1	
2	MS. DANIELLE RIVERA (Orcid ID : 0000-0002-9100-9945)
3	MR. IVAN PRATES (Orcid ID : 0000-0001-6314-8852)
4	DR. THOMAS J FIRNENO (Orcid ID : 0000-0002-4975-2794)
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10	Phylogenomics, introgression, and demographic history of South American true toads
11	(Rhinella)
12 13	Danielle Rivera ^{1,2} , Ivan Prates ³ , Thomas J. Firneno Jr ^{1,2} , Miguel Trefaut Rodrigues ⁴ , Janalee P. Caldwell ⁵ , Matthew K Fujita ^{1,2}
14	¹ Department of Biology, University of Texas at Arlington, Arlington, TX, USA
15	² Amphibian and Reptile Diversity Research Center, University of Texas at Arlington, TX, USA.
16	³ Department of Ecology and Evolutionary Biology and Museum of Zoology, University of
17	Michigan, Ann Arbor, MI, USA.
18	⁴ Departamento de Zoologia, Instituto de Biociências, Universidade de São Paulo, São Paulo,
19	SP, Brazil.
20	⁵ Sam Noble Museum & Department of Biology, University of Oklahoma, Norman, Oklahoma,
21	73072-7029, USA

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22 Correspondence

23 Danielle Rivera, Department of Biology, University of Texas at Arlington, Arlington, TX, USA.

24 Email: danielle.rivera@uta.edu

25 Running Head: Phylogenomics within *Rhinella marina*

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27 Abstract

28 The effects of genetic introgression on species boundaries and how they affect species' integrity 29 and persistence over evolutionary time have received increased attention. The increasing 30 availability of genomic data has revealed contrasting patterns of gene flow across genomic 31 regions, which impose challenges to inferences of evolutionary relationships and of patterns of 32 genetic admixture across lineages. By characterizing patterns of variation across thousands of 33 genomic loci in a widespread complex of true toads (*Rhinella*), we assess the true extent of 34 genetic introgression across species thought to hybridize to extreme degrees based on natural 35 history observations and multi-locus analyses. Comprehensive geographic sampling of five 36 large-ranged Neotropical taxa revealed multiple distinct evolutionary lineages that span large 37 geographic areas and, at times, distinct biomes. The inferred major clades and genetic clusters 38 largely correspond to currently recognized taxa; however, we also found evidence of cryptic 39 diversity within taxa. While previous phylogenetic studies revealed extensive mito-nuclear 40 discordance, our genetic clustering analyses uncovered several admixed individuals within major 41 genetic groups. Accordingly, historical demographic analyses supported that the evolutionary 42 history of these toads involved cross-taxon gene flow both at ancient and recent times. Lastly, 43 ABBA-BABA tests revealed widespread allele sharing across species boundaries, a pattern that 44 can be confidently attributed to genetic introgression as opposed to incomplete lineage sorting. 45 These results confirm previous assertions that the evolutionary history of *Rhinella* was 46 characterized by various levels of hybridization even across environmentally heterogeneous 47 regions, posing exciting questions about what factors prevent complete fusion of diverging yet 48 highly interdependent evolutionary lineages.

49 1 INTRODUCTION

50 How introgression affects reproductive isolation and speciation is an enduring question in 51 evolutionary biology. Reproductive isolation has long been viewed as the primary factor behind 52 lineage divergence and stable boundaries between closely related species (Avise et al., 1998; 53 Mayr, 1963; Rabosky, 2016). When closely related populations come into contact, however, 54 gene flow via hybridization can lead to the introgression of alleles (Mallet, 2005; O'Connell et 55 al., 2021). Introgression levels can vary starkly across genome regions. In particular, in the 56 presence of strong divergent selection, those loci underlying adaptive phenotypes can maintain 57 marked differentiation even with extensive gene flow among closely related populations (Feder 58 et al., 2012). Thus, these varying degrees of isolation across the genome may contribute to the 59 maintenance of species boundaries despite the homogenizing effects of gene flow (Yeaman & 60 Whitlock, 2011).

61 Differential introgression across genomic regions can lead to dramatic topological 62 discordance between genealogies inferred from distinct genes, as illustrated by instances of mitonuclear discordance (Bernardo et al., 2019; Bessa-Silva et al., 2020; Firneno et al., 2020). This 63 64 gene-tree heterogeneity must be accounted for as it can make reconstructing evolutionary 65 relationships and historical demography challenging (Carstens & Knowles, 2007; Firneno et al., 66 2020; Liu et al., 2010). The increasing availability of high-throughput sequencing datasets for 67 non-model organisms has improved our ability to discern patterns of introgression in closely 68 related species or populations (Firneno et al., 2020; Graham et al., 2018; Lavretsky et al., 2016) 69 and thus clarify phylogenetic relationships and species limits. This is especially so in large, 70 widely distributed species complexes with limited variation in external morphological traits and 71 hybridization blurring species limits (Guo et al., 2016; Phuong et al., 2017; Potter et al., 2016).

