis (Berry et al., 2017), we searched for gene orthologs in the  
236 metagenomic data of each of the Michigan inland lake isolate and the genomes obtained from NCBI using  
237 a custom ruby script available on this project's github page ([https://github.com/DenefLab/microcystis-](https://github.com/DenefLab/microcystis-oligotypes)  
238 [oligotypes](https://github.com/DenefLab/microcystis-oligotypes)). Extracted gene sequences were concatenated and aligned with MUSCLE using default  
239 parameters (Edgar et al., 2004). A phylogeny was constructed using RAxML v8.2.8 with a *Synechococcus*  
240 outgroup (Berry et al., 2017; Stamatakis, 2006), and a Newick tree was visualized using FigTree v1.4.3  
241 software (Rambaut, 2012).

242 For a subset of isolates collected from a relatively uniform environment (either Gull or  
243 Wintergreen Lakes on August 8, 2013), we measured the amount of genome variation among isolates with  
244 a pangenome analysis. We estimated the size of the core genome shared across isolates versus isolate-level  
245 variation by finding homologous gene clusters using the GET\_HOMOLOGUES software (Contreras-  
246 Moreira & Vinuesa, 2013). We used only a consensus subset of gene clusters identified by both  
247 OrthoMCL Markov Cluster and COGtriangle algorithms.

248 To detect positive selection in the two main branches that divide all 46 inland lake isolates of *M.*  
249 *aeruginosa*, we calculate genome-wide synonymous to non-synonymous substitution rate ratios with the  
250 posigene software package (Sahm et al., 2017). For this analysis, we supplied a smaller-scale phylogenetic  
251 tree with only eight genomes representative of the oligotrophic and eutrophic/mesotrophic branches from  
252 Berry et al. (2017) (K13-10, G13-05, G13-09, LL11-07, LG13-11, F13-15, K13-06, K13-05). We ran  
253 analyses separately for each branch and identified orthologs against the largest genome within each  
254 respective branch as the anchor species (i.e., G13-05 and K13-10).

255  
256 *Growth Assays:*

257 We aimed to detect whether growth rate of *M. aeruginosa* varies across populations according to  
258 phylogenetic group and a total phosphorus gradient. To assess variation across phylogenetic groups, we  
259 measured growth rates of 19 strains from our culture collection that had been isolated in 2011. To further  
260 validate our results of different growth rates across a total phosphorus gradient, we added an additional 12



261 strains that were assayed by Wilson et al. (2006). In total, these growth rate data represent 31 strains  
262 originating from 22 lakes throughout lower Michigan.

263 Common-garden growth assays to detect variation among phylogenetic groups of *M. aeruginosa*  
264 were conducted with the general design as follows. Fresh 20 mL cultures of isolates were initiated 7 days  
265 prior to an assay to ensure that *M. aeruginosa* was exponentially growing. One colony from each of 19  
266 strains of *Microcystis* were then inoculated via pipette (1  $\mu$ L) into randomized, separate wells containing  
267 0.5 mL sterile 0.5 $\times$  WC-S medium within 8-well chambered slides (Nunc Lab-Tek II Chamber Slide  
268 System) (Wilson et al., 2010). Once inoculated (day 0), colonies were photographed every 2 days for 6  
269 days at 100 $\times$  using a light microscope (Nikon Eclipse E600) interfaced with a digital camera (Diagnostic  
270 Instruments). See supplementary Fig. S3 for a sequence of digital micrographs depicting growth of a *M.*  
271 *aeruginosa* colony during a 6-day growth assay. Measurements, added to the images with computer  
272 software (Spot Advanced, Diagnostic Instruments), were made of colony surface area and depth (the  
273 straight line length perpendicular to the greatest linear dimension); colony volume ( $\mu\text{m}^3$ ) was determined  
274 as the product of surface area and depth (Wilson et al. 2010). Growth rate was determined as the slope of  
275 the linear regression of natural logarithm-transformed colony volumes over time.

276 Since coloniality is a characteristic trait of *M. aeruginosa* in nature (Wehr & Sheath, 2003), all  
277 growth assays were performed using colonial isolates that had been in culture for less than  $\sim$ 1.5 years.  
278 This contrasts with many previous lab studies of *M. aeruginosa* that have utilized older, single-celled  
279 culture collection strains. Furthermore, since all isolates employed in the experiments were the same age  
280 and were recently isolated, concerns arising from evolution in culture were minimized (Burkholder &  
281 Glibert, 2009; Demott & Mckinney, 2015; Lakeman et al., 2009). Assaying growth of individual *M.*  
282 *aeruginosa* colonies is necessary and advantageous because, unlike batch culture assays, this permits  
283 controlling for colony size and inoculation density effects on growth rate, since small colonies grow faster  
284 than large colonies (Wilson et al., 2010). To further minimize the confounding effects of initial colony size  
285 and shape on growth rate, round colonies of approximately the same equivalent diameter were selected for  
286 each isolate to the fullest extent possible using an ocular micrometer.

287  
288 *Statistical Analyses:*

289 We tested whether the phylogenetic grouping and lake of origin of *M. aeruginosa* isolates was  
290 predictive of growth rate, genome size, genome completeness and contamination, GC content, number and  
291 length of contigs, percentage of coding DNA, and number of sigma factors, paralogs, and SNPs with  
292 linear-mixed effects models using the lmer {lme4} function in R. To account for multiple *M. aeruginosa*  
293 isolates originating from the same lake, we used random effects terms for lake and collection date. All

294 percentage data was arcsine-square-root transformed to meet model assumptions. We also tested for a  
295 correlation between growth rate and total phosphorus concentration using linear regression. To test  
296 whether there was a correlation within the LL/LG group between genome completeness and percentage of  
297 polymorphic sites, we used robust linear regression (rlm {MASS} function in R) because it remains robust  
298 despite the presence of outliers or highly influential data points.

299 We next measured functional similarities among isolates of *M. aeruginosa* depending on lake  
300 origin and phylogenetic grouping. We calculated isolate dissimilarity using a Bray-Curtis distance metric  
301 on a matrix consisting of the number of genes within each of 1,820 different protein families. We excluded  
302 protein families that had zero variance across our 46 isolates. We tested whether isolates differed  
303 significantly in genome function by lake origin and phylogenetic grouping using analysis of variance of  
304 distance matrices with adonis {vegan}, which is a version of PERMANOVA that can accept categorical  
305 and continuous variables. We visualized clustering among isolates using principal coordinate analysis with  
306 the pcoa {ape} function in R. We also repeated these dissimilarity analyses with the bacterial phycosphere  
307 community, using both a Bray-Curtis distance metric of OTU community composition and a functional  
308 distance metric using protein families as done with host isolates. However, within the protein family  
309 phycosphere distance matrix, many pfams were not shared across samples (i.e., high beta-diversity or  
310 turnover). Such data structure often results in a strong arch or horseshoe shape in an ordination that is  
311 indicative of non-independent axes and is known as the Guttman effect. To facilitate data interpretation,  
312 we minimized the arching effect by using the stepacross flexible shortest path correction with a 'toolong'  
313 parameter of 0.75 {vegan} (Smith, 2017). Further, we then used these Bray-Curtis distances from the host  
314 and phycosphere data matrices to test whether host genomes that were more similar in function harbored  
315 phycosphere communities that were more similar in function. We assessed this association with a linear  
316 mixed-effects model using the lmer function in R and lake as the random effects term.

317 To determine whether certain protein families or Kegg Orthology terms within the host genome or  
318 phycosphere metagenome were associated with different phylogenetic groupings of *M. aeruginosa*, we ran  
319 analysis of variance models with a Benjamini-Hochberg false discovery rate correction in STAMP  
320 (Benjamini & Hochberg, 1995; Parks et al., 2014). To account for multiple samples collected within the  
321 same lake, rather than input gene counts for each host genome or phycosphere metagenome into these  
322 models, we instead averaged gene counts for each protein family across all genomes/metagenomes that  
323 both belonged to the same phylogenetic grouping and were collected from the same lake.

324

325 **Results:**

326

327 Structural Genome Variation within *M. aeruginosa*

328

329 Previously, we found two distinct phylogenetic groups of *M. aeruginosa* isolates from inland lakes  
330 of Michigan using a multi-locus sequence typing analysis (Berry et al., 2017). Here, we expand this  
331 analysis to include all publically available sequences of *M. aeruginosa* originating from multiple locations  
332 spanning six continents. We observed that our phylogenetic clustering of the 46 inland lake isolates  
333 corresponded with the trophic status of lake of origin, as determined by lake total phosphorus (Table S1).  
334 We note that although phosphorus and nitrogen can be co-limiting in lakes, we found no evidence that our  
335 phylogenetic clustering corresponded with either  $\text{NH}_4^+$  or  $\text{NO}_3^-$  concentrations (Table S1), and therefore  
336 use the classic, phosphorus-based definition for assigning freshwater trophic status (see methods). We  
337 refer to these isolates in three categories that correspond to the trophic status of the lake of origin on the  
338 one hand, and this phylogenetic clustering on the other. Nineteen of the 20 NCBI genomes clustered with  
339 17 isolates that originated from eutrophic and mesotrophic Michigan lakes, which we refer to as the ‘High  
340 Phosphorus Lake/High Phosphorus Genotype’ (HL/HG) group (Fig. S1). A second group contained an  
341 additional 29 isolates originating from oligotrophic, mesotrophic, and eutrophic Michigan lakes. Here, we  
342 subdivided this group into 18 isolates collected from oligotrophic lakes, which we refer to as the ‘Low  
343 Phosphorus Lake/Low Phosphorus Genotype’ (LL/LG) group, and into 11 isolates collected from  
344 eutrophic and mesotrophic lakes, which we refer to as the ‘High Phosphorus Lake/Low Phosphorus  
345 Genotype’ (HL/LG) group. Genomes within the LL/LG and HL/LG groups were less variable compared to  
346 other genomes within their respective groups (pairwise average nucleotide identity of strains within the  
347 group,  $\mu = 98.5 \pm 0.035$  s.e. and  $98.4 \pm 0.082$ , respectively), while genomes within the HL/HG group were  
348 notably more variable ( $95.6 \pm 0.09$ ). Based on average nucleotide identity of genomes belonging to  
349 different groups, the LL/LG and HL/LG groups were most similar ( $98.1 \pm 0.026$ ) while HL/HG was more  
350 distant to LL/LG ( $95.2 \pm 0.030$ ) and HL/LG ( $95.3 \pm 0.037$ ).

351 Metagenomic assemblies of *M. aeruginosa* from the LL/LG group, and to a lesser extent, HL/LG,  
352 were more fragmented than the HL/HG group (Fig. S4;  $F_{2,46} = 59.0$ ,  $p < 0.001$ ; # contigs in LL/LG  
353 genomes:  $\mu = 632 \pm 72$  s.d., HL/LG:  $\mu = 548 \pm 31$  s.d., HL/HG:  $\mu = 318 \pm 77$  s.d.). As assembly  
354 fragmentation is often caused by increased genomic heterogeneity, we determined the number of  
355 polymorphic sites in all assemblies. LL/LG isolates contained a higher percentage of polymorphic sites in  
356 their genomes compared to the other groups (Fig. 1A;  $F_{2,46} = 15.4$ ,  $p < 0.001$ ), and within these LL/LG  
357 isolates, higher percentages of polymorphic sites correlated with both lower estimated genome  
358 completeness (Fig. 1B; robust linear regression:  $F_{1,16} = 5.7$ ,  $p = 0.029$ ) and a greater number of fragments  
359 ( $F_{1,16} = 14.1$ ,  $p = 0.0018$ ,  $R^2 = 0.43$ ). LL/LG assemblies also contained a higher proportion of low sequence

360 read coverage contigs (Fig. 1C, and Fig. S5). As further evidence that these low coverage contigs were the  
361 result of exceptionally high heterogeneity causing separate contigs to be generated, we found that genomes  
362 with many low coverage contigs had a greater median Reference:Alternate allele ratio at polymorphic sites  
363 located throughout the genome. This indicates that high heterogeneity within a sequence caused the  
364 assembly to divide divergent groups of reads into two separate contigs, which would decrease occurrence  
365 of within-contig heterogeneity (Fig. 1D, Fig. S6). Further, in the 7 LL/LG genomes with many low  
366 coverage contigs (shown in Fig. S6), these contigs shared 100% average nucleotide identity to contigs of  
367 higher coverage within the same genome. As higher fragmentation leads to smaller average contig size,  
368 the disparate levels of heterogeneity and assembly fragmentation between phylogenetic groups could lead  
369 to inaccurate comparisons in structural or functional genome variation across groups. Therefore, in  
370 addition to contigs 2 kb in length and longer included in our initial binning analysis, we added any contigs  
371 that were less than 2 kb in length but annotated as *M. aeruginosa*. This approach had minimal effects for  
372 the HL/HG or HL/LG groups, but added numerous genes that had initially appeared absent to the LL/LG  
373 group (Fig. S4, compares genome metrics with and without these shorter fragments; Fig. S7 lists core  
374 genes occurring on short fragments in LL/LG genomes). Unless noted otherwise, all subsequent genome  
375 descriptions use these modified bins that included *M. aeruginosa*-annotated fragments less than 2 kb.

