Temporal refuges of a subordinate carnivore vary across rural-urban gradient Rumaan Malhotra^{1*}, Samantha Lima² and Nyeema C. Harris³

¹Ecology and Evolutionary Biology, University of Michigan 1101 N. University Ave, Ann Arbor, Michigan 48106
²Forestry and Natural Resources, Purdue University 715 W State St, West Lafayette, Indiana 47907
³Applied Wildlife Ecology Lab, School of the Environment, Yale University 195 Prospect St.

³Applied Wildlife Ecology Lab, School of the Environment, Yale University 195 Prospect St. New Haven, Connecticut 06511

*Correspondent: rumaanm@umich.edu

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the Version of Record. Please cite this article as doi: 10.1002/ece3.9310

This article is protected by copyright. All rights reserved.

Author Manuscript

Background: Animals exhibit variation in their space and time use across an urban-rural gradient. As the top-down influences of apex predators wane due to human-driven declines, landscape level anthropogenic pressures are rising. Human impacts can be analogous to apex predators in that humans can drive increased mortality in both prey species and carnivores, and impact communities through indirect fear effects and food subsidies. Here, we evaluate the time use of a common mesocarnivore across an urban rural gradient, and test whether it is influenced by the intensity of use of a larger carnivore. Methods: Using multiple camera-trap surveys, we compared the temporal response of a small carnivore, the raccoon (*Procyon lotor*), to the larger coyote (*Canis latrans*) at four study areas across Michigan that represented a gradient of pressure from humans. Results: We found that raccoon time use varied by study area and was most unique at the rural extreme. Raccoons consistently did not shift their activity pattern in response to covotes at the study area with the highest anthropogenic pressures despite considerable interannual variation, and instead showed stronger responses to coyotes at more rural study areas. Temporal shifts were characterized by raccoons being more diurnal in areas of high coyote activity. Conclusions: We conclude that raccoons may shift time use in the presence of coyotes, dependent on the level of anthropogenic pressure. Our results highlight that the variation in raccoon time use across the entirety of the urban-rural gradient needed to be considered, as anthropogenic pressures may dominate and obscure the dynamics of this interaction.

Keywords: coyote, niche, partitioning, landscape of fear, Michigan, raccoon

1 | INTRODUCTION

Urban-rural gradients provide comparisons between natural (e.g. top down predation) and anthropogenic forces (e.g. fear of humans) that structure wildlife communities through behavioral and ecological pathways (McDonnell and Pickett 1990; Ellington and Gehrt 2019). Non-consumptive effects that apex predators exert on prey or smaller competitors commonly manifest as antipredator behaviors (Wirsing et al. 2021). Similar to apex predators, humans can induce non-consumptive effects on subordinate species through changes in space and time use (Ciuti et al. 2012; Clinchy et al. 2016). However, humans are unique in their top-down pressures in that they can exert fear effects across trophic levels, superseding hierarchies in natural systems (Smith et al. 2017; Suraci et al. 2019). Thus far, urban-rural gradients have predominantly highlighted changes in diet and physical characteristics (e.g., body size) that can affect ecological interactions, or changes in biodiversity and species composition across taxa (Marzluff 2001; Urban et al. 2006; Gámez et al. 2022). Although not specifically cast in an urban-rural framework, there is further evidence that humans and built structures can alter animal behavior (Van Donselaar et al. 2018; Avilés-Rodríguez and Kolbe 2019). For example, global metaanalyses found that intensity of human pressure can drive increased nocturnality and reduce movement (Gaynor et al. 2018; Tucker et al. 2018). Altered time use due to humans can further translate into altered interspecific interactions, for example by increasing the total spatiotemporal overlap and thus the probability of encounter (Lewis et al. 2015; Gallo et al. 2019). In urban areas, where spatial overlap among species is inevitable due to the limited amount of habitat available, temporal partitioning may be particularly important for species' persistence (Adams and Thibault 2006; Santos et al. 2019; Stark et al. 2020). We leverage an urban-rural gradient

formed by human pressure to examine spatiotemporal dynamics between a widely distributed carnivore and a smaller sympatric competitor.