The increasing availability of genome-scale datasets has also fostered the development of model-based approaches to infer historical demographic events such as population size shifts and pulses of gene flow (Portik et al., 2017; Prates, Xue et al., 2016). These approaches have transformed our understanding of how landscape and climate changes have contributed to the assembly of regional species pools, for instance by limiting dispersal, promoting speciation, or

77 leading to lineage fusion (Graham et al., 2018; Lavretsky et al., 2016; Leaché et al., 2019; Portik 78 et al., 2017). One flexible approach involves simulating population histories to compare the fit of 79 empirical genome-scale data to data simulated under alternative biogeographical scenarios (Dal 80 Vechio et al., 2019; Portik et al., 2017; Prates et al., 2016). This modeling framework can 81 facilitate hypothesis testing, such as how climate-driven habitat shifts may have led to migration, 82 introgression, or isolation across geographic regions. These approaches have been instrumental 83 to shed light on the historical factors behind present-day spatial biodiversity patterns in regions 84 that concentrate large proportions of biodiversity. This is the case of the Neotropics, where 85 demographic inference has supported that Late-Quaternary climate fluctuations and Neogene 86 geomorphological change have played a major role in shaping species range limits, genetic 87 diversity levels, and lineage divergence (Gehara et al., 2017; Pirani et al., 2020; Prates, Xue et 88 al., 2016). Nevertheless, biogeographic investigations in the Neotropics have often shown 89 geographic and taxonomic bias, which questions the generality of the mechanisms invoked to 90 explain species richness and distributions. For instance, taxa with wide ranges across South 91 America's open vegetation biomes – the dry and highly seasonal Cerrado, Caatinga, and Chaco – 92 have received relatively less attention than rainforest biotas (Fonseca et al., 2018; Gehara et al., 93 2017; Werneck, 2011).

94 One example of a Neotropical clade whose biogeographic history remains poorly known 95 is the true South American toads, genus Rhinella (Bufonidae). Despite being the focus of a 96 handful of phylogeographic studies, the evolutionary relationships and species limits between 97 these toads remain elusive, perhaps due to wildly varying patterns of introgression and 98 hybridization across species (Maciel et al., 2010; Pereyra et al., 2016; Pereyra et al., 2021; 99 Sequeira et al., 2011; Vallinoto et al., 2009). As such, not only the evolutionary history of this 100 group is unclear, but so are the environmental and geographic factors that may have favored 101 introgression and its variation, or how hybridization may have contributed to lineage divergence 102 or fusion (Azevedo et al., 2003; Correa et al., 2012; Malone & Fontenot, 2008; Pereyra et al., 103 2016; Sequeira et al., 2011). *Rhinella* is composed of multiple species complexes that are each 104 distributed across much of the Neotropics. These groups are known to harbor high levels of 105 cryptic lineage diversity, as revealed by single and multi-locus genetic analyses (Maciel et al.,

106 2010; Pereyra et al. 2016; Pereyra et al. 2021; Vallinoto et al., 2009). Among them is the 107 Rhinella marina group, best known for the globally invasive species R. marina. Previous studies 108 of this group have identified both mitochondrial and nuclear introgression across species 109 (Azevedo et al., 2003; Maciel et al., 2010; Vallinoto et al., 2009). However, lack of data about 110 persisting genetically admixed populations in the wild makes it difficult to assess the magnitude 111 of presumed hybridization and how it affects species boundaries (Azevedo et al., 2003; Malone 112 & Fontenot, 2008; Pereyra et al. 2021). Despite the ecological diversity seen in *Rhinella*, with 113 taxa that span savannas, rainforests, and xeric shrublands, biogeographic analyses have largely 114 focused on taxa occurring within a single biome (Sequeira et al., 2011; Thomé et al., 2010), 115 which is also the case of other South American anuran clades (Fonseca et al., 2018; Gehara et al., 116 2017; Oliveira et al., 2018). As a result, how habitat transitions may contribute to patterns of 117 gene flow and species range limits remains unclear.

118 In this investigation, we focus on the *R. marina* group to investigate evolutionary 119 relationships, quantify the extent of hybridization, and examine whether landscape transitions 120 among South America's biomes impose limits to gene flow and species ranges. For this purpose, 121 we focus on *R. marina*, *R. poeppigii*, *R. horribilis*, *R. jimi*, and *R. schneideri* (also known as *R.* 122 *diptycha*), which have established contact zones throughout the continent. We infer population 123 structure, gene flow, and relationships based on geographically comprehensive sampling of 124 genomic variation within each taxon. We then proceed to test alternative historical hypotheses to 125 quantify plausible demographic events such as population size shifts and historical gene flow. 126 With this approach, we seek to answer the following questions: what are the levels of genetic 127 structure across and within each species? Do genomic data corroborate a pattern of widespread 128 admixture or introgression across these species, as previously suggested based on only a few 129 loci? Lastly, what historical demographic processes may explain species distributions and 130 genetic diversity patterns within this clade?

131

132 2 MATERIAL AND METHODS

133

134 **2.1 Sample collection**

- 135 Our sampling included 191 individuals belonging to the *Rhinella marina* species group, as
- 136 follows: 72 R. marina, 39 R. schneideri, 23 R. horribilis, 11 R. jimi, and nine R. cf. poeppigii, as
- 137 well as four *R. veredas*, eight *R. rubescens*, and 25 *R. icterica* which were used in the
- 138 mitochondrial analysis only. We also included samples from the *Rhinella granulosa* and *R*.
- 139 *margaritifera* major clades within *Rhinella* as outgroups in the divergence time estimation
- 140 analyses (see below). Within each species in the focal group consisting of *R. marina*, *R.*
- 141 *horribilis*, *R. schneideri*, *R. jimi*, and *R. poeppigii*, we sample multiple individuals from each
- 142 locality across their known ranges, with the exception of *R. cf. poeppigii*, which was identified as
- 143 distinct from *R. marina a posteriori* based on the genetic data (see Results)(Acevedo, et al. 2016;
- 144 Maciel, et al., 2010; Vallinoto, et al. 2010). Tissue samples were obtained from the MTR
- 145 herpetological tissue collection hosted at Instituto de Biociências, University of São Paulo
- 146 (IBUSP) with vouchers at Museum of Zoology, University of São Paulo, as well as from the
- 147 Amphibian and Reptile Diversity Research Center (ARDRC), and the Louisiana State University
- 148 Museum of Natural Science (LSUMNS)(Table S1).
- 149

150 2.2 DNA extraction, amplification, & sequencing

151 We extracted genomic DNA using a standard phenol-chloroform extraction protocol (Sambrook

152 & Russell, 2006). Fragments of the mitochondrial 16S were amplified using 16Sar and 16Sbr