376 Genomes from the LL/LG diverged subtly, though significantly, from the HL/HG group in several  
377 characteristics commonly assessed to detect genome streamlining. Further, the HL/LG group fell at  
378 intermediate levels in these streamlining characteristics, where HL/LG was more similar to LL/LG for  
379 certain characteristics and more similar to HL/HG for others. Genomes across groups were similar in size  
380 ( $p = 0.13$ ), however LL/LG genomes were less complete than those in HL/HG (Fig. 2;  $F_{2,46} = 17.40$ ,  $p <$   
381  $0.001$ ; LL/LG:  $\mu = 96.77\% \pm 1.49$  s.d. versus HL/HG:  $\mu = 99.27\% \pm 0.56$  s.d., and at intermediate  
382 completeness HL/LG:  $\mu = 98.91 \pm 0.59$  s.d.). Eleven genes considered core to the cyanobacterial phylum,  
383 which includes the *Microcystis* genus, were absent from multiple LL/LG and HL/LG isolates (Fig. S7).  
384 LL/LG isolates frequently lacked core genes for the enzyme 3-dehydroquinate dehydratase in the  
385 shikimate pathway, which blocks the biosynthesis of the aromatic amino acids (i.e., tryptophan, tyrosine,  
386 and phenylalanine), and the enzyme acetolactate synthase, which inhibits synthesis of the branch-chain  
387 amino acids (i.e., valine, leucine, and isoleucine). We confirmed these core genes were absent on even  
388 very small contigs (i.e.,  $< 2$  kb) by searching scaffolds below 2 kb in length (see Fig. S7 for a comparison  
389 of core genes inferred absent before and after inclusion of these shorter fragments). Additionally, LL/LG  
390 and HL/LG isolates also had greater nitrogen-use efficiency via a slightly lower GC content (Fig. 2;  $F_{2,46} =$   
391  $7.71$ ,  $p = 0.021$ ; LL/LG genomes:  $\mu = 42.48\% \pm 0.049$  s.d., HL/LG genomes:  $\mu = 42.51\% \pm 0.075$  s.d.,  
392 compared to HL/HG genomes:  $\mu = 42.71 \pm 0.15$ ). LL/LG and HL/LG genomes also contained a higher

393 percentage of coding versus non-coding DNA, while HL/HG was notably more variable (Fig. 2;  $F_{2,46} =$   
394  $5.00$ ,  $p = 0.082$ ; LL/LG genomes:  $\mu = 81.40\% \pm 0.17$  s.d. and HL/LG genomes:  $\mu = 81.47\% \pm 0.26$  s.d.  
395 compared to HL/HG genomes:  $\mu = 81.13\% \pm 0.69$  s.d.). Streamlined genomes often contain fewer  
396 paralogs, however we found similar percentages across groups (Fig. 2;  $F_{2,46} = 4.81$ ,  $p = 0.09$ ). Lastly,  
397 HL/LG genomes contained significantly more sigma factors than HL/HG genomes (Fig. 2;  $F_{2,46} = 8.78$ ,  $p$   
398  $= 0.012$ ; HL/LG genomes:  $\mu = 0.291\% \pm 0.01$  s.d. compared to HL/HG genomes:  $\mu = 0.266\% \pm 0.03$  s.d.,  
399 while LL/LG genomes were not significantly different from either group:  $\mu = 0.281\% \pm 0.02$  s.d.). Note  
400 that metrics requiring gene identification, including % coding DNA, sigma factors, and paralogs, were  
401 inferred using only the 2 kb+ bins due to low accuracy of gene calling on shorter fragments.

402

#### 403 Functional Genome Variation within *M. aeruginosa*

404

405 Isolates of *M. aeruginosa* differed in their genome-wide predicted gene functions based on protein  
406 families across the ten different lakes from which strains had originated (Note: four lakes were omitted as  
407 only one isolate per lake was sequenced; Fig. 3A; PCoA on a Bray-Curtis dissimilarity of genes  
408 categorized by protein family; adonis  $F_{9,40} = 8.56$ ,  $p < 0.01$ ,  $R^2 = 0.311$ ). Further, the three phylogenetic  
409 groups of *M. aeruginosa* were functionally distinct (Fig. 3B; adonis  $F_{2,45} = 9.01$ ,  $p < 0.01$ ,  $R^2 = 0.295$ ).  
410 LL/LG and HL/LG genomes were more similar to each other than either was to HL/HG genomes.  
411 Congruent with our phylogenetic results, HL/HG showed more variability among isolates (Fig. 3B).  
412 Group clustering was driven in part by the frequent absence of 11 core genes from LL/LG and HL/LG  
413 genomes (Fig. S7), where core genes were determined via checkM (Parks et al., 2015). Gene abundances  
414 within 671 protein families also varied significantly across phylogenetic groups (Fig. S8; FDR-corrected  
415  $p$ -values  $< 0.05$ ). Focusing our analysis on nutrient uptake and metabolism, we found that 16 LL/LG  
416 isolates, originating from 3 of 4 oligotrophic lakes, contained the alkaline phosphatase *phoA* (pfam00245).  
417 In contrast, none of the 28 isolates originating from phosphorus-rich lakes contained the alkaline  
418 phosphatase *phoA*. Further, while all isolates contained at least one gene within pfam05787 annotated as  
419 alkaline phosphatase *phoX*, all 29 LL/LG and HL/LG genomes contained a second gene within this  
420 protein family, while only 1 of the 17 HL/HG isolates contained a second *phoX* gene. Also pertinent given  
421 that phosphorus-limitation in oligotrophic lakes often co-occurs with nitrogen limitation, we found that  
422 LL/LG and HL/LG genomes also contain additional genes for nitrate and nitrite transport (K15576-8; Fig.  
423 S8). Additionally, a *nifU*-like protein involved in iron binding and FeS cluster formation occurred in 8 of  
424 18 LL/LG isolates across three different oligotrophic lakes but none occurred in the 27 HL/HG or HL/LG  
425 isolates. Further, considering the importance of colonial growth, we highlight the occurrence of several

426 genes regulating cell-cell recognition and adhesion in only LL/LG isolates. These genes include NeuB or  
427 sialic acid synthase, sialidases, an FRG1-like domain involved in underwater adhesion, and a gene in  
428 pfam03865 that is involved in secretion of adhesins.

429 Further, each phylogenetic group had different genes under positive selection. Genes under  
430 positive selection for the phylogenetic branch containing LL/LG isolates included a hybrid sensor histidine  
431 kinase/response regulator, chromosome segregation protein SMC, DNA-directed RNA polymerase  
432 subunit, NADPH-dependent glutamate synthase, and cell division protein ZipN/Ftn2/Arc6 (all FDR-  
433 corrected p-values < 0.05). In contrast, on the phosphorus-rich branch with a common ancestor of K13-05  
434 and K13-06, we found evidence of positive selection for a gene annotated to encode a Fe-S cluster  
435 assembly protein SufB, a domain of unknown function (DUF748), pyruvate phosphate dikinase  
436 PEP/pyruvate-binding protein, and a biosynthetic arginine decarboxylase (all FDR-correct p-values <  
437 0.05).

438

#### 439 Taxonomic and Functional Variation in the *M. aeruginosa* Phycosphere

440

441 Given clear divergence among isolates of *M. aeruginosa* across a phosphorus gradient, we asked  
442 whether associated bacteria in the phycosphere of *M. aeruginosa* also varied across this gradient. First, we  
443 found community-level divergence among the phycospheres associated with the three phylogenetic groups  
444 of hosts (Fig. 4A; adonis  $F_{2,49} = 2.4$ ,  $p = 0.001$ ,  $R^2 = 0.092$ ). Strongest separation was between the  
445 phycospheres of LL/LG hosts and that of both HL/HG and HL/LG hosts. We provide a taxonomic  
446 description of the core microbiome of *M. aeruginosa* in Fig. S9. Among HL/HG and HL/LG hosts,  
447 *Phycosocius bacilliformis* was the most abundant taxon in the phycosphere, comprising on average, 19.6%  
448 of the community (Table S3). In contrast, among LL/LG hosts, Caulobacterales and Cytophagaceae OTUs  
449 were more abundant, comprising 18.1 and 18.2% of the community respectively (Table S3).

450 Despite taxonomic differences among the phycosphere communities associated with hosts  
451 belonging to each phylogenetic group, we found no significant differences in protein functionality (Fig.  
452 4B; adonis not significant). Phycosphere functionality also did not differ by lake of origin (Fig. S9C;  
453 adonis not significant). In general, the phycospheres associated with different phylogenetic groups  
454 contained few unique functions (Fig. S10). Phycospheres of LL/LG hosts showed increased FeS cluster  
455 binding, as well as increased biosynthetic capacity for fatty acids, serine, threonine, histidine, and  
456 ubiquinone (Fig. S10). Nif genes indicative of nitrogen-fixation were also more commonly associated with  
457 the phycospheres of HL/LG (81%: 9 of 11 isolates) and LL/LG (50%: 9 of 18) than HL/HG (35%: 6 of 17  
458 isolates). However, 17 of these phycosphere communities contained only the nifA gene, which is the key

459 transcriptional regulator of *nif* genes, but may also regulate genes not involved in nitrogen fixation  
460 (Nienaber et al., 2000).

461 We chose *P. bacilliformis* for more in-depth investigation of function because this taxon occurred  
462 in all 46 phycospheres and was found, on average, at the greatest relative abundance via 16S marker gene  
463 surveys (Fig. S9). We detected substantial genome variation within *P. bacilliformis*. In contrast to a lab  
464 contaminant that would be identical in all cultures, we identified 7 genotypes that were each at least 96%  
465 complete and at least 0.70% divergent from all other genotypes (Table S4). To shed light on the  
466 predominance of *P. bacilliformis* in the *M. aeruginosa* phycosphere, we compared the metabolic  
467 capacities of each of 7 *M. aeruginosa* hosts with each of their associated genotypes of *P. bacilliformis*. We  
468 found that while no hosts harbored genes for the synthesis of threonine or serine, each *P. bacilliformis*  
469 genome showed the capacity to biosynthesize each of these amino acids via L-serine synthesis from 3-  
470 phosphoglycerate, L-threonine synthesis from L-homoserine, and L-homoserine synthesis from L-  
471 aspartate. Additionally, only one *M. aeruginosa* genome showed evidence of asparagine synthesis, while  
472 all *P. bacilliformis* genomes contained pathways for tRNA-dependent asparagine synthesis. Four genomes  
473 of *P. bacilliformis* were also indicative of galactose utilization via the Leloir Pathway (Table S4).

474

#### 475 Functional Interdependence between *M. aeruginosa* and the Phycosphere

476

477 In addition to assessing host and host microbiome function independently, we aimed to determine  
478 whether genome variation within a host provides further predictive power beyond the lake environment  
479 and time in predicting the functional capacity of the *M. aeruginosa* phycosphere. We selected 8 isolates  
480 from oligotrophic Gull Lake and 5 isolates from eutrophic Wintergreen Lake collected on August 8, 2013.  
481 This approach retained substantial genetic variation among host isolates on which to test our hypothesis.  
482 Among Gull isolates, 46.5% of homologous gene clusters were shared across all 8 isolates and 59.6% of  
483 clusters were shared across 7 isolates. Similarly, among Wintergreen isolates, 50.9% of homologous gene  
484 clusters were shared across all 5 isolates and 68.6% of clusters were shared across 4 isolates. Functional  
485 divergence between hosts showed a weak, positive correlation with functional divergence between  
486 phycosphere communities (Fig. 4C: linear mixed-effects model, marginal  $R^2 = 0.093$ ,  $p = 0.060$  or  
487 similarly, best fitting model with polynomial term, marginal  $R^2 = 0.14$ ,  $p = 0.065$ ). We show results from  
488 the eutrophic and oligotrophic lake with the greatest number of isolates collected on a single date because  
489 we found that there was a sharp reduction in detectability of a significant relationship as sample size  
490 decreased. For example, among Gull Lake 2013 isolates, we tested all subsetted sample-sizes of 4, 5, 6,

491 and 7 isolates and found that the probability of finding a significant result was 15.7, 25, 60.7, and 100%  
492 respectively.

493

#### 494 Physiological Variation within *M. aeruginosa*

495

496 One colony (source lake TP = 16.6  $\mu\text{g L}^{-1}$ ) exhibited negative growth during the experiment and  
497 was omitted from analysis, leaving  $n = 18$  for the experiment. We found that the observed genetic  
498 differences in *M. aeruginosa* corresponded with physiological differences among isolates, with maximum  
499 intrinsic growth rate being significantly and positively correlated with the total phosphorus concentration  
500 (TP, an index of trophic status) of the source lake (linear regression:  $n = 18$ ,  $p = 0.030$ ,  $R^2 = 0.26$ ).