As a highly adaptive mesocarnivore, covotes (*Canis latrans*) exploit a wide range of habitats and exhibit tolerance to disturbance (Bekoff and Gese 2003; Flores-Morales et al. 2019). Coyotes exemplify mesopredator release, a phenomenon in which subordinate carnivores increase in abundance and distribution once the suppressive effects of larger carnivores are removed (Crooks and Soulé 1999; Prugh et al. 2009). For example, the recent range expansion of coyotes aligns with human-caused extirpation of gray wolves (Canis lupus). Though coyotes are subordinate to larger carnivores where they are sympatric, they are aggressors towards several smaller carnivores and account for high rates of mortality for some species (e.g., Vulpes velox, Vulpes macrotis) (Bekoff and Gese 2003; Berger 2007). As a result, coyotes are commonly cited as a species that can act as both a mesopredator or an apex predator in their community, depending on the presence of larger carnivores such as the gray wolf (Prugh et al. 2009; Roemer et al. 2009; Colborn et al. 2020) or mountain lion (Puma concolor) (Elbroch and Kusler 2018; Ruprecht et al. 2021). Similarly, raccoons (*Procyon lotor*) exhibit tolerance to human pressures and spatially overlap with coyotes through much of their North American range (Timm et al. 2017; Kays 2018). Covote-raccoon interactions are interesting because both species are widespread, and exhibit a size difference that should typify intraguild aggression or predation (Donadio and Buskirk 2006). Despite this, we lack evidence for any sort of spatial or temporal partitioning between coyotes and raccoons (Gehrt and Clark 2003; Shedden et al. 2020). There has yet to be a study that examines the temporal dynamics of these two species across the urbanrural gradient.

Raccoons exhibit spatiotemporal variation in behavioral attributes, leading us to the expectation that the response of raccoons to covotes may vary by habitat and other characteristics across study areas (Beasley et al. 2011). Based on a lack of avoidance behavior by raccoons or raccoon mortality due to coyotes, Gehrt and Prange (2007) argued that raccoons and coyotes do not fit into the mesopredator release hypothesis. There is little evidence that coyotes act as a control on the abundance or spatial use of raccoons (Lesmeister et al. 2015). Telemetry studies of raccoons have found some evidence of mortality due to covotes, but only as a rare occurrence (Gehrt and Clark 2003; Prange et al. 2003). In North Carolina, raccoons exhibited low levels of vigilance despite their temporal overlap with coyotes (Chitwood et al. 2020). Given that covotes pose some risk to raccoons based on size and sympatry, but that overall risk is low, we tested whether raccoons shifted their activity based on heterogeneity in coyote risk within a study area. Employing a camera survey across an urban-rural gradient, we tested whether raccoon time use differed between intensities of coyote spatial use. Specifically, we tested the variation in raccoon and coyote time use across two scales: between study areas (across the urban-rural gradient) and within study area, between years (interannual variation) (Figure 1). As anthropogenic pressures increase, our knowledge of contemporary baseline ecological interactions becomes dated. Thus, it becomes essential to understand how these competitive interactions compare across landscapes with varying human pressures. Therefore, we tested three hypotheses based on wildlife prioritizing the avoidance of human activity in their temporal activity:

1) Raccoon time use at the most urban study area will significantly differ from the other three study areas.

2) Interannual variation in both raccoon and coyote activity will be lowest at the urban study area.

3) Raccoon time use will shift in areas of high coyote activity at the rural end of the urban-rural gradient, but not at the urban end.

2 | METHODS

2.1 | Study area

We investigated raccoon temporal dynamics across differing levels of coyote activity at four study areas across the state of Michigan, USA (Figure 2) which represent an urban-rural gradient.

1) The Huron Mountain Club (HMC) is a privately-owned property along the southern shore of Lake Superior, encompassing around 6,900 hectares in Marquette County, Michigan, USA. This study area has a wide variety of habitats including beech-sugar maple hardwood forests, aspen dominated stands, and coniferous boreal forests. Sympatric large predators include: gray wolves, black bears (*Ursus americanus*), and coyotes. Anthropogenic pressures are limited to a small, seasonally occupied area of human habitation near the north central part of the property. Hunting and fishing occur on the property, and the intensity is presumably low due to restrictive public access.