- 153 primers and sequenced on an ABI 3730xL (Primer information and PCR conditions in the
- 154 Supplementary Text S1). Sequences were edited and aligned in Geneious Prime 2020.0.4
- 155 (Identification and Accession numbers in Supplementary Table S1). We generated double-digest
- 156 restriction-site associated DNA sequencing (ddRADseq) data following (Peterson, et al., 2012),
- 157 with modifications as described in Streicher et al. (2014). Briefly, 200-500 ng of DNA were
- 158 digested using the *Sbf*I (restriction site 5'-CCTGCAGG-3') and *Msp*I (restriction site
- 159 5'-CCGG-3') restriction enzymes in a single reaction using the manufacturer's recommended
- 160 buffer (New England Biolabs) for 5 hr at 37°C. Digested DNA was bead-purified before ligating
- 161 barcodes and index adaptors, then samples with the same index were pooled and size-selected

162 (415-515 bp) on a Blue Pippin Prep size selector (Sage Science). Final Library preparation was

- 163 analyzed and quantified on a BioAnalyzer (Agilent) and Qubit Fluorometer 4 (Thermo Fisher
- 164 Scientific). The resulting 100 bp single-end libraries were sequenced at MedGenome on an
- 165 Illumina HiSeq2500.

166 We used the command line version of ipyrad v. 0.9.45 (Eaton & Overcast, 2020) 167 (available at https://ipyrad.readthedocs.io) to de-multiplex and assign reads to individuals based on sequence barcodes (allowing no mismatches from individual barcodes), perform reference 168 169 read assembly (minimum clustering similarity threshold = 0.90), align reads into loci, and call 170 single nucleotide polymorphisms (SNPs). As a reference, we used the *Rhinella marina* genome 171 (Edwards et al., 2018). A minimum Phred quality score (= 33), sequence coverage (= 6x), read 172 length (= 35 bp), and maximum proportion of heterozygous sites per locus (= 0.5) were enforced, 173 while ensuring that variable sites had no more than two alleles (i.e., a diploid genome). 174 Following the initial assembly, we used Matrix Condenser (de Medeiros & Farrell, 2018) to 175 assess levels of missing data across samples and then re-assembled our dataset to ensure a 176 minimum sample coverage of less than 35% missing loci within each sample and at least 75% of 177 samples at each locus. This strategy resulted in a final dataset composed of 49,376 SNPs at 3,318 178 RAD loci with less than 12% missing data. Additionally, Weir and Cockerham mean F_{ST} estimates for the ddRADseq dataset using VCFTools (Danecek et al., 2011) and Nei's G_{ST} for 179 180 the mitochondrial dataset were calculated using the R package mmod (Winter, 2012).

181

182 **2.3 Inferring population structure and genetic admixture**

183 Based on the ddRAD data, we used a genetic clustering approach to estimate the number of

184 demes and if admixture was present among them. We assembled a SNP dataset as described

- above but excluding outgroups and using only one SNP per RAD locus to maximize sampling of
- 186 independent SNPs. This approach resulted in a dataset composed of 3,314 SNPs. Genetic
- 187 clustering was performed using the maximum likelihood method ADMIXTURE, testing up to 15
- populations with 20 replicates per K and a 10-fold cross-validation (Alexander, et al., 2009;
- 189 Portik, 2016). The best K was determined by assessing the replicate with the lowest cross-This article is protected by copyright. All rights reserved

190 validation error. To further characterize population structure, we used the non-parametric method

191 of discriminant analysis of principal components (DAPC), implemented in the R package

192 adegenet (Jombart & Ahmed, 2011; Jombart, et al., 2010). The *find.clusters* function was

193 used to test the fit of 1-15 clusters (K). The K with the lowest Bayesian information criterion

194 (BIC) score was considered the best-fit number of demes. The resulting ancestry coefficient

195 matrices (Q-matrices) were then imported into QGIS (QGIS Development Team 2020. QGIS

196 Geographic Information System. Open Source Geospatial Foundation Project.

http://qgis.osgeo.org) to make average-per-locality pie-charts indicating admixture levels at each
sampled locality for each species.

199

200 2.4 Phylogenetic Analyses

201 We reconstructed maximum likelihood phylogenies for both the mitochondrial and the unlinked

202 ddRADseq loci datasets using IQTREE v2.1.2, utilizing the built-in model selection tool

203 ModelFinder Plus, implementing 1000 ultrafast bootstraps (Hoang et al., 2018;

204 Kalyaanamoorthy et al., 2017; Nguyen et al., 2015). We specified that all partitions share the

same branch lengths and selected the best-fit partitioning scheme by merging partitions (which

206 implements the greedy algorithm of PartitionFinder), testing the "MrBayes" substitution model

set and considering the top 10% partition schemes using the fast relaxed clustering algorithm

from PartitionFinder2 to save computational time (Chernomor et al., 2016; Lanfear et al., 2012;

Lanfear et al., 2014; Lanfear et al., 2017). In addition, we performed phylogenetic inference

under a Bayesian framework for both datasets using MrBayes 3.2.6 (Ronquist et al. 2012),

211 implementing three independent runs of four Markov chains of 10 million generations each and

sampling every 1,000 generations with the first 25% generations discarded as burn-in. We used

213 Tracer 1.7 (Rambaut et al. 2018) to assess whether Markov chain mixing was adequate (effective

sample sizes > 200) and to visually assess model parameter stationarity and convergence

between runs. We then summarized a 50% majority-rule consensus tree.