501 Observed growth rates ranged nearly 7-fold from 0.08  $\text{d}^{-1}$  (source lake TP = 7.9  $\mu\text{g L}^{-1}$ ) to 0.55  $\text{d}^{-1}$  (source  
502 lake TP = 196.1  $\mu\text{g L}^{-1}$ ). More specifically, LL/LG and HL/LG isolates tended to grow more slowly at  
503 saturating resource levels than HL/HG isolates (Fig. 5; linear mixed-effects regression:  $F_{2,17} = 2.8$ ,  $p =$   
504 0.053,  $R^2 = 0.27$ ; LL/LG isolates:  $\mu_{\text{max}} = 0.162 \pm 0.038$  s.e., HL/LG isolates:  $\mu_{\text{max}} = 0.198 \pm 0.012$ , and  
505 HL/HG isolates:  $\mu_{\text{max}} = 0.327 \pm 0.061$ ). This correlation between  $\mu_{\text{max}}$  and lake TP remained evident when  
506 adding strains from an additional 11 Michigan lakes reported by Wilson et al. (2006) (see analysis in Fig.  
507 S11). In contrast to the evident correlation with lake TP, growth rate did not correspond with either  $\text{NH}_4^+$   
508 or  $\text{NO}_3^-$  concentrations of the source lake (linear regressions:  $n = 18$ ,  $p = 0.41$ ,  $R^2 = 0.036$  and  $n = 18$ ,  $p =$   
509 0.70,  $R^2 = 0.011$  respectively). Further, the HL/HG group was notably variable in growth rate. Possible  
510 drivers of this variability are unclear. The four strains exhibiting slower growth rates originated from four  
511 different lakes. Further, strains most closely related to each other according to the phylogeny shown in  
512 Fig. 2 fell on opposite sides of the growth gradient ( $\mu_{\text{max}}$  of W11-03 = 0.115 versus W11\_06 = 0.489, and  
513 L211-101 = 0.218 versus L211-11 = 0.419).

514

#### 515 **Discussion:**

516

517 Biodiversity at the within-species scale can have broad consequences for community structure and  
518 ecosystem function (Bolnick et al., 2011; Crutsinger et al., 2006; Whitham et al., 2006). Understanding  
519 factors that correspond with intraspecific variation across niche space is therefore key to understanding the  
520 maintenance of vital functions across environmental gradients. We show that intraspecific niche  
521 divergence can culminate from varied ecological and evolutionary factors. In *M. aeruginosa*, these  
522 contributing factors include genome evolution and frequently, an altered life history strategy of clonal vs.  
523 non-clonal colony formation. We also showed shifts in the host microbiome due to environmental



524 selection, as well as correlations between host genome variation and microbiome function. Independent,  
525 and perhaps synergistic, effects of each of these factors play important roles in the fitness of *M.*  
526 *aeruginosa* across a phosphorus gradient of freshwater lakes in Michigan. These results shed light on the  
527 potential importance of intraspecific variation in regulating dynamics of cyanobacterial harmful algal  
528 blooms.

529 Our study broadens understanding of this environmentally important cyanobacterium from  
530 genotypes of *M. aeruginosa* that are adapted to phosphorus-rich environments (Meyer et al., 2017) to  
531 genotypes of *M. aeruginosa* that evidently correspond with survival and colony formation of *Microcystis*  
532 in low-phosphorus environments. Architecture of these genomes is indicative of selective pressure towards  
533 more streamlined cellular efficiency. Although genome size is the most notable feature of streamlining  
534 theory, more broadly, streamlining refers to selection in resource-poor environments towards reductions in  
535 resource use, cell and genome size, and cell complexity (Giovannoni et al. 2014). Compared to well  
536 described examples of streamlining in marine cyanobacteria (i.e. *Prochlorococcus*, Rocap et al., 2013),  
537 our results in *M. aeruginosa* are much smaller in effect size. Subtle shifts in our dataset are not surprising  
538 considering we surveyed intraspecific differences across a relatively small geographic region.

539 Genomes from the LL/LG and HL/LG groups lacked potentially important functions compared to  
540 the HL/HG group. LL/LG and HL/LG genomes frequently lacked 11 core cyanobacterial genes, however  
541 the database that identifies 'core' genes consists mostly of *M. aeruginosa* isolates collected from eutrophic  
542 environments. LL/LG and HL/LG genomes frequently lacked *cobS*, which is necessary for synthesis of  
543 cobalamin, Vitamin B12. Frequent absence of *mutT*, involved in preventing DNA mutations, in LL/LG  
544 and HL/LG isolates suggests a higher mutational rate may be beneficial for adaptation to oligotrophic  
545 environments (Denamur & Matic, 2006). However, because *mutT* prevents AT to GC transversions  
546 (Yanofsky et al., 1966), it is unclear how LL/LG and HL/LG genomes lacking *mutT* would maintain a  
547 lower GC content than HL/HG isolates. Future study is warranted into whether subsequent or concurrent  
548 genetic changes occurred with the loss of *mutT* to facilitate increased mutational rate while maintaining  
549 low GC content.

550 In addition to gene loss and reduced genome complexity, acquisition of new gene functions may  
551 contribute to survival under phosphorus-limited environments. Most notable, we found that LL/LG  
552 isolates contained alkaline phosphatases that were absent in HL/HG and HL/LG genomes. Of three known  
553 alkaline phosphatase protein families (PhoA, PhoD, and PhoX (Luo et al., 2009)), only PhoX has been  
554 previously described in *M. aeruginosa* (Harke et al., 2012), and is also common among bacteria of the  
555 oligotrophic open ocean (Kathuria & Martiny, 2011; Luo et al., 2009; Sebastian & Ammerman, 2009).  
556 Two factors that may be facilitating LL/LG survival in low-P environments are that all LL/LG and HL/LG

557 isolates gained a second PhoX gene, and that LL/LG isolates gained an additional gene annotated as PhoA  
558 (pfam00245). Considering that PhoA requires activation by zinc and magnesium ions, while PhoX  
559 requires bioavailable calcium, different alkaline phosphatases may be more advantageous for *M.*  
560 *aeruginosa* inhabiting different environments.

561 We also had two notable findings in regards to histidine kinases and response regulators, which  
562 enable cells to detect and respond to changes in its environment such as nutrients and light. A gene  
563 annotated to histidine kinase KdpD was found to occur in significantly less abundance in the LL/LG group  
564 than the HL/HG or HL/LG groups, which is aligned with findings of a substantial loss of response  
565 regulation in the streamlined cyanobacterium *Prochlorococcus* (Mary and Vault, 2003). On the other  
566 hand, a gene annotated to histidine kinase SsrA was under positive selection only in the LL/LG group,  
567 which is aligned with the finding that certain histidine kinases are essential for survival of the  
568 cyanobacterium *Synechococcus* under extreme nutrient limitation (Schwarz and Grossman, 1998).

569 We also inferred differences in life history traits among phylogenetic groups. We infer that in  
570 contrast to clonal colony formation in HL/HG and HL/LG isolates, LL/LG *M. aeruginosa* colonies more  
571 frequently (though not exclusively) initially form through non-clonal cellular adhesion. Colony formation  
572 can occur either gradually via clonal growth or rapidly via cell adhesion, and both modes of colony  
573 formation can be induced by abiotic and biotic stressors including predation (Xiao et al., 2018). In light of  
574 lower growth rates among LL/LG isolates, cell adhesion may be compensate for slower clonal colony  
575 growth. The increased frequency of polymorphic sites observed in LL/LG genomes are beyond what could  
576 be explained by *de novo* mutations during lab culturing (~ 1,200 generations between field collection and  
577 sequencing (Baldia et al., 2007)). We also found similar heterogeneity between isolates collected in 2011  
578 and 2013 despite differences in culturing time. Further supporting an increased occurrence of genomic  
579 heterogeneity among LL/LG colonies, we found increased frequencies of genes in LL/LG genomes that  
580 may facilitate non-clonal colony formation: (a) NeuB synthesizes sialic acids, which are a component of  
581 cyanobacterial extracellular polymeric substances (Zippel & Neu, 2011; Strom et al., 2017), and play  
582 important roles in cellular recognition and adhesion (Gunawan et al., 2005), (b) Sialidases may facilitate  
583 cellular adhesion by uncovering carbohydrate receptors that are recognized by bacterial adhesins (Vimr,  
584 1994), (c) pfam06229 (FRG1-like domain) that contains a *Hydra spp.* gene linked to this freshwater  
585 cnidarian's ability to adhere to underwater surfaces (Rodrigues et al., 2016), and (d) a bacterial adhesin  
586 (pfam03865: Haemolysin secretion/activation protein ShlB/FhaC/HecB) (Moslavac et al., 2005) that  
587 facilitates adhesion to other cells in pathogenic and symbiotic interactions (Hooper & Gordon, 2001).  
588 Genes regulating cell recognition and adhesion should be less important for clonal colonies, where  
589 daughter cells remain attached after binary fission (Kessel & Eloff, 1975). We note that all colonies that

590 we isolated, including the evidently non-clonal LL/LG colonies, were of the distinctive, tightly-packed  
591 morphology typical of natural *M. aeruginosa* colonies. In contrast, experimentally induced non-clonal  
592 colonies are amorphous, loose aggregations of cells that do not resemble *M. aeruginosa* morphologies  
593 common in nature. The non-amorphous shapes of the non-clonal LL/LG colonies suggests that growth  
594 occurred mostly by cell division but may have included some level of aggregation only at earlier stages of  
595 colony development. Lastly, the colonies that we isolated may have been descendants from the initial  
596 colonies that formed via some level of cell aggregation.

597 We also note that the HL/HG group of *M. aeruginosa* was more highly variable than the LL/LG  
598 group, which corresponds to the variation in nutrient levels observed in high versus low nutrient lakes.  
599 HL/HG isolates were more variable according to our MLST phylogeny as well as the genome wide  
600 metrics, GC content and percentage of coding DNA. Greater genome variation in the HL/HG group also  
601 translated into greater variance in function, according to protein family annotation, as well as physiology  
602 ( $\mu_{max}$ ). Low-nutrient lakes generally always have low concentrations of bioavailable nutrients,  
603 particularly in the summer epilimnion of deep, stratified lakes where co-limitation by both N and P may be  
604 especially common. In contrast, high-nutrient lakes have higher average concentrations than low-nutrient  
605 lakes, but also have much greater variance, including occasional low concentrations following periods of  
606 intense phytoplankton growth (Sarnelle, 1992). The lowest nutrient levels tend to occur in the epilimnion  
607 during late summer, where and when *M. aeruginosa* tends to reach peak abundance (Sarnelle, 1992).

608 In contrast to clear genomic divergence of *M. aeruginosa* across the phosphorus gradient, we  
609 found subtle changes in the phycosphere. Functional convergence despite taxonomic divergence as we see  
610 in this study has also been noted in other study systems, including the human gut microbiome (Turnbaugh  
611 et al. 2009), however there are also several limitations to our approach that may explain this result. First,  
612 although the functional component of our study analyzes pfam profiles based on near complete genomes  
613 of phycosphere bacteria, sequence variation of proteins belonging to the same pfam, either present in  
614 different species or in different strains of the same species may alter phycosphere functionality in ways not  
615 reflected by our current analysis. In addition, the phycosphere of *M. aeruginosa* may be similar in genome  
616 functions across the trophic gradient, but may express different genes when associated with hosts  
617 belonging to different phylogenetic groups. Analyses of these differences are beyond the scope of this  
618 study, but are worth further study. Second, to generate adequate quantities of DNA for sequencing, and to  
619 enable physiological characterization, all isolates were cultured under common garden, laboratory-based  
620 conditions. Phycosphere community composition likely changed during this transition from the natural  
621 environment. Yet, as a host genotype effect on overall taxonomic and specific functional gene content  
622 remained by both phylogenetic group and lake of origin, these changes were constrained by the limits

623 imposed by the natural community associated with each individual colony from which the cultures were  
624 started. Future studies repeating our analysis on sequence data generated directly from individual colonies  
625 collected from the environment will clarify the extent to which laboratory culturing impacted the host  
626 genotypic signal.

627 Despite these limitations, phycosphere community composition remained representative of  
628 heterotrophic bacteria associated with blooms of *M. aeruginosa*. For example, two of the three most  
629 abundant heterotrophic taxa across our isolate collection, *P. bacilliformis* and Cytophagaceae (Table S3),  
630 are strongly associated with blooms of *M. aeruginosa* (Tanabe et al., 2015; Berry et al., 2016). We had  
631 hypothesized that the phycosphere could facilitate survival of a streamlined host with atypical nutrient  
632 requirements caused by gene loss, e.g. the loss of key genes for amino acid biosynthesis in LL/LG isolates.  
633 Our hypothesis is based on the Black Queen Hypothesis which proposed that streamlined bacteria may  
634 compensate for gene losses through increased community connectivity (Morris et al., 2012). In contrast, we  
635 found convergence in phycosphere gene functions across the nutrient gradient. This convergence in  
636 phycosphere function may be driven in part to providing uniform culturing conditions with standardized  
637 nutrients and vitamins. In contrast to function, taxonomic composition of the phycosphere varied across  
638 the gradient. Considering a common garden likely caused some degree of convergence among  
639 phycospheres, this magnitude of taxonomic divergence likely underestimates divergence under natural  
640 environmental conditions. Functional convergence despite taxonomic divergence suggests that hosts may  
641 select for certain essential functions among the available pool of heterotrophic bacteria, which themselves  
642 are strongly shaped by lake environmental conditions (Crump et al., 2007). This host-mediated selection  
643 of phycosphere function was also apparent in how functional similarity of the host phycosphere was  
644 weakly predicted by the functional similarity of the host. Although this was a weak prediction that  
645 accounted for less than 10% of the total variance among phycosphere functionality, these results  
646 correspond with findings that intraspecific plant variation can have small but significant influences on the  
647 rhizosphere of maize and *Arabidopsis* genotypes (Lundberg et al., 2012; Peiffer et al., 2013). Different  
648 species of phytoplankton hosts have been shown to harbor distinct phycosphere communities (Jasti et al.  
649 2005; Eigemann et al. 2013), but the relative effects of intra- versus interspecific-level variation on the  
650 phycosphere has not been directly studied.