2) The University of Michigan Biological Station (UMBS), a ~4,000 hectare research station and forest in Pellston County, Michigan, USA served as one of our intermediate disturbance study areas. With repeated logging and fire disturbance until 1923, the secondary forest is a mix of transitional hardwood and boreal forests. Douglas and Burt lakes along the north and south, and the town of Pellston and a major highway along west and east, respectively border this study area. Large co-occurring predators include: black bears, coyotes, and coyote-

wolf hybrids. We were able to distinguish the few known coyote-wolf hybrids in the area due to them having collars from a different study, which were visible in the camera trap images (Wheeldon et al. 2012). Human pressures resulted from regulated research infrastructures for climate monitoring and housing facilities with low levels concentrated seasonally during the summer.

3) The Shiawassee National Wildlife Refuge (SNWR) is a 9,870 hectare wildlife refuge managed by the US Fish and Wildlife Service. The refuge is comprised of forested hardwood wetlands and prairie. The city of Saginaw abuts the northern edge of the refuge and is surrounded by agricultural land for crop farming. The only large native predator present is the coyote. Anthropogenic pressures, in addition to the urban and ex-urban nature of the boundaries, are in the form of recreational visitors and hunting. Public hunting for deer and waterfowl, and furbearer trapping are permissible on the refuge in accordance with lawful seasons.

4) The Detroit Metro Parks (DMP) is a noncontiguous collection of greenspaces interspersed throughout southeast Michigan that is managed by the Detroit Parks and Recreation Department. We chose twenty-five of these parks that varied in size from ~1.6 - 480 hectares, tree cover, human visitation, and degree of disturbance. Roads, buildings, or a riverine edge bound all parks. The only large native predator present is the coyote. Strong anthropogenic pressures are present in the form of the surrounding urban matrix as well as the associated presence of humans and domestic pets across parks..

2.2 | Camera trap survey

We deployed remotely-triggered camera traps (Reconyx© PC 850, 850C, 900, 900C) throughout each study area with camera placement and sampling design proportional to study area size (Table 1). Our study uses data from three surveys at DMP (2017, 2018, 2020), three

surveys at SNWR (2016, 2017, 2018), two surveys at UMBS (2015, 2016), and four surveys at HMC (2016, 2017, 2018, and 2019). Unbaited camera traps were affixed to trees > 0.5m diameter and placed 0.5-1.0 m off the ground. Study area-specific placement of camera traps was determined by signs of animal activity such as game trails and scat. Camera trap settings included: high sensitivity, one-second lapse between three pictures in a trigger, and a 15-second quiet period between triggers.

Image identifications were initially crowd-sourced and filtered for carnivores using a public-science program called *Michigan ZoomIN* in combination with a consensus algorithm and expert validation (Gadsden et al. 2021). Carnivore species identifications were also sorted and confirmed by at least two independent researchers in the Applied Wildlife Ecology Lab.

2.3 | Temporal activity

Time stamps associated with the camera trap images were used to conduct temporal analyses. Prior to all analyses, a 30-minute delay between triggers was introduced for every species to account for pseudoreplication, given the tendency of some animals to remain in front of the camera trap and trigger it multiple times. Since surveys were conducted across different times of the year, we scaled times to sunrise and sunset times using the *sunTimes* function in the 'circular' package in R (Agostinelli and Lund 2017).

2.31 | Variation between study areas

For each study species, we first compiled all triggers from each survey within a study area to have an aggregate across years of overall temporal activity at each study area. We then compared raccoon and coyote temporal activity between study areas (Figure 1a) using the Mardia-Watson-Wheeler (MWW) test, which is a nonparametric test of differences in the angular means between samples of circular data using the 'circular' package in R (version 4.1.0). When the W value is high it results in a significant p value (p < 0.05), which we conclude to mean that the compared temporal activities are different.

2.3.2 | Seasonal and yearly variation

Our multi-site camera study allowed us to compare differences in temporal activity of our study species based on landscape level differences along an urban-rural gradient. Comparing between seasons can confound inferences from the analyses, due to different seasons potentially resulting in different detection rates (Marcus Rowcliffe et al. 2011). While we did not have identical seasonal coverage for every study area, the multiple surveys at every study area resulted in coverage for the entire year at every study area with the exception of UMBS (Figure S1). To determine if there was consistency at study areas regardless of season and year, we compared temporal activity of each of our study species between each survey within each study area, and then looked for broader patterns across study areas (Figure 1b).