216 To estimate divergence dates and inform the delimitation of species boundaries, we 217 conducted Bayesian divergence dating analyses based on the mtDNA dataset in BEAST2 using This article is protected by copyright. All rights reserved an HKY model of nucleotide substitution, a log-normal relaxed molecular clock, and a
coalescent constant population size prior. We follow Pramuk et al. (2008) by enforcing a
minimum age for the root node between the *Rhinella marina* and *R. granulosa* species
complexes based on a *Rhinella marina* fossil from the Clarendonian North American Stage of
the middle Miocene (ca. 11 mya), described by Sanchiz (1998), and employed a log-normally
distributed prior with a standard deviation of 0.5. We ran this analysis for 20 million generations

sampling every 1000 generations. Runs were assessed using TRACER v1.6 (Rambaut &

225 Drummond, 2009) to examine convergence. We then summarized a maximum clade credibility

tree using TreeAnnotator discarding the first 25% of trees as burn-in (Bouckaert et al., 2019;

227 Stamatakis, 2014). All phylogenetic tree-based methods were analyzed on Cipres (Miller et al.,

228 2010).

229

230 **2.5 Demographic modeling with** ∂a∂i

231 We used the diffusion-approximation method ∂a∂i (Gutenkunst, et al., 2009) to test alternative

hypotheses of population history within the *Rhinella marina* clade based on species ranges,

233 previously reported potential hybridization events, common biogeographic patterns exhibited by

amphibians across this region, as well as computational limitations. Using both two- and three-

dimensional joint site frequency spectra (2D- and 3D-JSFS), we divided the dataset into two

236 population subsets: one comprised of *R. marina, R. horribilis*, and *R. jimi*; and another

comprised of *R. schneideri* and *R. cf. poeppigii*. Folded-JSFS datasets were used in all $\partial a \partial i$ analyses.

We filtered the ddRAD data for each subset to allow no more than 35% missing data from any sample, removed singletons, and selected one SNP per locus using VCFtools (Danecek et al., 2011; Gutenkunst et al., 2009; Portik et al., 2017). We then used the *stacks_pipeline* Python script from Portik et al. (2017) to create the SNP input file for $\partial a \partial i$. We used the python script easySFS (https://github.com/isaacovercast/easySFS) to determine the projection size of each population, which was determined by balancing a downscaled sample size that maximized the number of segregating sites (Gutenkunst et al., 2009; Marth et al., 2004). In $\partial a \partial i$, we then

tested a range of extrapolation grid sizes (40-100 in 10-unit increments, e.g., 50, 60, 70 to 100,
110, 120) in the divergence-with-no-migration model to determine the appropriate grid size by
selecting the model with the highest log-likelihood, implementing 4 rounds of optimization
totaling 100 replicates. Once an optimal grid size was determined, each tested model was run 3
times independently.

251 For the subset composed of R. marina, R. horribilis, and R. jimi, we used a 3D-JSFS to 252 test models incorporating gene flow at different times, including those accounting for ancient 253 migration, recent secondary contact, and past simultaneous divergence of all lineages (Fig. S5). 254 In addition to a model of 1) divergence with no migration, we tested the following models: 2) 255 divergence with continuous symmetric gene flow between all populations; 3) divergence with 256 continuous symmetric gene flow between geographically adjacent populations; 4) isolation 257 followed by secondary contact; 5) simultaneous divergence in isolation followed by more recent 258 secondary contact between adjacent populations; 6) simultaneous divergence with continuous 259 symmetric migration between adjacent populations; 7) ancient migration with very recent 260 isolation; 8) ancient migration with a longer period of recent isolation; 9) a short ancient period 261 of migration followed by a long period of isolation; and 10) ancient migration followed by 262 lineage isolation and population size change across two epochs (Barratt et al., 2018; Portik et al., 263 2017).

264 For the subset composed of *R. schneideri* and *R. cf. poeppigii*, we tested 2D-JSFS models 265 incorporating differing migration levels at different time periods (Fig. S6). In addition to a model 266 of 1) divergence with no migration, we tested the following models: 2) divergence with 267 continuous symmetric migration; 3) divergence with continuous asymmetric migration; 4) 268 divergence with continuous symmetric migration and a varying rate of migration across two 269 epochs; 5) divergence with continuous asymmetric migration and a varying rate of migration 270 across two epochs; 6) divergence in isolation, followed by symmetric secondary contact; 7) 271 divergence in isolation, followed by asymmetric secondary contact; 8) ancient symmetric 272 migration then subsequent isolation; 9) ancient asymmetric migration then subsequent isolation; 273 10) divergence in isolation followed by symmetric secondary contact with subsequent isolation;

and 11) divergence in isolation followed by asymmetric secondary contact with subsequent
isolation (Charles et al., 2018; Portik et al., 2017).

Best-fit models were chosen based on log-likelihood values, which we assumed to be the true likelihood (and not composite likelihood) given that we have kept only one SNP per RAD locus. Replicates with the consistently highest likelihood scores were used to calculate and compare models using the Akaike information criterion (AIC).

280

281 **2.6 Inferring gene flow**

282 To further explore potential hybridization between taxa, we inferred Patterson's D statistic, or 283 the ABBA-BABA statistic, and the related admixture fraction estimates, or f_4 -ratio statistics, 284 based on the ddRAD data using Dsuite (Malinsky, et al., 2020; Patterson et al., 2012). Tests 285 were designed with a 4-taxon fixed phylogeny (((P1,P2)P3)O), wherein a typical ancestral ("A") 286 and derived ("B") allele pattern should follow BBAA. Under incomplete lineage sorting, 287 conflicting ABBA and BABA patterns should occur in equal frequencies, resulting in a D 288 statistic = 0. If, however, introgression between P3 and P1 or P2 has occurred, there should be an 289 excess of these patterns and a D statistic significantly different from 0, with significance detected 290 using a block-jackknifing approach (Durand, et al., 2011; Green et al., 2010; Malinsky et al., 291 2020; Patterson et al., 2012). We used the f-branch or $f_b(C)$ metric to tease apart potentially 292 correlated f_4 -ratio statistics and estimate gene flow events between internal branches on the 293 phylogeny (Malinsky et al., 2018; Martin et al., 2013). Dsuite uses a VCF file and a 294 jackknifing approach to assess correlations in allele frequencies between closely-related species 295 (Malinsky et al., 2020). Within Dsuite, we used the *Dtrios* and *Fbranch* programs to identify 296 introgression between all combinations of species, as well as potential direction of gene flow, 297 specifying *Rhinella veredas* as an outgroup and applying the Benjamini-Hochberg (BH) 298 correction to control for the false discovery rate.