651 One notable functional contribution of LL/LG and HL/LG *M. aeruginosa* phycospheres that may  
652 explain survival across a nutrient gradient was an increased occurrence of nitrogen-fixation genes.  
653 However, much of this pattern was driven by the increased occurrence of only a single *nif* gene, *nifA*, so it  
654 is unclear what role these phycosphere taxa may contribute towards nitrogen assimilation. Additionally,  
655 genome investigation into *Phycosocius bacilliformis*, which we found in the phycosphere of each of our

656 *M. aeruginosa* isolates, suggests that complementary amino acid biosynthesis may be one component of  
657 this symbiotic interaction. Previously identified in surveys of *M. aeruginosa* blooms, *P. bacilliformis* can  
658 increase growth of colonial green algae (Tanabe et al., 2015). Several genotypes of *P. bacilliformis* appear  
659 to derive a significant source of energy from galactose, which is the primary component of the  
660 polysaccharide-based mucilage that binds cells of colonial *M. aeruginosa* (Plude et al., 1991; Rohrlack et  
661 al., 1999).

662 From a methodological perspective, our case study emphasizes the need for cautious interpretation  
663 of metagenome-assembled genomes. Comparing genome traits and functions across phylogenetic groups  
664 was challenging due to fundamental differences in genome architecture and assembly. For example,  
665 repetitive elements are a well known technical complication in genome assembly (Treangen & Salzberg,  
666 2012). Our LL/LG assemblies were the most fragmented due to a larger number and/or problematic  
667 locations of repetitive elements, as well as a higher incidence of colony heterogeneity. *Microcystis*  
668 *aeruginosa* is noted for an unusually high percentage of mobile repetitive elements (Kaneko et al., 2007).  
669 Such varied genomic architecture can lead to biased inferences. For example, when we initially binned  
670 only contigs over 4kb in length, because LL/LG genomes contained a sizable portion of contigs under 4kb  
671 in length, we incorrectly inferred that LL/LG genomes were smaller in size and lacked many more gene  
672 functions. In-depth investigation of these patterns revealed that many genes seemingly missing from  
673 LL/LG genomes were merely on contigs shorter than 4kb. This led us to the cautious approach of binning  
674 2 kb fragments and adding even shorter contigs that were taxonomically classified as *M. aeruginosa*.

675 Genome data of *M. aeruginosa*, which dominates freshwater harmful algal blooms worldwide, has  
676 previously existed only for isolates that originated from phosphorus-rich environments. Our findings that  
677 *M. aeruginosa* inhabiting oligotrophic environments differ in genome structure, function, and life history  
678 compared to isolates derived from eutrophic environments, has important implications for understanding  
679 the dynamics of harmful algal blooms and their response to ongoing global change. Particularly notable,  
680 we found that divergence among *M. aeruginosa* isolates results in a growth tradeoff. HL/HG isolates  
681 retain the ability for rapid growth when resources are high (i.e., a greater  $\mu_{\text{MAX}}$  that matches rate estimates  
682 reported by Wilson et al. 2006 & 2010, Reynolds 2006, Seip and Reynolds 1995 for colonial *Microcystis*).  
683 LL/LG and HL/LG isolates have acquired the ability to subsist when resources are low but at the cost of  
684 an ability to increase growth when resources become more readily available. Similar growth tradeoffs  
685 have been observed in streamlined bacteria (Giovannoni et al., 2005). This tradeoff is regarded as a key  
686 property of oligotrophs versus copiotrophs, which instead have growth rates more responsive to nutrient  
687 flux (Koch, 2001). Further, this growth trade-off is an important future direction to consider in the context  
688 of harmful algal blooms. These blooms can last several months, during which time bloom development

689 drives down available nutrients in the water column to low levels (Heisler et al., 2008; Sarnelle, 1992),  
690 thus constructing a niche for low-nutrient adapted genotypes. Our findings are especially notable  
691 considering these distinct architectures occurred within a species of cyanobacterium within a single lake.  
692 Co-occurrence of such isolates adapted to thrive in different micro-environments may have critical  
693 implications for the temporal variability and spatial extent of toxic cyanobacterial blooms. For example,  
694 the HL/LG-type isolates residing in eutrophic and mesotrophic lakes may be key players in extending the  
695 duration of blooms after HL/HG-type isolates have depleted phosphorus to levels that would otherwise  
696 lead to a recovery period or dominance of other non-toxic phytoplankton. Such physiological adaptation to  
697 low-phosphorus conditions also helps explain the recent expansion of *M. aeruginosa* in oligotrophic lakes  
698 invaded by dreissenid mussels, lakes which are otherwise an uncharacteristic habitat for this  
699 cyanobacterium (Raikow et al., 2004; Knoll et al. 2008; Sarnelle et al., 2010).

700 Overall, evolutionary divergence of *M. aeruginosa* corresponds with maintenance of high fitness  
701 across a wide phosphorus gradient. Evolutionary changes included direct effects on the host genome that  
702 increased nutrient-use efficiency and nutrient assimilation. However, genomic changes within the  
703 organism do not operate independently from its ecology. Changes within *M. aeruginosa* genomes may  
704 have facilitated changes in the behavioral ecology of the cyanobacterium by acquiring gene functions that  
705 have seemingly enabled an altered life history strategy of non-clonal colony formation. Host genome  
706 changes further correspond with changes in the symbiotic and/or commensal ecological interactions  
707 between the host and the host microbiome, in which function of the host is linked with function of the host  
708 microbiome. These findings demonstrate the intricate and non-independence of ecological and  
709 evolutionary processes that may facilitate intraspecific niche divergence.

710

#### 711 **Acknowledgments:**

712

713 This project was supported by funding from NOAA distributed through the Cooperative Institute for Great  
714 Lakes Research (NA17OAR4320152) and the National Science Foundation (Division of Environmental  
715 Biology-1737680) to VJD, a Dow Sustainability Postdoctoral Fellowship to SLJ, the National Science  
716 Foundation (Division of Environmental Biology-0841864, Division of Environmental Biology-0841944)  
717 to OS, and the Gull Lake Quality Organization and the Robert C. Ball and Betty A. Ball Fisheries and  
718 Wildlife Fellowship at Michigan State University to JDW.

719

#### 720 **References:**

721 Anantharaman, K., Duhaime, M. B., Breier, J. A., Wendt, K. A., Toner, B. M. & Dick, G. J.  
722 (2014). Sulfur oxidation genes in diverse deep-sea viruses. *Science*, *344*, 757-760.

723 Baldia, S. F., Evangelista, A. D., Aralar, E.V., & Santiago, A. E. (2007). Nitrogen and phosphorus  
724 utilization in the cyanobacterium *Microcystis aeruginosa* isolated from Laguna de Bay, Philippines.  
725 *Journal of Applied Phycology*, *19*, 607-613.

726 Barrett, R. D. H., & Schluter, D. (2008). Adaptation from standing genetic variation. *Trends in*  
727 *Ecology & Evolution* *23*, 38-44.

728 Bassar, R. D., Marshall, M. C., López-Sepulcre, A., Zandonà, E., Auer, S. K., Travis, J., Pringle,  
729 C. M., Flecker, A. S., Thomas, S. A., Fraser, D. F., & Reznick, D. N. (2010). Local adaptation in  
730 Trinidadian guppies alters ecosystem processes. *Proceedings of the National Academy of Sciences*, *107*,  
731 3616-3621.

732 Benjamini, Y., & Hochberg, Y. (1995). Controlling the False Discovery Rate: A Practical and  
733 Powerful Approach to Multiple Testing. *Journal of the Royal Statistical Society Series B*, *57*, 289-300.

734 Bergmann, G. T., Bates, S. T., Eilers, K. G., Lauber, C. L., Caporaso, J. G., Walters, W. A.,  
735 Knight, R., & Fierer, N. (2011). The under-recognized dominance of Verrucomicrobia in soil bacterial  
736 communities. *Soil Biology and Biochemistry*, *43*, 1450-1455.

737 Berry, M. A., White, J. D., Davis, T. W., Jain, S., Johengen, T. H., Dick, G. J., Sarnelle, O., &  
738 Denef, V. J. (2017). Are oligotypes meaningful ecological and phylogenetic units? A case study of  
739 *Microcystis* in freshwater lakes. *Frontiers in Microbiology*, *8*, 1-7.

740 Berry, M. A., Davis, T. W., Cory, R. M., Duhaime, M. B., Johengen, T. H., Kling, G. W., Marino,  
741 J. A., Den Uyl, P. A., Gossiaux, D., Dick, G. J., & Denef, V. J. (2017). Cyanobacterial harmful algal  
742 blooms are a biological disturbance to western Lake Erie bacterial communities. *Environmental*  
743 *Microbiology*, *19*, 1149 – 1162.

744 Bolnick, D. I., Amarasekare, P., Araújo, M. S., Bürger, R., Levine, J. M., Novak, M., Rudolf, V.  
745 H. M., Schreiber, S. J., Urban, M. C., & Vasseur, D. A. 2011. Why intraspecific trait variation matters in  
746 community ecology. *Trends in Ecology and Evolution*, *26*, 183-192.

747 Buchfink, B., Xie, C., & Huson, D. H. (2014). Fast and sensitive protein alignment using  
748 DIAMOND. *Nature Methods*, *12*, 59-60.

749 Burkholder, J. M., & Gilbert, P. M. (2009). The importance of intraspecific variability in harmful  
750 algae- preface to a collection of topical papers. *Harmful Algae*, *8*, 744-745.

751 Cai, H., Jiang, H., Krumholz, L. R., & Yang, Z. (2014). Bacterial community composition of size-  
752 fractionated aggregates within the phycosphere of cyanobacterial blooms in a eutrophic freshwater lake.  
753 *Plos ONE*, *9*, e102879.

754 Chislock, M. F., O. Sarnelle, B. Olsen, E. Doster, & A. E. Wilson. 2013. Large effects of  
755 consumer offense on ecosystem structure and function. *Ecology* 94: 2375-2380

756 Contreras-Moreira, B., Vinuesa, P. (2013). GET\_HOMOLOGUES, a versatile software package  
757 for scalable and robust microbial pan-genome analysis. *Applied and Environmental Microbiology*, 79,  
758 7696-7701.

759 Costas, E., Lopez-Rodas, V., Javier Toro, F., & Flores-Moya, A. (2008). The number of cells in  
760 colonies of the cyanobacterium *Microcystis aeruginosa* satisfies Benford's Law. *Aquatic Botany*, 89, 341-  
761 343.

762 Crump, B. C., Adams, H. E., Hobbie, J. E., & Kling, G. W. (2007). Biogeography of  
763 bacterioplankton in lakes and streams of anarctic tundra catchment. *Ecology*, 88, 1365-1378.

764 Crutsinger, G. M., Collins, M. D., Fordyce, J. A., Gompert, Z., Nice, C. C., & Sanders, N. J.  
765 (2006). Plant genotypic diversity predicts community structure and governs an ecosystem process.  
766 *Science*, 313, 966-968.

767 DeMott, W. R., & Mckinney, E. N. (2015). Use it or lose it? Loss of grazing defenses during  
768 laboratory culture of the digestion-resistant green alga *Oocystis*. *Journal of Plankton Research*, 37, 399-  
769 408.

770 Denamur, E., & Matic, I. (2006). Evolution of mutation rates in bacteria. *Molecular Microbiology*,  
771 60, 820-827.

772 Dick, G. J., Andersson, A. F., Baker, B. J., Simmons, S. L., Thomas, B. C., Yelton, A. P., &  
773 Banfield, J. F. (2009). Community-wide analysis of microbial genome sequence signatures. *Genome*  
774 *Biology*, 10, R85.

775 Dittami, S. M., Duboscq-Bidot, L., Perennou, M., Gobet, A., Corre, E., Boyen, C., & Tonon, T.  
776 (2016). Host-microbe interactions as a driver of acclimation to salinity gradients in brown algal cultures.  
777 *ISME Journal*, 10, 51-63.

778 Drake, J. W., Charlesworth, B., Charlesworth, D., & Crow, J. F. (1998). Rates of spontaneous  
779 mutation. *Genetics*, 148, 1667-1686.

780 Edgar, R. C. (2004). MUSCLE: Multiple sequence alignment with high accuracy and high  
781 throughput. *Nucleic Acids Research*, 32, 1792-1797.

782 Eigemann, F., Hilt, S., Salka, I. & Grossart, H.-P. (2013). Bacterial community composition  
783 associated with freshwater algae: Species specificity versus dependency on environmental conditions and  
784 source community. *FEMS Microbiology Ecology*, 32, 650 - 663.