2.3.3 | Coyotes on raccoon temporal activity

For each survey, we used a kernel density estimation for the independent coyote triggers and designated the cameras that fell within the top quartile as 'HIGH' coyote intensity of use zones in ArcGIS Pro (version 2.3.1). We used this rather than a fixed cutoff value of expected detection rate because our study areas spanned the entirety of the urban-rural gradient and expected detection rates for coyote vary depending on the composition of a study area (Magle et al. 2014). Coyote triggers were checked for spatial independence using Moran's I prior to kernel density estimation. We compared raccoon temporal activity between the high coyote cameras and the rest of the study area using the MWW test (Figure 1b,c). For additional evidence that temporal shifts by raccoons were due to avoidance of coyotes, we then compared the overlap between

coyote and raccoon time use in the two raccoon test groups from the MWW test. To do this, we calculated an overlap (Δ) coefficient of temporal activity for coyotes and raccoons within each group ('HIGH' and 'LOW' coyote intensity of use) along with 95% confidence intervals generated from 10,000 parametric bootstraps of the temporal distribution models. Δ values range from 0 to 1, with 0 indicating completely distinct and non-overlapping temporal activity between comparison groups, and 1 indicating complete overlap. Δ_1 was used for comparisons when one of the sample groups had less than 50 triggers; otherwise Δ_4 was used to estimate temporal overlap (Ridout and Linkie 2009). Finally, the activity distributions were visually assessed to determine qualitative characteristics of shifts (e.g. raccoons shifting towards increased nocturnality in high coyote zones).

3 | RESULTS

We obtained 1,378 coyote and 11,136 raccoon triggers with a 30-minute quiet period across 12 surveys in 82,595 trap nights (HMC- 36,868; UMBS- 12,953; SNWR- 12,477; DMP- 20,297) from 2015-2020. Raccoons and coyotes were the most common carnivores in almost every survey, comprising 57-98% of all the carnivore triggers. In Detroit, where domestic dogs (*Canis familiaris*) and cats (*Felis catus*) comprised 35% of the triggers, coyotes were the fourth most common carnivore species after raccoons, cats, and dogs.

3.1 | Coyote intensity of use

We found no evidence of spatial autocorrelation in coyote detections based on non-significance in Moran's I results across all surveys. Kernel density estimates indicated coyotes were distributed non-randomly in space (Figure 2). At DMP with heavy anthropogenic pressure (average 77 coyote triggers per camera in "HIGH" coyote zones), coyote spatial activity was concentrated in two heavily forested parks and had few human triggers compared to the rest of the surveyed parks in Detroit. In contrast, at HMC which had the least amount of anthropogenic pressure and the presence of wolves, the highest coyote spatial activity occurred in a recreation area that contained several buildings and homes but had few overall triggers (average 3 coyote triggers per camera in "HIGH" coyote zones). Coyote spatial activity formed distinct zones in SNWR and UMBS as well, and the location of hotspots varied by survey. Hotspots at these two study areas were not associated with any discernible landscape level measures of anthropogenic pressures. Raccoon triggers were recorded within both the low and high zones of coyote spatial activity across all study areas.

3.2 | Variation in temporal activity between study areas

Raccoon activity at each study area was unique, showing significant differences in every pairwise comparison of study areas from MWW tests (Table 1a). We expected raccoon activity to be the most distinct at the most urban study area, DMP. Instead, we found that raccoon activity was most unique at HMC (the opposite end of the gradient), showing considerably more use of the diurnal period (Figure 3a). Overlap comparisons between the sites reflected this. Overlap in raccoon temporal activity between HMC and the other three study areas was relatively low, with confidence intervals for these comparisons showing 76-87% overlap. In contrast, comparisons between UMBS, SNWR, and DMP were significantly higher, with confidence intervals for comparisons between these sites showing 87-96% overlap .