299

300 **3 RESULTS**

301

302 **3.1 Phylogenetic relationships**

303 The 16S phylogeny (160 individuals; 481 base pairs) suggested little phylogenetic structure 304 within the *Rhinella marina* complex. One clade included most of the *R. horribilis* samples, while 305 individuals from the remaining taxa formed a polytomy (Fig. S1). Maximum likelihood and 306 Bayesian phylogenies based on the ddRADseq dataset (128 individuals) resulted in fully 307 concordant phylogenies (Fig. 1). These analyses inferred six highly supported clades, two 308 corresponding to *R. marina* and the other four corresponding to *Rhinella schneideri*, *R.* 309 horribilis, R. jimi, and a clade tentatively assigned to R. cf. poeppigii (BS = 100; PP = 1.0; Fig. 310 1-3). These putative R. poeppigii samples were originally identified as R. marina, which would 311 render *R. marina* to be paraphyletic; however, after re-examining these specimens 312 morphologically, we were able to positively identify samples from western Amazonia in Brazil's 313 state of Acre as *R. poeppigii*, while closely related samples from eastern localities in the state of 314 Pará were morphologically more similar to R. marina (Fig. S7). Pairwise Nei's G_{ST} estimates for 315 the 16S data were much lower than the Weir and Cockerham weighted F_{ST} estimates for the 316 ddRADseq data. Across all taxa, the average pairwise G_{ST} for the mitochondrial data was 0.117 317 (0.025-0.228) while the average pairwise F_{ST} for the nuclear data was 0.506 (0.379-0.843) (Table 318 S2).

319 The time-calibrated phylogeny based on the 16S mitochondrial data dated the root of 320 Rhinella marina at 5.58 mya (95% HPD: 2.75–9.40 million years ago; Fig 4). Though many 321 relationships had poor support due to lack of variability within the locus, some clades showed 322 high support, including a clade with most of the *R*. *horribilis* samples, which was dated at 1.03 323 mya (95% HPD: 0.35–2.04 mya). Two samples not included in this clade were samples 324 distributed in the northern Andes, which clustered with other *R. marina* samples (Fig 4). One 325 highly supported clade consisted of two clades with a divergence date of 0.67 mya (95% HPD: 326 0.18–1.48 mya): one clade consisted of Rhinella cf. poeppigii samples from eastern Amazonia, 327 and other consisted of the western Amazonia R. poeppigii sample as well as R. marina from 328 southern Amazonia (Fig. 4). Additionally, R. granulosa is estimated to be sister to the R. marina

329 complex, with *R. margaritifera* more distantly related. Due to the lack of variation within the *R.* 330 *marina* group, we interpret dates within this complex with caution.

331

332 3.2 Population structure

333 Despite the high posterior probabilities of each clade in our ddRAD tree, the ADMIXTURE 334 results supported genetic admixture both within and across multiple taxa within the *Rhinella* 335 *marina* complex (Fig. 1), with a best-fit K of 7. Each ddRAD clade corresponded to a cluster, 336 except for the Rhinella schneideri clade which consisted of two non-reciprocally monophyletic 337 units. *Rhinella horribilis* (blue, Fig. 1-2) showed admixture from the northern cluster of *R*. 338 marina into one northern Andes locality. One cluster of Rhinella marina was relegated to 339 northern Amazonia (light green, Fig. 1-2), while the other cluster showed a cline of admixture 340 across its western and southern Amazonia clades (light green to purple, Fig. 1-2) and admixture 341 from R. *jimi* and R. cf. poeppigii (dark green and orange, Fig. 1). The two genetic clusters within 342 R. schneideri (pink and yellow, Fig. 1,3) followed an east-west admixture gradient across the 343 Cerrado to the northern Atlantic Forest, as well as intermediate ecotones. Rhinella jimi occurs 344 mostly in the semi-arid Caatinga shrublands of northeastern Brazil, but also in the adjacent 345 coastal rainforest (dark green, Fig. 1-2). The DAPC analysis supported this clustering scheme as 346 well; however, BIC scores suggested similar support for six to eight clusters (Fig. S4). The seven 347 clusters recovered were concordant with phylogenetic structure (Fig. 1).

348

349 **3.3 Demographic inference**

350 For the subset composed of *R. marina, R. horribilis*, and *R. jimi*, the best 3D-JFSF model was

351 one that incorporated ancient migration with a short period of recent isolation since divergence,

with a log-likelihood of -1572.69 and AIC of 3165.38 (Fig. 5, Table S3). This model included an

ancient period of migration between all lineages (mA, Fig. 5), then another period of migration

between geographically adjacent species after the divergence between *R. marina* and *R. jimi*, and

355 then subsequent lineage isolation. Parameter estimates indicated a much longer ancient period of

migration between all lineages with smaller migration rates (T1 = 10.82; mA = 0.05) compared to the shorter time of adjacent-species migration with higher rates of migration (T2 = 0.12; m1 = 1.36; m2 = 0.85) and the shortest period of isolation (T3 = 0.10) (Table S3).

For the subset composed of *R. schneideri* and *R. cf. poeppigii*, the best 2D-JFSF model incorporated divergence in isolation followed by secondary contact with asymmetric gene flow, with a log-likelihood of -539.27 and AIC of 1090.54 (Fig. 5, Table S3). Parameter estimates inferred a period of divergence in isolation (T1 = 0.07) with a shorter period of secondary contact (T2 = 0.01) and a much higher rate of migration from *R. cf. poeppigii* into *R. schneideri* (m12 = 15.5) than from *R. schneideri* into *R. cf. poeppigii* (m21 = 1.82) (Table S3).