785 Franguel, L., Quillardet, P., Castets, A., Humbert, J., Matthijs, H. C. P., Cortez, D., Tolonen, A.,  
786 Zhang, C., Gribaldo, S., Kehr, J., Zilliges, Y., Ziemert, N., Becker, S., Talla, E., Latifi, A., Billault, A.,



787 Lepelletier, A., Dittmann, E., Bouchier, C., & Tandeau de Marsac, N. (2008). Highly plastic genome of  
788 *Microcystis aeruginosa* PCC 7806, a ubiquitous toxic freshwater cyanobacterium. *BMC Genomics*, *9*, 274.

789 Giovannoni, S. J., Thrash, J. C., & Temperton, B. (2014). Implications of streamlining theory for  
790 microbial ecology. *ISME Journal*, *8*, 1553-1565.

791 Giovannoni, S. J., Tripp, H. J., Givan, S., Podar, M., Vergin, K. L., Baptiasa, D., Bibbs, L. Eads,  
792 J., Richardson, T. H., Noordewier, M., & Rappé, M. S. (2005). Genome streamlining in a cosmopolitan  
793 oceanic bacterium. *Science*, *309*, 1242-1245.

794 Gunawan, J., Simard, D., Gilbert, M., Lovering, A. L., Wakarchuk, W. W., Tanner, M. E., &  
795 Strynadka, N. C. J. (2005). Structural and mechanistic analysis of sialic acid synthase NeuB from  
796 *Neisseria meningitidis* in complex with Mn<sup>2+</sup>, phosphoenolpyruvate, and N-acetylmannosaminol.  
797 *Journal of Biological Chemistry*, *280*, 3555-3563.

798 Harke, M. J., Berry, D. L., Ammerman, J. W., & Gobler, C. J. (2012). Molecular response of the  
799 bloom-forming cyanobacterium, *Microcystis aeruginosa*, to phosphorus limitation. *Microbial Ecology*, *63*,  
800 188-198.

801 Heisler, J., Glibert, P. M., Burkholder, J. M., Anderson, D. M., Cochlan, W., Dennison, W.C.,  
802 Dortch, Q., Gobler, C. J., Heil, C. A., Humphries, E., Lewitus, A., Magnien, R., Marshall, H. G., Sellner,  
803 K., Stockwell, D. A., Stoecker, D. K., & Suddleson, M. (2008). Eutrophication and harmful algal blooms:  
804 A scientific consensus. *Harmful Algae*, *8*, 3-13.

805 Hooper, L. V., & Gordon, J. I. (2001). Glycans as legislators of host–microbial interactions:  
806 Spanning the spectrum from symbiosis to pathogenicity. *Glycobiology*, *11*, 1-10.

807 Humbert, J. F., Barbe, V., Latifi, A., Gugger, M., Calteau, A., Coursin, T., Lajus, A., Castelli, V.,  
808 Oztas, S., Samson, G. & Longin, C. (2013). A tribute to disorder in the genome of the bloom-forming  
809 freshwater cyanobacterium *Microcystis aeruginosa*. *PLoS One*, *8*, p.e70747.

810 Huntemann, M., Ivanova, N. N., Mavromatis, K., Tripp, H. J., Paez-Espino, D., Palaniappan, K.,  
811 Szeto, E., Pillay, M., Chen, I. M., Pati, A., Nielsen, T., Markowitz, V. M., & Kyrpides, N. C. (2015). The  
812 standard operating procedure of the DOE-JGI Microbial Genome Annotation Pipeline (MGAP v.4).  
813 *Standards in Genomic Sciences*, *10*, 86.

814 Jackrel, S. L., White, J. D., Evans, J. T., Buffin, K., Hayden, K., Sarnelle, O. & Deneff, V. J.  
815 (2019). *Microcystis* cultures isolated from Michigan inland lakes genome sequencing, assembly, and  
816 targeted locus. *NCBI SRA*.

817 Jasti, S., Sieracki, M. E., Poulton, N. J., Giewat, M. W., & Rooney-Varga, J. N. (2005).  
818 Phylogenetic diversity and specificity of bacteria closely associated with *Alexandrium* spp. and other  
819 phytoplankton. *Applied and Environmental Microbiology*, *71*, 3483-3494.

820 Johnson, Z. I., Zinser, E. R., Coe, A., McNulty, N. P., Woodward, E. M. S., & Chisholm, S. W.  
821 (2006). Niche partitioning among *Prochlorococcus* ecotypes along ocean-scale environmental gradients.  
822 *Science*, 311, 1737-1740.

823 Joshi, N. A., & Fass, J. N. (2011). Sickle: A sliding-window, adaptive, quality-based trimming  
824 tool for FastQ files. V.1.33.

825 Kaneko, T., Nakajima, N., Okamoto, S., Suzuki, I., Tanabe, Y., & Tamaoki, M., Nakamura, Y.,  
826 Kasai, F., Watanabe, A., Kawashima, K., Kishida, Y., Ono, A., Shimizu, Y., Takahashi, C., Minami, C.,  
827 Fujishiro, T., Kohara, M., Nakazaki, N., Nakayama, S., Yamada, M., Tabata, & Watanabe, M. M. (2007).  
828 Complete genomic structure of the bloom-forming toxic cyanobacterium *Microcystis aeruginosa* NIES-  
829 843. *DNA Research*, 14, 247-256.

830 Kathuria, S., & Martiny, A. C. (2011). Prevalence of a calcium-based alkaline phosphatase  
831 associated with the marine cyanobacterium *Prochlorococcus* and other ocean bacteria. *Environmental*  
832 *Microbiology*, 13, 74-83.

833 Kessel, M., & Eloff, J. N. (1975). The ultrastructure and development of the colonial sheath of  
834 *Microcystis marginata*. *Archives of microbiology*, 106, 209-214.

835 Knoll, L. B., Sarnelle, O. Hamilton, S. K., Scheele, C. E. H., Wilson, A. E., Rose, J. B., &  
836 Morgan, M. R. (2008). Invasive zebra mussels (*Dreissena polymorpha*) increase cyanobacterial toxin  
837 concentrations in low-nutrient lakes. *Canadian Journal of Fisheries and Aquatic Sciences*, 65, 448-455.

838 Koch, A. L. (2001). Oligotrophs versus copiotrophs. *Bioessays*, 23, 657-661.

839 Kozich, J. J., Westcott, S. L., Baxter, N. T., Highlander, S. K., & Schloss, P. D. (2013).  
840 Development of a dual-index sequencing strategy and curation pipeline for analyzing amplicon sequence  
841 data on the MiSeq Illumina sequencing platform. *Applied and Environmental Microbiology*, 79, 5112-  
842 5120.

843 Laczny, C. C., Sternal, T., Plugaru, V., Gawron, P., Atashpendar, A., Margossian, H. H.,  
844 Cordonado, S., van der Maaten, L., Vlassis, N & Wilmes P. (2015). VizBin-an application for reference-  
845 independent visualization and human-augmented binning of metagenomic data. *Microbiome*, 3,1.

846 Lakeman, M. B., von Dassow, P., & Cattolico, R. A. (2009). The strain concept in phytoplankton  
847 ecology. *Harmful Algae*, 8, 746-758.

848 Lamichhaney, S., Berglund, J., Almén, M. S., Maqbool, K., Grabherr, M., Martinez-Barrio, A.,  
849 Promerova, M., Rubin, C., Wang, C., Zamani, N., Grant, B. R., Grant, P. R., Webster, M. T., & L.  
850 Andersson. (2015). Evolution of Darwin's finches and their beaks revealed by genome sequencing.  
851 *Nature*, 518, 371.

852           Lau, J.A., & Lennon, J. T. (2012). Rapid responses of soil microorganisms improve plant fitness  
853 in novel environments. *Proceedings of the National Academy of Sciences*, *109*, 14058-14062.

854           Li, H., & Durbin, R. (2009). Fast and accurate short read alignment with Burrows–Wheeler  
855 transform. *Bioinformatics*, *25*, 1754-1760.

856           Li, H., Handsaker, B., Wysoker, A., Fennell, T., Ruan, J., Homer, N., Marth, G., Abecasis, G., &  
857 Durbin R. (2009). The sequence alignment/map format and SAMtools. *Bioinformatics*, *25*, 2078-2079.

858           Louati, I., Pascault, N., Debroas, D., Bernard, D., Humbert, J. F., & Leloup, J. (2015). Structural  
859 diversity of bacterial communities associated with bloom-forming freshwater cyanobacteria differs  
860 according to the cyanobacterial genus. *Plos ONE*, *10*, e0140614.

861           Lundberg, D. S., Lebeis, S. L., Herrera Paredes, S., Yourstone, S., Gehring, J., Malfatti, S.,  
862 Tremblay, J., Engelbrekston, A., Kunin, V., Glavina del Rio, T., Edgar, R. C., Eickhorst, T., Ley, R. E.,  
863 Hugenholtz, P., Green Tringe, S., & Dangl J. L. (2012). Defining the core *Arabidopsis thaliana* root  
864 microbiome. *Nature*, *488*, 86-90.

865           Luo, H., Benner, R., Long, R. A., & Hu, J. (2009). Subcellular localization of marine bacterial  
866 alkaline phosphatases. *Proceedings of the National Academy of Sciences*, *106*, 21219-21223.

867           Masango, M. G., Myrburgh, J. G., Labuschagne, L., Govender, D., Bengis, R. G., & Naicker, D.  
868 (2010). Assessment of *Microcystis* bloom toxicity associated with wildlife mortality in the Kruger  
869 National Park, South Africa. *Journal of Wildlife Diseases*, *46*, 95-102.

870           McMurdie, P. J., & Homes, S. (2014). Waste not, want not: Why rarefying microbiome data is  
871 inadmissible. *PLoS Computational Biology*, *10*, e1003531.

872           Menzel, D. W., & Corwin, N. (1965). The measurement of total phosphorus in seawater based on  
873 the liberation of organically bound fractions by persulfate oxidation 1. *Limnology and Oceanography*, *10*,  
874 280-282.

875           Meyer, K. A., Davis, T. W., Watson, S. B., Deneff, V. J., Berry, M. A., & Dick, G. J. (2017).  
876 Genome sequences of lower Great Lakes *Microcystis* sp. reveal strain-specific genes that are present and  
877 expressed in western Lake Erie blooms. *PLoS ONE*, *12*, e0183859.

878           Michalak, A. M., Anderson, E. J., Beletsky, D., Boland, S., Bosch, N.S., Bridgeman, T.B.,  
879 Chaffin, J.D., Cho, K., Confesor, R., Daloğlu, I., DePinto, J.V., et al. (2013). Record-setting algal bloom  
880 in Lake Erie caused by agricultural and meteorological trends consistent with expected future conditions.  
881 *Proceedings of the National Academy of Sciences*, *110*, 6448-6452.

882           Morris, J. J., Lenski, R. E., & Zinser, E. R. (2012). The Black Queen Hypothesis: evolution of  
883 dependencies through adaptive gene loss. *MBio*, *3*, e00036-12.

884 Moslavac, S., Mirus, O., Bredemeier, R., Soll, J., von Haeseler, A., & Schleiff, E. (2005).  
885 Conserved pore-forming regions in polypeptide-transporting proteins. *The FEBS Journal*, 272, 1367-1378.

886 Murphy, J., & Riley, J. P. (1962). A modified single solution method for the determination of  
887 phosphate in natural waters. *Analytica Chimica Acta*, 27, 31-36.

888 Newton, R. J., Jones, S. E., Eiler, A., McMahon, K. D., & Bertilsson, S. (2011). A guide to the  
889 natural history of freshwater lake bacteria. *Microbiology and Molecular Biology Reviews*, 75, 14–49.

890 Nienaber, A., Huber, A., Göttfert, M., Hennecke, H., & Fischer, H. M. (2000). Three new NifA-  
891 regulated genes in the *Bradyrhizobium japonicum* symbiotic gene region discovered by competitive DNA-  
892 RNA hybridization. *Journal of Bacteriology*, 182, 1472-1480.

893 Olm, M. R., Brown, C. T., Brooks, B., & Banfield, J. F. (2017). dRep: A tool for fast and accurate  
894 genomic comparisons that enables improved genome recovery from metagenomes through de-replication.  
895 *ISME Journal*, 11, 2864-2868.

896 Parks, D. H., Imelfort, M., Skennerton, C. T., Hugenholtz, P., & Tyson, G. W. (2015). CheckM:  
897 Assessing the quality of microbial genomes recovered from isolates, single cells, and metagenomes.  
898 *Genome Research*, 25, 1-13.

899 Parks, D. H., Tyson, G. W., Hugenholtz, P., & Beiko, R. G. (2014). STAMP: Statistical analysis  
900 of taxonomic and functional profiles. *Bioinformatics*, 30, 3123-3124.

901 Peiffer, J. A., Spor, A., Koren, O., Jin, Z., Green Tringe, S., Dangl, J. L., Buckler, E. S., & Ley, R.  
902 E. (2013). Diversity and heritability of the maize rhizosphere microbiome under field conditions.  
903 *Proceedings of the National Academy of Sciences*, 110, 6548-6553.

904 Pfennig, D. W., Wund, M. A., Snell-Rood, E. C., Cruickshank, T., Schlichting, C. D., & Moczek,  
905 A.P. 2010. Phenotypic plasticity impacts on diversification and speciation. *Trends in Ecology & Evolution*  
906 25, 459-467.