Coyote activity showed a markedly different pattern than raccoon activity did across study areas (Table 1c, Figure 3b). Coyote time use at DMP (the most urban study area) and at HMC (the most rural study area) showed significantly higher overlap than any other area comparisons, which was marked by increased diurnal activity. Coyote time use at the intermediate study areas (SNWR and UMBS) also showed high overlap. Overall, coyote time use fell into two distinct groups: one reflecting the extremes of the gradient, and another reflecting the intermediate.

3.3 | Seasonal/annual variation in temporal activity

Raccoon activity varied significantly by survey and year for every survey on the more urban side of the gradient (SNWR and DMP), but not on the rural end of the spectrum (UMBS and HMC) (Table 2A). At the most urban end of the urban-rural gradient, raccoon activity was significantly different between every year surveyed at DMP and SNWR. At UMBS, the comparison between the two years approached significance (W = 5.53, p = 0.063). For HMC the results varied, depending on the years compared. For example, 2016/2017 and 2017/2018 comparisons showed that raccoon time use varied significantly between these years, while raccoon time use between 2018 and 2019 was similar (W = 3.03, p = 0.220). These results refuted our hypothesis that interannual variation would be weakest for raccoons at DMP, instead showing that interannual variation is stronger at the urban end of the gradient.

Coyote activity was generally more consistent across surveys and years than raccoon temporal activity (Table 2b). Similar to the results for raccoon activity, there was interannual variation in coyote activity on the urban end of the gradient. However, in contrast with the raccoon activity results, this was restricted to only the most urban study area (DMP). The lone exception was the comparison between the HMC 2019 and HMC 2016 surveys, which also showed a significant difference in coyote time use (W = 11.043, p = 0.004). This result

highlighted a broader trend for coyote temporal activity: there was no significant difference in the temporal activity of coyotes between surveys in a study area unless the surveys were more than a year apart.

3.4 | Coyote use on raccoon temporal activity

Overall, our hypothesis for raccoon-coyote temporal interactions was largely correct, with raccoons at DMP (the most urban study area) consistently exhibiting no shift in time use relative to coyote intensity of use zones. However, there was reduced overlap between coyotes and raccoons within the high coyote zone. Results for the other areas varied by survey year (Table 3). Below, we first present for each study area the results for the comparison of raccoon activity between the high and low coyote zone. Then we provide the comparison of raccoon and coyote temporal activity within the high coyote zone (relative to the same comparison in the low coyote zone), to determine if there is evidence that a shift in raccoon activity between zones is due to temporal avoidance of coyotes (Figure 4).

HMC: At the most rural study area, we found results for the effects of coyotes varied by survey. The 2016 and 2017 surveys exhibited no shifts, while surveys in 2018 and 2019 showed significant shifts in raccoon activity between coyote low and high zones (W = 15.12, 10.02, p < 0.001 respectively) (Table 3). Results were consistent even when the 2017 survey was separated into summer and winter survey seasons since it covered an entire year, indicating no shifts in raccoon activity between coyote zones. When comparing coyote and raccoon temporal activity within each zone the 2018 survey showed some evidence of decreased temporal overlap between coyote and raccoons in the high coyote zone, while for 2019 the confidence intervals were too

Author Manuscript

wide to be meaningful (Figure 5).

UMBS: For both surveys, we found there were significant shifts in raccoon activity between coyote zones (W = 9.63, p < 0.001 for 2016, and W = 7.39, p = 0.025 for 2015). Both surveys also showed evidence of reduced temporal overlap between coyote and raccoons in the high coyote zone.

SNWR: We found that again, results varied by survey, with two out of three surveys showing significant shifts in raccoon activity between coyote zones; 2016 (W = 6.08, p = 0.047) and 2018 (W = 10.46, p < 0.001) showed shifts, while in 2017 (W = 3.65, p = 0.162) raccoons did not shift activity. Only the 2018 survey showed evidence of reduced temporal overlap between coyotes and raccoons in the high coyote zone.

DMP: We found that raccoons exhibited no shifts in activity between coyote zones consistently across for all three years surveyed in our study. Curiously, two out of the three surveys (2018, 2020) showed evidence of reduced overlap between raccoons and coyotes in the high coyote zone, with the difference reaching significance in the 2020 survey (Δ 4 CI in the high coyote zone: 0.46-0.58 vs. low coyote zone: 0.61-0.80).