365

366 3.4 D-statistics

367 Nearly all topological trios tested (((P1,P2)P3)O) had significant *D*-statistics (Table S4),

368 indicating that the majority of gene flow within this group is not due to incomplete lineage

369 sorting. The *R. jimi-marina-horribilis* trio was not significant (p > 0.05), indicating that we could

370 not reject the null hypothesis of no gene flow, with ABBA-BABA patterns arising solely due to

371 incomplete lineage sorting (Malinsky et al., 2020). D-statistics for all significant trios ranged

from 0.12 to 0.49 (Table S4). The highest *D*-statistics were for *R. horribilis-jimi-schneideri*

373 (0.49), *R. marina-jimi-schneideri* (0.37), and *R. schneideri-poeppigii-marina* (0.30). The *f*_b(C)

374 statistic is a summary of f_4 admixture ratios and shows excess allele sharing between the branch

375 on the y-axis and the sample on the x-axis (Malinsky et al. 2018). The $f_b(C)$ statistics indicated

the highest percentages of gene flow between *R. cf. poeppigii* and *R. marina* (11%), between *R.*

377 *cf. poeppigii* and *R. horribilis* (8%), and between *R. jimi* and *R. schneideri* (7%) (Fig. 6, Table

378

379

380 4 DISCUSSION

S5).

381 Based on comprehensive geographic and genomic sampling within a clade of South American

382 toads, this investigation found evidence of multiple distinct evolutionary lineages that span large This article is protected by copyright. All rights reserved 383 geographic areas and, at times, distinct biomes. The inferred major clades and genetic clusters

- 384 largely correspond to currently recognized taxa within *Rhinella*; however, we also found
- evidence of potentially cryptic diversity within *R. marina*, *R. schneideri*, and potentially *R*.
- 386 *poeppigii*. Genetic clustering analyses suggested that many of the inferred groups include
- 387 admixed individuals. Accordingly, demographic analyses supported that the evolutionary history
- 388 of these toads involved cross-taxon gene flow both at ancient (in the case of *R. marina, R.*
- 389 *horribilis*, and *R. jimi*) and recent (in the case of *R. schneideri* and *R. cf. poeppigii*) times. Both
- 390 demographic inference and ABBA-BABA tests inferred patterns of genetic introgression across
- 391 species, supporting previous assertions that the evolutionary history of *Rhinella* was
- 392 characterized by various levels of hybridization (Pereyra et al. 2016; Sequeira et al., 2011).
- 393

394 4.1 Phylogenetic patterns and species boundaries

395 The phylogenetic findings of this study improve our knowledge about species diversity and 396 distributions in South America. Our sampling validates previous reports of Rhinella poeppigii 397 present in western Amazonia (Venâncio et al., 2017). Rhinella poeppigii has a history of both 398 taxonomic uncertainty and misidentification, due to its similarity to *R. marina* (De la Riva, 2002; 399 Venâncio et al., 2017; Venegas & Ron, 2014). After the first individuals were identified and 400 collected in Ecuador, subsequent specimens collected in the region that were previously 401 misidentified were discovered at Museo de Zoología, Pontificia Universidad Católica del 402 Ecuador (QCAZ) (Venegas & Ron, 2014). In this study we included another individual from 403 Porto Walter, Acre, Brazil, which further corroborates *R. poeppigii* extending into Brazil. 404 Furthermore, we uncovered a group of R. cf. poeppigii specimens in eastern Amazonia near the 405 Belo Monte Hydroelectric dam on the Xingu River (Fig. 3). These samples, however, do not 406 display distinct R. poeppigii morphology, and in fact are more similar morphologically to R. 407 marina, to which they were originally assigned (Fig. S7). Unfortunately, as sampling of this 408 clade was initially unintentional, we sampled only a few individuals that could confidently be 409 called *R. poeppigii* that did not occur across the range of *R. poeppigii*, which may be 410 misrepresenting the genetic admixture visualized within this clade (Fig. 1,3). Given this

restricted sampling and the more than 2,000 km distance in sampled individuals, it may be that
eastern *R. cf. poeppigii* is actually a yet undescribed cryptic species within the *Rhinella marina*complex.

414 The mitochondrial 16S rRNA marker has been used extensively for identification and 415 barcoding of amphibians (Maya-Soriano et al., 2012; Rockney et al., 2015; Vences et al., 2005). 416 Despite this marker being extremely useful in taxonomic identification for a number of closely 417 related species (Firneno & Townsend, 2019), even within the Rhinella genus (Pereyra et al. 418 2016), there is an inherent lack of diversity recovered across all focal species within the R. 419 *marina* complex (Fig. S1-S2). It is possible that purifying selection has acted on this region of 420 the mitochondrial genome, thereby greatly reducing genetic diversity across the complex 421 (Charlesworth et al., 1995; Cvijović et al., 2018). Considering that processes like purifying 422 selection can also reduce genetic diversity at linked neutral sites, previous estimates of potential 423 introgression within Rhinella species using mitochondrial data may be similarly affected 424 (Cvijović et al., 2018). This phenomenon could have resulted in an overestimation of shared loci 425 by any other means, such as hybridization, as opposed to a constraint on particular loci. With the 426 16S fragment sequenced being relatively short (~480 bp), an analysis of the entire 16S rRNA 427 gene or even the whole-mitochondrial genome in this group could prove useful in disentangling 428 the reasons for such low genetic diversity seen here.