907 Plude, J. L., Parker, D. L., Schommer, O. J., Timmerman, R. J., Hagstrom, S. A., Joers, J. M., &  
908 Hnasko, R. (1991). Chemical characterization of polysaccharide from the slime layer of the  
909 cyanobacterium *Microcystis flos-aquae* C3-40. *Applied and Environmental Microbiology*, 57, 1696-1700.

910 Post, D. M., Palkovacs, E. P., Schielke, E. G., & Dodson, S. I. (2008). Intraspecific variation in a  
911 predator affects community structure and cascading trophic interactions. *Ecology*, 89, 2019-2032.

912 Pritchard, L., Glover, R. H., Humphris, S., Elphinstone, J. G., & Toth, I. K. (2016). Genomics and  
913 taxonomy in diagnostics for food security: Soft-rotting enterobacterial plant pathogens. *Analytical*  
914 *Methods*, 8, 12 – 24.

915           Qin, B., Zhu, G., Gao, G., Zhang, Y., Li, W., Paerl, H. W., & Carmichael, W. W. (2010). A  
916 drinking water crisis in Lake Taihu, China: Linkage to climatic variability and lake management.  
917 *Environmental Management*, 45, 105-112.

918           Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., et al. (2012). The SILVA  
919 ribosomal RNA gene database project: improved data processing and web-based tools. *Nucleic acids  
920 research*, 41, D590-D596.

921           Raikow, D. F., Sarnelle, O., Wilson, A. E., & Hamilton, S.K. (2004). Dominance of the noxious  
922 cyanobacterium *Microcystis aeruginosa* in low-nutrient lakes is associated with exotic zebra mussels.  
923 *Limnology and Oceanography*, 49, 482-487.

924           Rambaut, A. (2012). FigTree v1. 4. Available via <http://tree.bio.ed.ac.uk/software/figtree/>.  
925 Accessed Dec 12, 2018.

926           Richter, M., & Rosselló-Móra, R. (2009). Shifting the genomic gold standard for the prokaryotic  
927 species definition. *Proceedings of the National Academy of Sciences*, 106, 19126-19131.

928           Reynolds, C. S. (2006). The ecology of phytoplankton. Cambridge University Press.

929           Rocap, G., Larimer, F. W., Lamerdin, J., Malfatti, S., Chain, P., Ahlgren, N. A. et al. (2003).  
930 Genome divergence in two *Prochlorococcus* ecotypes reflects oceanic niche differentiation. *Nature*, 424,  
931 1042-1047.

932           Rohrback, T., Henning, M., & Kohl, J. G. (1999). Mechanisms of the inhibitory effect of the  
933 cyanobacterium *Microcystis aeruginosa* on *Daphnia galeata*'s ingestion rate. *Journal of Plankton  
934 Research*, 21, 1489-1500.

935           Sahm, A., Berns, M., Platzer, M., & Szafranski, K. (2017). PosiGene: Automated and easy-to-use  
936 pipeline for genome-wide detection of positively selected genes. *Nucleic Acids Research*, 45, 1-11.

937           Sarnelle, O., Morrison, J., Kaul, R., Horst, G., Wandell, H. & Bednarz, R. (2010). Citizen  
938 monitoring: Testing hypotheses about the interactive influences of eutrophication and mussel invasion on  
939 a cyanobacterial toxin in lakes. *Water Research*, 44, 141-150.

940           Sarnelle, O., Wilson, A. E., Hamilton, S. K., Knoll, L. B., & Raikow, D.F. (2005). Complex  
941 interactions between the zebra mussel, *Dreissena polymorpha*, and the harmful phytoplankton, *Microcystis  
942 aeruginosa*. *Limnology and Oceanography*, 50, 896-904.

943           Sarnelle, O. (1992). Contrasting effects of *Daphnia* ratios of nitrogen to phosphorus in a  
944 eutrophic, hard-water lake. *Limnology and Oceanography* 37:1527-1542.

945           Schloss, P. D., Westcott, S. L., Ryabin, T., Hall, J. R., Hartmann, M., Hollister, E. B. et al.  
946 (2009). Introducing mothur: open-source, platform-independent, community-supported software for

947 describing and comparing microbial communities. *Applied and Environmental Microbiology*, 75, 7537-  
948 7541.

949 Sebastian, M., & Ammerman, J. W. (2009). The alkaline phosphatase PhoX is more widely  
950 distributed in marine bacteria than the classical PhoA. *ISME Journal*, 3, 563.

951 Seip, K. L., & Reynolds, C. S. (1995). Phytoplankton functional attributes along trophic gradient  
952 and season. *Limnology and Oceanography*, 40, 589-597.

953 Seymour, J. R., Amin, S. A., Raina, J., & Stocker, R. (2017). Zooming in on the phycosphere: the  
954 ecological interface for phytoplankton–bacteria relationships. *Nature Microbiology*, 2, 17065.

955 Shapiro, B. J., & Polz, M. F. (2014). Ordering microbial diversity into ecologically and  
956 genetically cohesive units. *Trends in Microbiology*, 22, 235 – 247.

957 Smith, R. J. (2017). Solutions for loss of information in high-beta-diversity community data.  
958 *Methods in Ecology and Evolution*, 8, 68-74.

959 Soranno, P. A., Bacon, L. C., Beauchene, M., Bednar, K. E., Bissell, E. G., Boudreau, C. K., Boyer  
960 M. G. et al. (2017). LAGOS-NE: A multi-scaled geospatial and temporal database of lake ecological  
961 context and water quality for thousands of US lakes. *GigaScience*, 6, 12.

962 Stamatakis, A. (2006). RAxML-VI-HPC: Maximum likelihood-based phylogenetic analyses with  
963 thousands of taxa and mixed models. *Bioinformatics*, 22, 2688-2690.

964 Steffen, M. M., et al. (2017). Ecophysiological examination of the Lake Erie *Microcystis* bloom in  
965 2014: Linkages between biology and the water supply shutdown of Toledo, OH. *Environmental Science  
966 and Technology*, 51, 6745-6755.

967 Sterner, R. W. (2008). On the phosphorus limitation paradigm of lakes. *International Review of  
968 Hydrology*, 93, 433-445.

969 Strom, S., Bright, K., Fredrickson, K., & Brahamsha, B. (2017). The *Synechococcus* cell surface  
970 protein SwmA increases vulnerability to predation by flagellates and ciliates. *Limnology and  
971 Oceanography*, 62, 784-794.

972 Tanabe, Y., Okazaki, Y., Yoshida, M., Matsuura, H., Kai, A., Shiratori, T., Ishida, K., Nakano, S.,  
973 & Watanabe, M. M. (2015). A novel alphaproteobacterial ectosymbiont promotes the growth of the  
974 hydrocarbon-rich green alga *Botryococcus braunii*. *Scientific Reports*, 5, 10467.

975 Treangen, T. J., & Salzberg, S. L. (2012). Repetitive DNA and next-generation sequencing:  
976 computational challenges and solutions. *Nature Reviews Genetics*, 13, 36-46.

977 Turnbaugh, P. J., Hamady, M., Yatsunenko, T., Cantarel, B. L., Duncan, A., Ley, R. E., Sogin, M.  
978 L., Jones, W. J., Roe, B. A., Affourtit, J. P., Egholm, M., Henrissat, B., Heath, A. C., Knight, R., &  
979 Gordon, J. I. (2009). A core gut microbiome in obese and lean twins. *Nature*, 457, 480-484.

980 Vimr, E. R. (1994). Microbial sialidases: Does bigger always mean better? *Trends in*  
981 *Microbiology*, 2, 271-277.

982 Vox, M., Hesselman, M. C., te Beek, T. A., van Passel, M. W. J., & Eyre-Walker, A. (2015).  
983 Rates of lateral genes transfer in prokaryotes: High but why? *Trends in Microbiology*, 23, 598-605.

984 Wang, Q., Garrity, G. M., Tiedje, J. M., & Cole, J. R. (2007). Naive Bayesian classifier for rapid  
985 assignment of rRNA sequences into the new bacterial taxonomy. *Applied and Environmental*  
986 *Microbiology*, 5261-5267.

987 Wehr, J. D., & Sheath, R. G. (2003). Freshwater algae of North America: Ecology and  
988 classification. Academic Press, San Deigo.

989 Wetzel, R. G. (2001). Limnology: Lake and River Ecosystems. Ed. 3. Elsevier Academic Press,  
990 San Diego, CA.

991 White, J. D., Kaul, R. B., Knoll, L. B., Wilson, A. E., & Sarnelle, O. (2011). Large variation in  
992 vulnerability to grazing within a population of the colonial phytoplankter, *Microcystis aeruginosa*.  
993 *Limnology and Oceanography*, 56, 1714-1724.

994 Whitham, T. G., Bailey, J. K., Schweitzer, J. A., Shuster, S. M., Bangert, R. K., LeRoy, C. J.,  
995 Lonsdorf, E. V., Allan, G. J., DiFazio, S. P., Potts, B. M., Fisher, D. G., Gehring, C. A., Lindroth, R. L.,  
996 Marks, J. C., Hart, S. C., Wimp, G. M., & Wooley, S. C. (2006). A framework for community and  
997 ecosystem genetics: From genes to ecosystems. *Nature Reviews Genetics*, 7, 510-523.

998 Wilson, A. E., Kaul, R. B., & Sarnelle, O. (2010). Growth rate consequences of coloniality in a  
999 harmful phytoplankter. *PLoS ONE*, 5, e8679.

1000 Wilson, A. E., Sarnelle, O., Neilan, B. A., Salmon, T. P., Gehringer, M. M., & Hay M. E. (2005).  
1001 Genetic variation of the bloom-forming cyanobacterium *Microcystis aeruginosa* within and among lakes:  
1002 Implications for harmful algal blooms. *Applied and Environmental Microbiology*, 71, 6126-6133.

1003 Wilson, A. E., W. A. Wilson, & Hay, M. E. (2006). Intraspecific variation in growth and  
1004 morphology of the bloom-forming cyanobacterium *Microcystis aeruginosa*. *Applied & Environmental*  
1005 *Microbiology*, 72, 7386-7389.

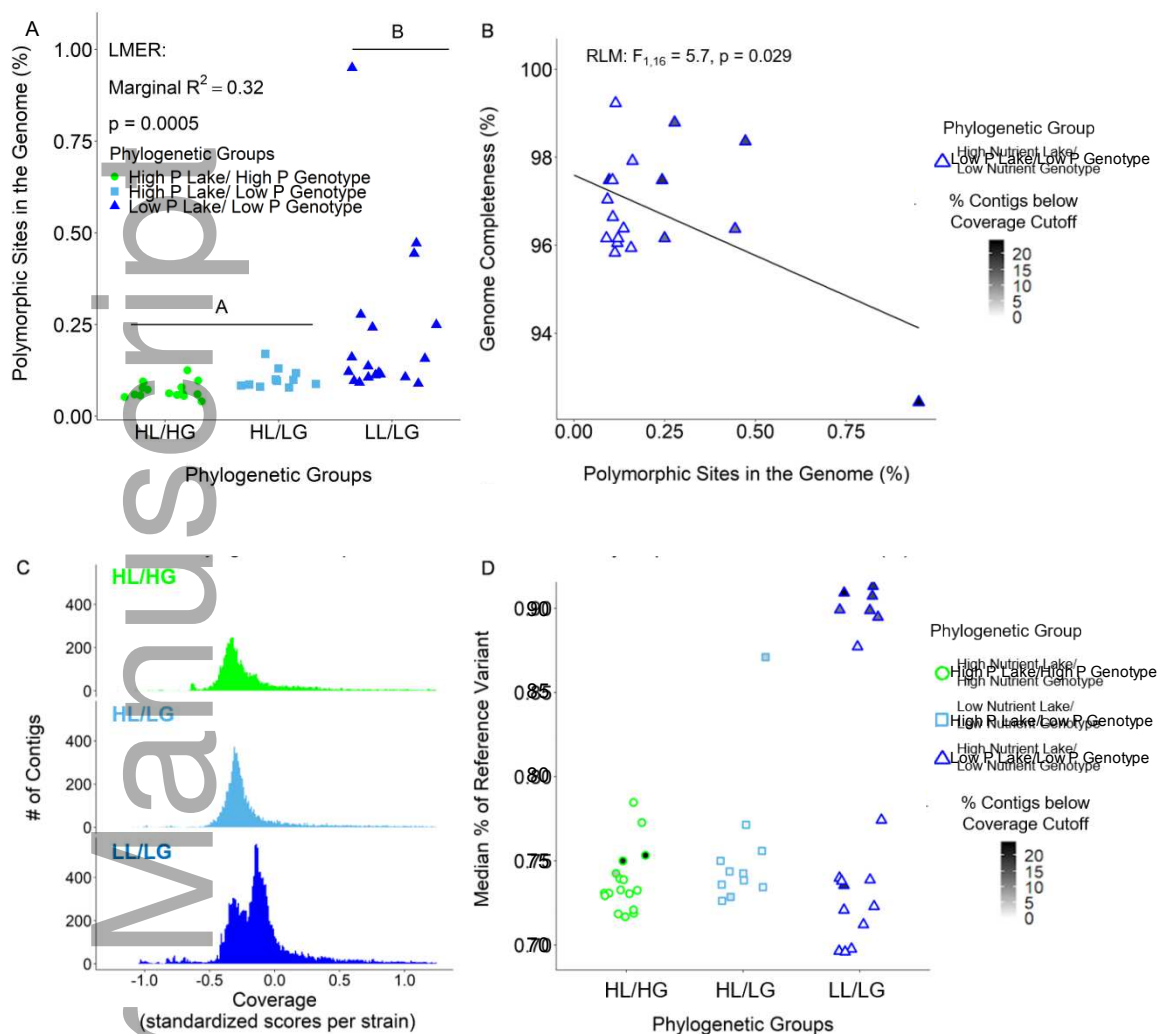
1006 Xiao, M., Li, M., & Reynolds, C. S. (2018). Colony formation in the cyanobacterium *Microcystis*.  
1007 *Biological Reviews*, 93, 1399-1420.