4 | DISCUSSION

Behavioral adjustments in diet, spatial, and temporal use can reduce competition for resources to promote coexistence (Inouye 1978). We tested for spatial and interannual variation in the time use of raccoons across an urban-rural gradient and measured the use of temporal refuges by raccoons in the presence of coyotes across that same gradient. As expected, we found that raccoon time use varied both across the gradient and among years. More importantly, we highlight that there were consistent patterns across the urban-rural gradient in raccoon temporal

response to coyotes. We found that at the most urban study area (DMP), raccoons consistently did not shift their temporal activity in response to coyotes, despite significant interannual variation in raccoon activity. In contrast, all other study areas showed some evidence of behavioral plasticity in raccoon time use with the intensity of coyote spatial use. Our results lend some support to findings that non-consumptive or fear effects are present within the hierarchy of the carnivore guild (Gordon et al. 2015), but are overall better explained by human-associated factors.

4.1 | Variation in temporal activity across study areas

Urban systems represent an extreme of human pressures, and the continuing increase in urban habitat makes understanding the unique behaviors and ecologies of wildlife in urban spaces such as Detroit, Michigan particularly important. For example, Breck et al. (2019) found that coyotes at urban study areas are bolder in comparison to their rural counterparts. Thus, urban coyotes may be less constrained by the fear of humans in their space and time use. This may explain our surprising result that coyotes were less nocturnal in Detroit compared to the intermediate study areas, the opposite of what we would expect for avoidance of humans (Gaynor et al. 2018). This temporal activity pattern better fits the similar result from HMC, where the activity pattern we found would be consistent with the avoidance of wolves. Fowler and colleagues (2021) found little evidence of spatial partitioning between coyotes and wolves in the Upper Peninsula of Michigan (where HMC is located), so temporal partitioning is a plausible coexistence mechanism. The increased nocturnality at the intermediate study areas may be due to the lack of a larger natural predator (the wolf) in combination with higher hunting pressure due to lower tolerance of rural hunters for coyotes (Drake et al. 2019).

Compared to coyotes, the temporal activity of raccoons seems to consistently become more nocturnal with increasing human pressure. Raccoons have been shown to display a fear response to dog vocalizations in playback experiments, and increasing nocturnality may reflect avoidance of domestic dogs (which are largely diurnal) at the urban end of the gradient (Suraci et al. 2016). Surprisingly, it was not the human-dominated urban system that was the most unique in raccoon temporal use amongst the study areas, but instead the more pristine HMC in northern Michigan. The overall raccoon activity pattern showed considerable use of the diurnal period, resulting in low overlap with other study areas. We could similarly attribute this result to the low amount of human and domestic dog presence in the study area.

4.2 | Interannual variation in temporal activity

HMC was the study area with the greatest interannual variation in raccoon response to coyotes out of the four study areas. One explanation is the a lack of human impact in the form of food subsidies, as raccoons rely heavily on anthropogenic foods in some systems (Demeny et al. 2019; Manlick and Pauli 2020). The availability of resources can modulate the strength of competition, and so annual variation in food resources could drive the avoidance response of raccoons to coyotes (Newsome et al. 2015). At the other three study areas, human food waste and other human-derived subsidies likely offset years that may otherwise be relatively resource-poor for raccoons (Oro et al. 2013). Unlike UMBS and SNWR, which have nearby towns, HMC is isolated, surrounded by forest and with the few cabins on the property only seasonally occupied. However, overall raccoon activity (without the consideration of coyote spatial use) showed the greatest interannual variation at the other end of the urban-rural gradient at the two more urban study areas, which would contradict the explanation of anthropogenic food subsidies unless these resources exhibited annual or seasonal differences in the availability. A more plausible

explanation is that the level of dependence of raccoons on anthropogenic food subsidies varies based on the season, driving differences in activity patterns. A seasonal dependence on food subsidies would further account for variation between years surveyed in the same fall-winter season (such as DMP '17 and '18), since onset of cold weather and snow varies annually.

While there was also some interannual variation in coyote temporal activity at DMP, the general consistency in coyote time use from year to year could indicate that coyotes are either less plastic in the temporal niche, are tracking resources and threats spatially, or that there is little variation in resources and threats over seasons and years.