429 By contrast, despite evidence of admixture both within and between species, nuclear data 430 estimated a phylogeny with substantial structure and support (Fig. 1). When compared to other 431 phylogenies generated with single or multi-locus datasets, high-throughput sequencing of the 432 Rhinella marina complex has revealed a surprising amount of genetic complexity, introgression, 433 and interspecific resolution (Bessa-Silva et al., 2020; Maciel et al., 2010; Vallinoto et al., 2009). 434 These patterns suggest that in groups with such complex demographic histories, and especially 435 those with a likelihood of hybridization between divergent populations or species, large-scale 436 genetic data can be very useful in disentangling relationships and histories. These types of data 437 could be utilized in future studies to identify potentially adaptive regions of the genome that 438 correlate with phenotypic or ecological differences between populations.

439

440 **4.2 Biogeographic drivers of species range limits**

441 Inferred species range limits can be attributed to both present-day spatial environmental 442 gradients and the history of topographic change in South America, as suggested for a number of 443 other South American taxa (Carnaval et al., 2009; Fonseca et al., 2018; Gehara et al., 2014; 444 Prates, Rivera et al., 2016). Mitochondrial divergence time analyses are consistent with the idea 445 that the Andean uplift contributed to divergence between *R. marina* and *R. horribilis* (Fig. 4); pronounced genetic divergence between populations on each side of the Andean chain supports 446 447 the recent recognition of *R. horribilis* as a taxon distinct from *R. marina* (Vallinoto et al., 2009). 448 While the Andes likely limits contemporary gene flow between these two taxa, our finding of 449 admixture between them suggests that the northern Andes may be a semi-permeable barrier (Fig. 450 2), in agreement with patterns seen in other organisms (Acevedo et al., 2016; Bessa-Silva et al., 451 2020; Maciel et al., 2010). Additionally, like other amphibians (Noonan & Wray, 2006) and 452 reptiles (Gamble et al., 2008), the extensive fluvial network formed in western Amazonia by 453 periodic Miocene flooding, known as the Pebas formation, may have contributed to divergence 454 not only between R. horribilis and R. marina, but also between the northeast and south-455 southwestern Amazonian clades within R. marina (Vallinoto et al., 2009; Wesselingh & Salo, 456 2006). *Rhinella marina*, which is comprised of two well-supported clades, is distributed across 457 Amazonian climates, which are known to have asynchronous historical eastern-western climatic 458 cycles and have had an effect on species composition and genetic diversity within the biome 459 (Cheng et al., 2013; Prates, Rivera et al., 2016). Considering that the distinct clades have a 460 northern-southern distribution, as opposed to an eastern-western distribution, however, it may be 461 more plausible that geographic barriers, such as fluctuating fluvial networks from the Miocene 462 through the Pleistocene (Cooke et al., 2012; Lundberg et al., 1998), have had a higher impact in 463 promoting divergence between these clades within *R. marina*.

Similar to what is observed within *R. marina*, we see patterns of species distributed
across environmental gradients repeated across the phylogeny; *R. schneideri* is distributed across
the Cerrado, through Cerrado-Caatinga-Atlantic Forest ecotones, and into the northern Atlantic

467 Forest, with an east-west gradient of admixture (Fig. 1,3). The Seasonally Dry Tropical Forests 468 and savannas of South America have been known to harbor complex and cryptic genetic 469 diversity and have been especially affected by Quaternary climate fluctuations (Bandeira et al., 470 2021; Fonseca et al., 2018; Gehara et al., 2017; Prado et al., 2012; Vasconcellos et al., 2019; 471 Werneck et al., 2015). Considering the phylogenetic pattern that we see within *R. schneideri*, we 472 can posit that this species expanded eastward during Plio-Pleistocene climate change (Bandeira 473 et al., 2021; Lisiecki & Raymo, 2007). Paleoclimatic modeling of the biogeographic history and 474 niche of R. schneideri on a finer scale is recommended to validate this hypothesis.

475 A puzzling biogeographic pattern that emerged from our results is the extremely disjunct 476 distribution between R. poeppigii in western Amazonia and its sister clade, R. cf. poeppigii, from 477 eastern Brazilian Amazonia, more than 2,000 km apart. This mysterious pattern has also been 478 reported for other herpetofaunal species, including the lizards Anolis trachyderma (Ribeiro-479 Júnior, 2015) and Potamites ecpleopus (Ribeiro-Júnior & Amaral, 2017) and the horned treefrog 480 Hemiphractus scutatus (de Lima Moraes & Pavan, 2018). Despite this large geographic distance, 481 as well as the effects of contrasting climatic seasonality between the eastern and western 482 localities in this region on other herpetofauna (Cheng et al., 2013; Prates, Rivera et al., 2016; 483 Wang et al., 2017), this and other studies indicate limited genetic divergence across disjunct 484 regions (de Lima Moraes & Pavan, 2018). A comprehensive analysis of museum specimens and 485 available tissues from these areas, in conjunction with a more thorough sampling of *R. poeppigii* 486 across its range, will be required to confirm this unexpected pattern of genetic divergence within 487 this group.

488

489 **4.3 Hybridization and introgression**

490 The interspecific relationships inferred with historical demographic modeling suggest extremely

491 varied patterns of migration and hybridization through time within the *Rhinella marina* complex.

492 Our study indicates that species within this group have diverged across multiple biomes and

493 amassed significant genetic differentiation despite continuous gene flow among species (Fig. 5).