1008 Yanofsky, C., Cox E. C., & Horn, V. (1966). The unusual mutagenic specificity of an *E. coli*  
1009 mutator gene. *Proceedings of the National Academy of Sciences*, 55, 274-281.

1010 Zippel, B., & Neu, T. R. (2011). Characterization of glycoconjugates of extracellular polymeric  
1011 substances in tufa-associated biofilms by using fluorescence lectin-binding analysis. *Applied and*  
1012 *Environmental Microbiology*, 77, 505-516.

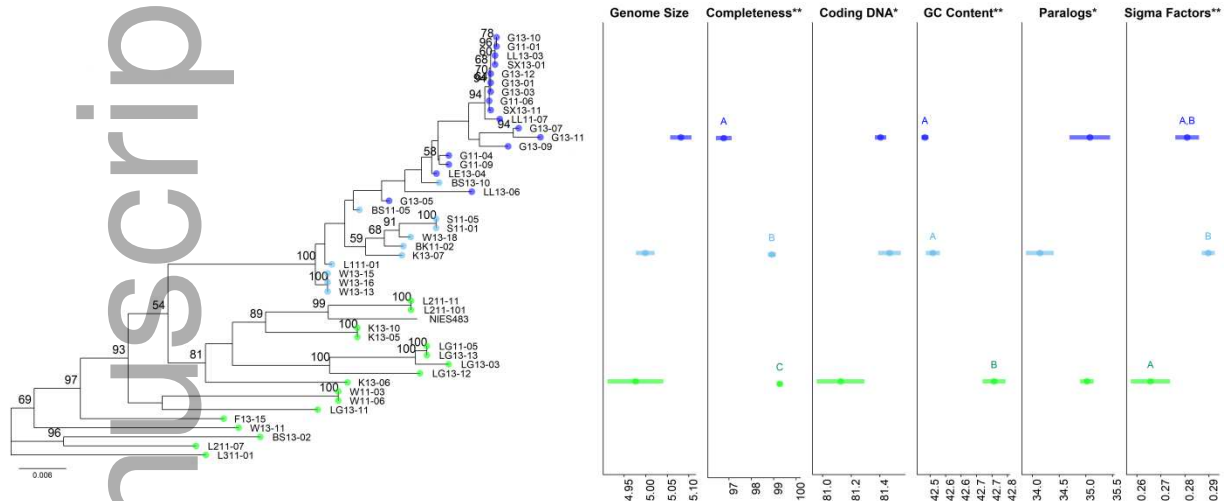
1013  
1014 **Data Accessibility:**  
1015 - Metagenomes sequences: NCBI SRA study # SRP092358, accession # SRR9658105 - SRR9658151.  
1016 - 16S Sequences: NCBI SRA project # PRJNA351875, accession # SRX6419375 - SRX6419329.  
1017 - Metadata and analysis scripts: [https://github.com/sjackrel/Microcystis\\_Inland\\_Lake\\_Genome\\_Evolution](https://github.com/sjackrel/Microcystis_Inland_Lake_Genome_Evolution).  
1018  
1019 **Author Contributions:**  
1020 S.L.J. analyzed all sequencing data; J.D.W. collected all *M. aeruginosa* samples and field metadata,  
1021 maintained lab cultures, performed growth rate experiments, and extracted DNA; J.T.E. created a custom  
1022 genome annotation pipeline; K.B. and K.H. extracted *M. aeruginosa* bins from metagenome datasets and  
1023 evaluated experimental tools; J.D.W., O.S. and V.J.D. designed study; S.L.J. and V.J.D wrote the  
1024 manuscript; J.D.W. and O.S. contributed to editing the manuscript.  
1025 **FIGURE 1** Population genomic analysis of *Microcystis aeruginosa* sequence bins. (A) Polymorphic sites,  
1026 including single nucleotide variants, insertions, and deletions, are more common in genomes of the Low  
1027 Phosphorus Lake/Low Phosphorus Genotype (LL/LG) phylogenetic groups, suggesting non-clonal  
1028 heterogeneity. (B) Within the LL/LG group, frequency of polymorphic sites within the genome was  
1029 inversely correlated with genome completeness estimates by checkM. Note we used robust linear  
1030 regression to ensure that statistical significance was not overly influenced by the single value in the  
1031 bottom right-hand corner, however excluding this point entirely renders the association insignificant. (C)  
1032 Extensive polymorphisms due to isolate heterogeneity lead to many low-coverage contigs representing  
1033 alternate assembly paths in LL/LG genomes and was visually evident as a bimodal distribution of contig  
1034 coverage. X-axis is a standardized Z-score with  $\mu = 0$ , s.d. = 1. (D) This frequent division of a single  
1035 contig into two separate contigs when extensive polymorphism occurred caused an increase in the  
1036 genome-wide percent occurrence of the reference versus alternate allele variant in the main sequence bin,  
1037 as sequence reads carrying the alternate nucleotide were now no longer aligned to the main sequence bin  
1038 contigs but to the corresponding low coverage contigs. The Y-axis sums forward (F) and reverse (R)  
1039 strands using the equation:  $(F+R \text{ of Reference})/[(F+R \text{ of Reference})+(F+R \text{ of Alternate})]$ .



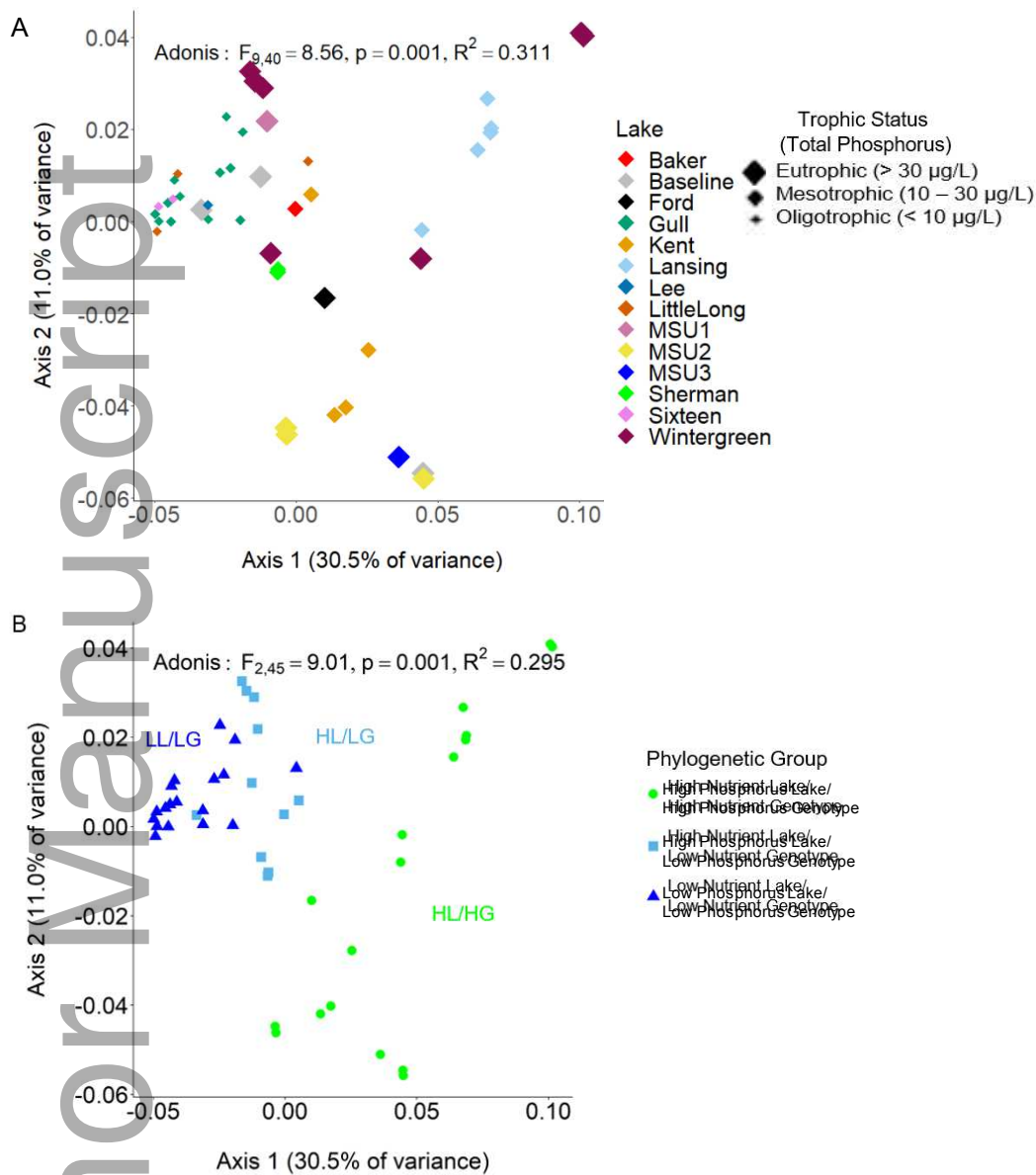


1040  
 1041 **FIGURE 2** Divergent genome structure across a phylogeny of 46 isolates of *Microcystis aeruginosa*  
 1042 collected from 14 inland lakes in Michigan, USA. Multi-locus sequencing typing was used to infer  
 1043 evolutionary history with RAxML based on five concatenated housekeeping genes (FtsZ, glnA, gltX, gyrB  
 1044 and pgi). Dark blue: isolates from oligotrophic lakes ('Low Phosphorus Lake, Low Phosphorus Genotype  
 1045 LL/LG'); light blue: isolates from phosphorus-rich lakes, but related to oligotrophic isolates ('High  
 1046 Phosphorus Lake/Low Phosphorus Genotype, HL/LG'); green: isolates from phosphorus-rich lakes  
 1047 ('High Phosphorus Lake/High Phosphorus Genotype, HL/HG'). All significant trends, as determined  
 1048 using linear mixed effects models that control for collection date and lake of origin, are noted with one  
 1049 asterisk at the  $p < 0.10$  level and two asterisks at the  $p < 0.05$  level. Values for each metric are shown for  
 1050 each phylogenetic group ( $\mu \pm 1$  S.E.). See Fig. S12 for values for each strain. Except for genome size,  
 1051 which is shown in megabases, all metrics are percentage data. Note that genome size, completeness, and  
 1052 GC content consider all contigs, regardless of length, while coding DNA, paralogs, and sigma factors as a

1053 percentage of total genes considers only contigs 2 kb in length and longer. Significance of post-hoc  
 1054 pairwise comparisons are noted with lettering, where groups sharing the same letter do not significantly  
 1055 differ from each other. Nineteen of the 20 publicly available sequences collected worldwide were most  
 1056 closely related to the HL/HG group (Fig. S1).



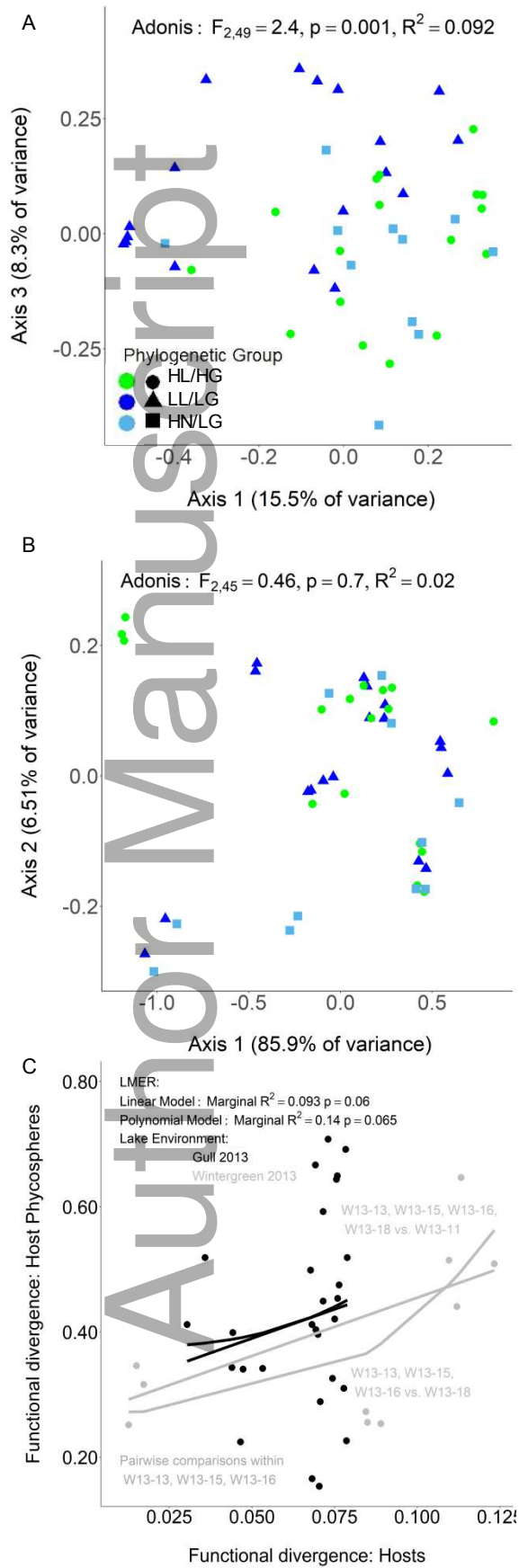
**FIGURE 3** Functional analysis of *M. aeruginosa* genomes. (A) Isolates cluster based on genome-wide protein function according to lake of origin and corresponding lake trophic status. Note that all LL/LG isolates originated from oligotrophic lakes that we defined as TP < 10 µg/L, while HL/HG and HL/LG isolates, which frequently co-occurred, originated from mesotrophic and eutrophic lakes that we defined as TP ≥ 10 µg/L. (B) Isolates also cluster in genome-wide protein functions based on the phylogenetic groups shown in Fig. 2. All genes within each genome were assigned to protein families, with points shown closer in principal coordinate space sharing more similarity in protein family composition. Significance of separation was determined using analysis of variance on distance matrices, i.e. adonis.