4.3 | Coyote spatial use on raccoon temporal activity

As the largest wild carnivore in Detroit, the coyote has the potential to act as a fear source for other wild carnivores. In absence of shifts in raccoon activity at our DMP study area, it seems that the fear effect does not extend to raccoons, consistent with the findings of Chitwood and colleagues (2020). Given that we did find some evidence of temporal avoidance in our other study areas, it is possible that fear of coyotes is not strong enough to elicit a shift in raccoon time use in the face of a stronger force; the most obvious in an urban system being humans and domestic dogs, as reflected by raccoons at DMP having the least diurnal activity (Figure 3a) (Gaynor et al. 2018; Nix et al. 2018; Sévêque et al. 2022). Despite raccoon activity consistently being similar between zones of coyote intensity of use, raccoon activity did seem to show somewhat reduced overlap with coyote activity in the high coyote intensity of use areas. This implies that coyotes were potentially using time differently depending on how heavily used an area was by conspecifics. A plausible explanation would be intraspecific competition or aggression (e.g. Newsome et al. 2019, dingoes in Australia), or it could more generally suggest coyotes are more plastic in their time use than raccoons in urban systems (McClennen et al.

2001). The latter is supported by the higher sensitivity of coyotes to human activity; although both species are cosmopolitan, raccoons are more human tolerant than coyotes (Crooks 2002; Randa and Yunger 2006; Green et al. 2022).

Our results highlight broad patterns in raccoon temporal use between zones of high and low coyote activity. The mechanisms that underlie these patterns require further study and a temporal shift could very likely have more nuance than simple avoidance by a subordinate carnivore. A shift in temporal use by a subordinate (as shown in our SNWR and DMP study areas) might instead reflect the pursuit of an alternate resource (e.g. avoiding exploitative competition by pursuing different prey) rather than avoidance of the antagonistic interaction itself (Newsome et al. 2015). While our results indicate the response of the raccoon to be driven by a larger predator, it does not preclude an interaction between top-down and bottom-up forces, which may be important to understanding what raccoons are directly responding to across study areas and survey seasons (Elmhagen and Rushton 2007). For example, resource availability, such as the abundance of small mammal prey, fluctuates with season and could be a driver of varying levels of competition between coyotes and raccoons (Batzli 1992; Fedriani et al. 2000; Sovie et al. 2019). Seasonal variation in temporal response may explain the divergent result for the 2017 SNWR survey, which occurred during the summer months. The other two surveys at the study area occurred during the fall and the spring, periods which are associated with heightened resource gathering for the imminent winter, and heightened coyote aggression because of the coyote breeding season (Way et al. 2001). Pairing dietary studies that explore the seasonal variation in coyote and raccoon diets across all study areas with spatiotemporal analyses would elucidate if seasonal variation in resource availability drives resource partitioning between these species.

Though the two study areas at the opposite ends of the gradient (i.e., HMC and DMP) best highlight the variation in raccoon temporal activity and temporal response to coyotes, there were study area specific patterns for the entire gradient. We intended for our sampled study areas to represent opposing gradients of humans and native apex predator presence, which were reflected in the amount of built structures and which carnivores were captured on camera at each study area. Since we did not explicitly test for the effect of the relative activity of apex predators and humans, we cannot discount the possibility that factors other than top-down forces drove the urban-rural gradient we observed in our results. Study areas varied in vegetative cover, topography, latitude, and distribution of resources. However, differences in the sources of top-down forces are the most obvious and likely ecological factor that differs between the study areas for generalist species such as raccoons and coyotes. Similar outcomes have been reported for other coyote-subordinate predator systems when compared across study areas that vary in the presence of an apex predator (Shores et al. 2019).

4.1 | Study limitations

The limitations of our study primarily centered around interannual variation and seasonality being confounded. We quantified the interannual variation in raccoon and coyote activity within each study area based on the year of the survey (Table 2). However, since study areas were not surveyed during the same seasons from year to year, the variation that we found could be attributed to either season or year. Furthermore, coyote density and thus activity may fluctuate by year or season. The number of detections of a species is correlated with the trap success, and in our study coyote trap success was fairly consistent within study areas (but see HMC'16 and SNWR'18) (Table S1). When comparing study areas using aggregated raccoon and coyote

activity across surveys, the differences in the carnivore community (as mentioned in the methods) at each study area could be a further confounding factor.