494 Many of the species within the *Rhinella marina* complex also have a shared introgressive history

495 (Fig. 1,6; Table S4-S5). Hypothesis testing of demographic models suggests that the R. marina-496 horribilis-jimi clade continued to exchange genes throughout its dispersion across the continent, 497 and species within this clade exchanged genes with other species within the complex (Fig. 1,5-6, 498 Table S5). *Rhinella* species have long been observed to have overlapping ranges with the 499 potential for hybridization, especially given the propensity for these toads to participate in 500 "mating balls", wherein different toad species - and sometimes from different genera -501 seemingly mate with one another (Fontenot et al., 2011; Maciel et al., 2010; Pereyra et al. 2016; 502 Sequeira et al., 2011; Thomé et al., 2012). There is a relative lack of data, however, surrounding 503 behavioral aspects of intraspecific interactions among species outside of these mating balls. 504 Despite evidence of gene flow between species, there was no evidence of population-wide 505 hybridization or the presence of hybrid species within our sampling. Potential proposed 506 hybridization events have been reported within or between Rhinella species groups, such as 507 within the R. granulosa complex (Guerra et al., 2011; Pereyra et al. 2016) and the R. crucifer 508 complex (Júnior et al., 2004; Thomé et al., 2012), where either instances of morphologically 509 intermediate individuals or hybrid populations have been reported. Much of the speculation 510 surrounding hybridization in Neotropical toads has been accompanied by a lack of data from 511 natural populations to assess the biological reality of presumed hybrid species (Fontenot et al., 512 2011; Malone & Fontenot, 2008; Thomé et al., 2012). Within the R. marina group, however, we 513 found that recurrent gene flow between species at low levels is much more prevalent than the 514 persistence of repeated gene flow in massive multi-species mating events.

515

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526 DATA AVAILABILITY STATEMENT

- 527 16S sequences and demultiplexed genetic data are available through NCBI (BioProject
- 528 PRJNA772164; Table S1) and associated scripts and input files are available through dryad:
- 529 https://doi.org/10.5061/dryad.7pvmcvdtp.

530 AUTHOR CONTRIBUTIONS

- 531 DR and MKF conceived and designed the study. DR, MKF, MTR and IP acquired funding.
- 532 MTR, IP, and JPC provided samples. DR, IP, and TJF wrote software used for analysis. DR and
- 533 TJF analyzed the data. DR, IP, and MKF wrote the manuscript with input from all authors.

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799 TABLES AND FIGURES

- 800 FIGURE 1 (A) Maximum likelihood phylogeny of *Rhinella marina* complex focal species using
- 801 ddRADseq data and corresponding ADMIXTURE plot (K=7). Black circles on the phylogeny
- denote ML bootstrap support (BS) > 95 and Bayesian posterior probability (PP) > 0.95. (B)
- 803 DAPC plot (K=7).

804 **FIGURE 2** Locality map for the subset depicting average ADMIXTURE cluster assignments per

805 locality (K=7) for *Rhinella horribilis*, *R. marina*, and *R. jimi*. Colors correspond to Figure 1. Map

806 partitioned into biomes (Central America, Northern Andes, Northern Amazonia, Western

807 Amazonia, Eastern Amazonia, Southern Amazonia, Pantanal, Chaco, Cerrado, Caatinga,

808 Northern Atlantic Forest, Southern Atlantic Forest).

809 FIGURE 3 Locality map for the subset depicting average ADMIXTURE cluster assignments per

810 locality (K=7) for *Rhinella poeppigii* and *R. schneideri*. Colors correspond to Figure 1. Map

811 partitioned into biomes (Central America, Northern Andes, Northern Amazonia, Western

812 Amazonia, Eastern Amazonia, Southern Amazonia, Pantanal, Chaco, Cerrado, Caatinga,

813 Northern Atlantic Forest, Southern Atlantic Forest). R. poeppigii range adapted from (De la

814 Riva, 2002; Venâncio et al., 2017; Venegas & Ron, 2014).

815 FIGURE 4 Time calibrated phylogeny based on mitochondrial 16S data. Black circles indicate

816 PP > 0.95, * indicates the fossil calibration, and the bars represent the 95% HPD, which are also 817 in parentheses. Colors correspond to the phylogeny in Fig. 1.

818 FIGURE 5 Optimal demographic models and residual plots for the (A) 3D-JSFS analysis of

819 *Rhinella horribilis, R. marina*, and *R. jimi*, and (B) 2D-JSFS analysis of *R. poeppigii* and *R.*

820 schneideri.

FIGURE 6 The f_b statistic (summary of f_4 admixture ratios). Grey color corresponds to tests that are not possible because of constraints on the phylogeny. * indicates a significant result.

TABLE 1 Optimal demographic models and estimated parameters. Abbr: $LL = log-likelihood; \theta$

824 (4N_{ref}µL) = the effective mutation rate of the reference population (ancestral population); nul,

825 nu2 = effective population sizes under the constant population size model; nuA = effective

population sizes of the ancestral population; mA, m1, m2 = migration rates between the ancestral

(A), first (1) or second (2) population; m12 = migration rate from population two to population

828 one; m21 = migration rate from population one to population two; T1, T2, T3 = unscaled time

829 between demographic events.

Model

2D	LL	AIC	<u>0</u>	<u>nu1</u>	<u>nu2</u>		<u>m12</u>	<u>m21</u>	
divergence in isolation with									
continuous asymmetric	-539.27	1090.5	1424.5	0.01	0.05		15.5	1.82	
secondary contact									
3D	<u>LL</u>	<u>AIC</u>	<u>0</u>	<u>nu1</u>	<u>nu2</u>	<u>nu3</u>	<u>nuA</u>	<u>mA</u>	<u>m1</u>
ancient migration ^{Thistfr tiple itesto} isolation	tected by co -1572.7	opyright. A 3165.4	ll rights res 95.33	erved 0.57	0.15	1.31	6.15	0.05	1.36



R. horribilis
R. jimi
R. marina (a)
R. marina (b)



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1200

N Amazonia

W Amazonia

Central America mec_16280_f2.pdf

Andes

SAmazonia

Cerrado

🖤 Caatinga

E Amazonia

Known R. poeppigii range R. c.f. poeppigii R. schneideri (a) R. schneideri (b) W Amazonia



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Caatinga

Cerrado

Atlantic Forest mec_16280_f4.pdf



mec_16280_f5.pdf