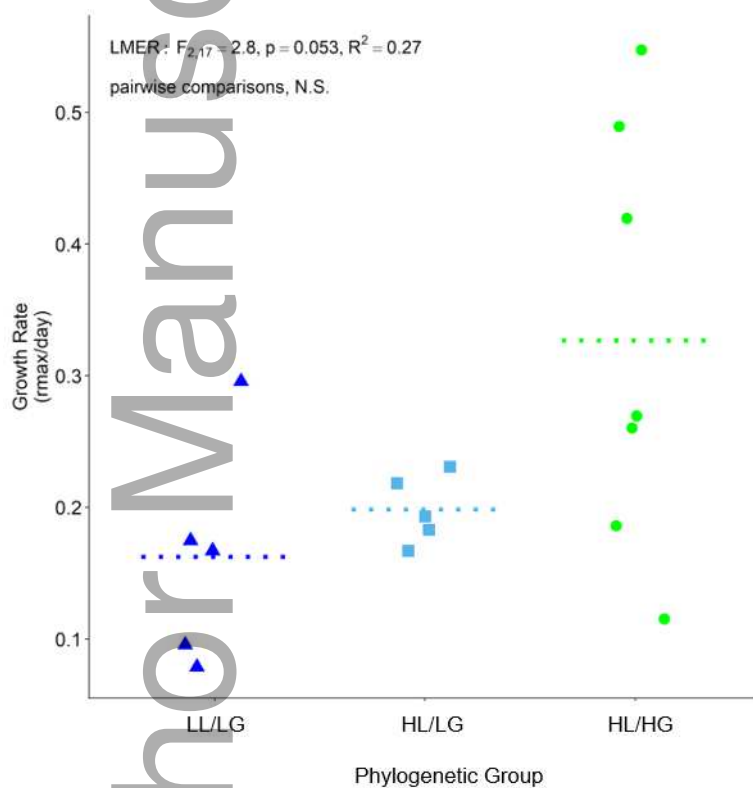
1057 **FIGURE 4** Taxonomic and functional analysis of *Microcystis aeruginosa*-associated microbiomes. (A)  
 1058 Isolates of *M. aeruginosa* belonging to different phylogenetic groups harbored taxonomically different  
 1059 communities of phycosphere bacteria. (B) Despite taxonomic differences, phycosphere communities  
 1060 associated with each of the three phylogenetic groups of *M. aeruginosa* tended to have similar protein  
 1061 functions. (C) When controlling for the local pool of phycosphere bacteria, host genomes that were more  
 1062 functionally similar to each other tended to harbor more functionally similar phycosphere communities.  
 1063 The x- and y- axes list functional distance in terms of Bray-Curtis distances from principal coordinate  
 1064 analyses that used the number of genes annotated to different protein families. This analysis used a subset

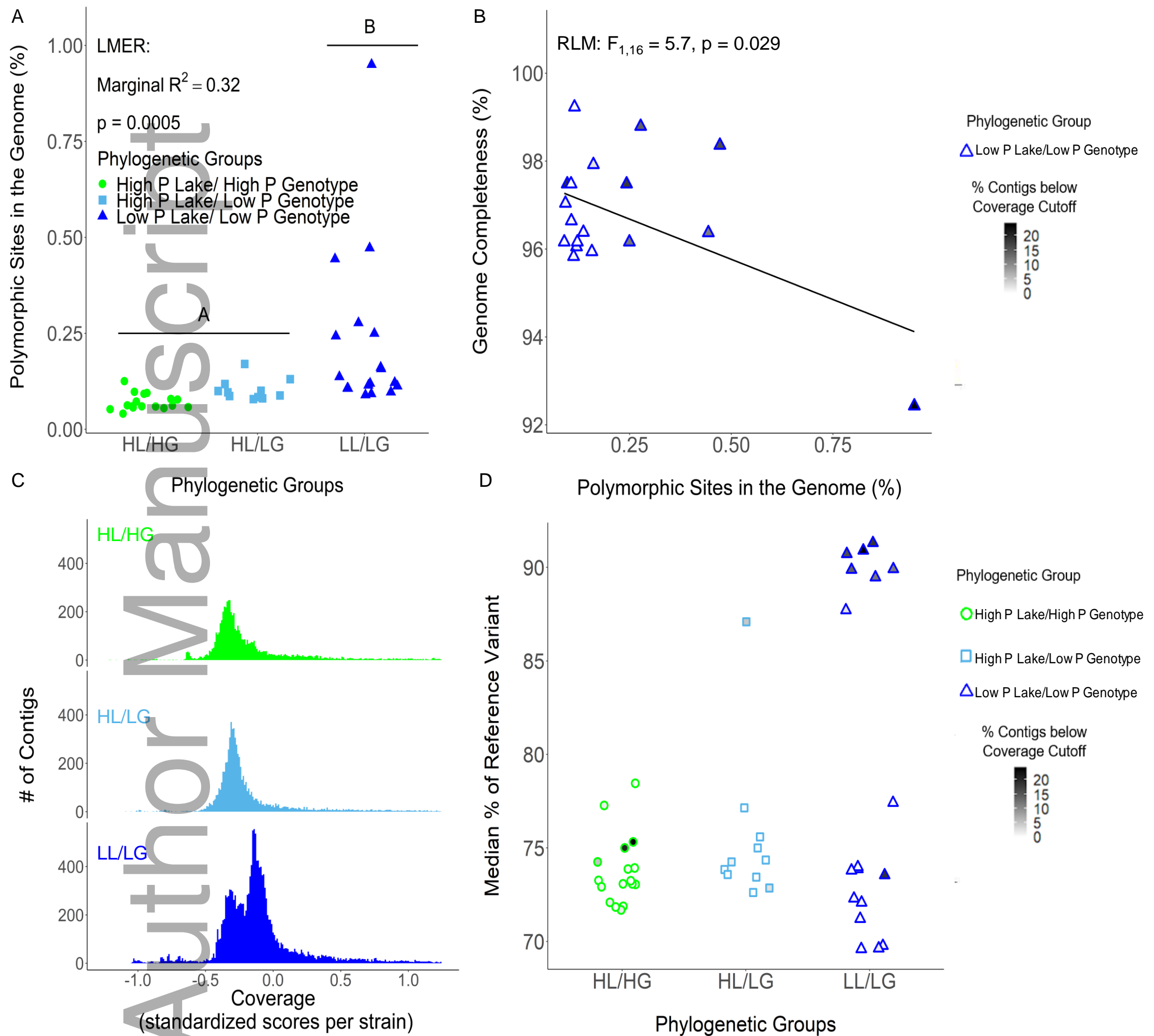
1065 of samples that were collected on the same day and from the same lake (5 isolates from eutrophic  
1066 Wintergreen Lake, 8 isolates from oligotrophic Gull Lake, respectively). Note that we show both the best  
1067 fitting linear model and polynomial model, as results were similar. Also note that taxonomic data using a  
1068 16S rRNA gene survey of bacterial taxa is shown in (A), while functional data shown in (B) and (C) uses  
1069 metagenome data in which all genes have been annotated to protein families to confer gene function.

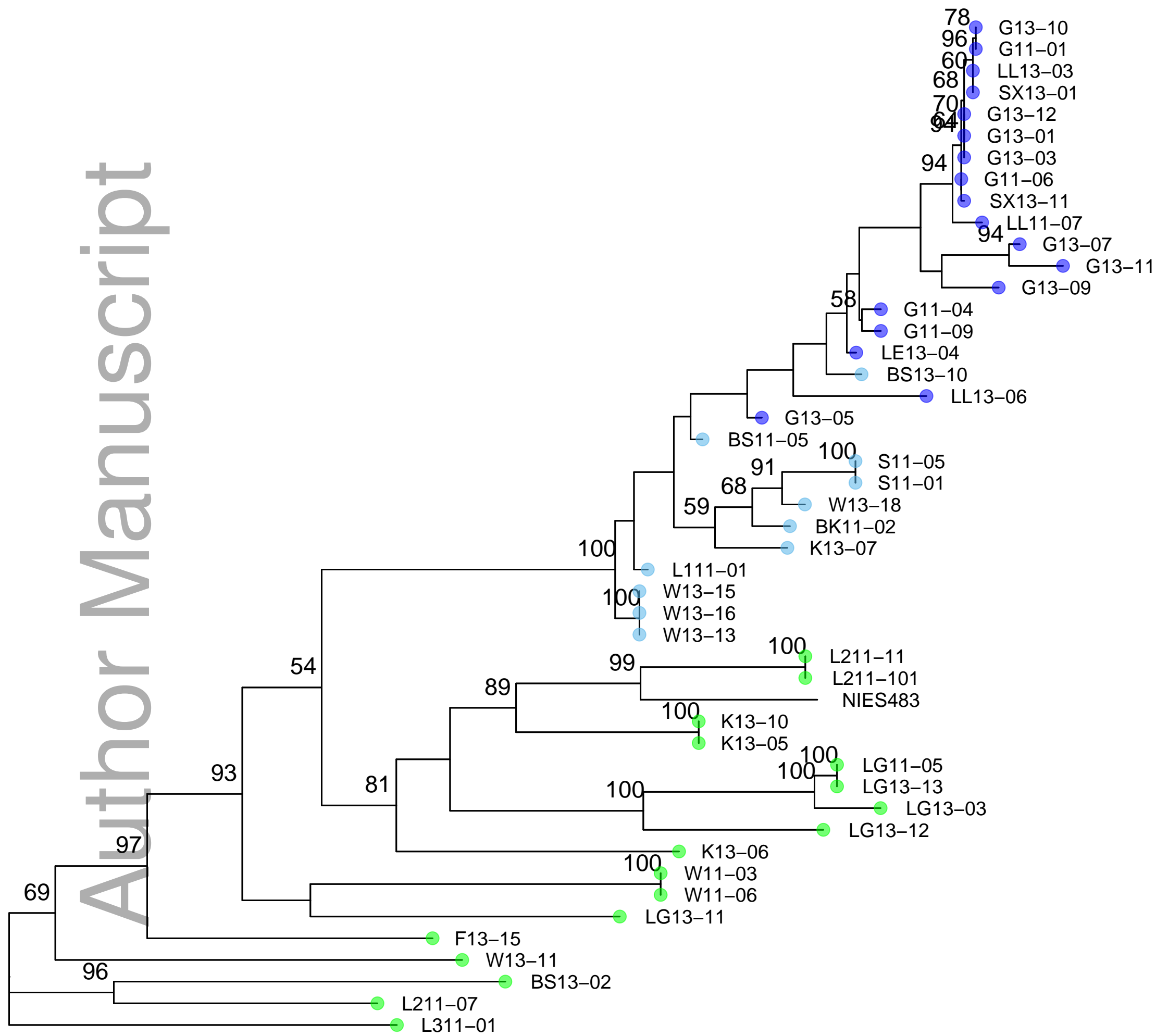
Author Manuscript



1070 **FIGURE 5** Growth rate measurement of *M. aeruginosa* isolates. Isolates of *Microcystis aeruginosa*  
1071 belonging to the Low Phosphorus Lake/Low Phosphorus Genotype (LL/LG) phylogenetic group grew  
1072 slower than isolates belonging to the High Phosphorus Lake/High Phosphorus Genotype (HL/HG)  
1073 phylogenetic group. Isolates from the High Phosphorus Lake/Low Phosphorus Genotype (HL/LG) group,  
1074 which originated from phosphorus-rich lakes but had a genomic architecture more closely resembling  
1075 isolates in oligotrophic lakes, grew at intermediate rates. A linear mixed-effects model controls for Lake as  
1076 a random effect to account for multiple isolates originating from a single lake. Group means illustrated  
1077 with a dashed line.







0.006  
This article is protected by copyright. All rights reserved