5 | CONCLUSION

We conclude that raccoons may shift their time use in the presence of coyotes. However, it is less clear whether this is done to reduce temporal overlap with coyotes. On the surface, our results seemingly contradict recent works that suggest that coyotes are not an important intraguild predator for raccoons, and that raccoons thus do not partition time to avoid coyotes (Gehrt and Clark 2003; Chitwood et al. 2020). Instead, we suggest that time use shifts may be at a fine scale, and whether they are present depends on a suite of factors. Therefore for a behaviorally plastic species such as the raccoon, it is difficult to make broad conclusions about time use without considering the considerable variation across the urban-rural gradient they inhabit. Similarly for the covote, their role as an intraguild aggressor for raccoons is not static across the urban-rural gradient. Instead, the competitive dominance of covotes is likely dependent on the amount of human pressure and the presence of other larger competitors. Ultimately, as the human footprint on the planet continues to deepen, we need to continue reevaluating interactions across the gradient that it creates. The paradigm in conservation is also shifting to include *in situ* conservation of species in urban habitats, rather than considering these areas solely as suboptimal sink habitats (Magle et al. 2012; Athreya et al. 2013; Mormile and Hill 2017). Studies comparing the ecological roles of species within a community between urban and natural systems are timely. Such work will prove invaluable in understanding how wildlife communities in these novel habitats differ not just in composition, but also in their function.

ACKNOWLEDGMENTS

First, we recognize implementing our camera survey and field research was conducted on lands originally belonging to the People of the Three Fires. Our work was made possible by the generous support and permission of the Detroit Metro Parks, the Shiawassee National Wildlife Refuge, the University of Michigan Biological Station, and the Huron Mountain Club. We also thank the National Science Foundation (IOS #2140322) and the Huron Mountain Wildlife Foundation for financial support. We would like to thank the Applied Wildlife Ecology Lab at the University of Michigan for assistance with fieldwork, image classification and overall project feedback. In particular, K. Mills and S. Gámez provided valuable commentary and general edits. We also thank the countless volunteers for assistance with fieldwork including G. Gadsden, T. Carey, H. Boone and Michigan ZoomIN online community.

AUTHORS' CONTRIBUTIONS

NCH designed the sampling protocol and secured funding, NCH and RM formulated the idea, NCH, RM, and SL carried out fieldwork and data processing, RM carried out analyses and wrote the manuscript, NH and SL provided heavy editorial duties.

CONFLICT OF INTEREST

All authors declare that they have no conflict of interest.

DATA ACCESSIBILITY

The datasets used during this study can be accessed at https://doi.org/10.5061/dryad.hx3ffbghc

REFERENCES

Adams RA, Thibault KM. 2006. Temporal resource partitioning by bats at water holes. J Zool. 270(3):466–472. doi:10.1111/j.1469-7998.2006.00152.x.

Agostinelli C, Lund U. 2017. R package 'circular': circular statistics (version 0.4-93). URL Httpsr-Forge R-Proj Orgprojectscircular.

Athreya V, Odden M, Linnell JDC, Krishnaswamy J, Karanth U. 2013. Big Cats in Our Backyards: Persistence of Large Carnivores in a Human Dominated Landscape in India. PLOS ONE. 8(3):e57872. doi:10.1371/journal.pone.0057872.

Avilés-Rodríguez KJ, Kolbe JJ. 2019. Escape in the city: urbanization alters the escape behavior of Anolis lizards. Urban Ecosyst. 22(4):733–742. doi:10.1007/s11252-019-00845-x.

Batzli GO. 1992. Dynamics of Small Mammal Populations: A Review. In: McCullough DR, Barrett RH, editors. Wildlife 2001: Populations. Dordrecht: Springer Netherlands. p. 831–850. [accessed 2020 Jul 1]. https://doi.org/10.1007/978-94-011-2868-1_63.

Beasley JC, Olson ZH, Dharmarajan G, Eagan TS, Rhodes OE. 2011. Spatio-temporal variation in the demographic attributes of a generalist mesopredator. Landsc Ecol. 26(7):937–950.

Bekoff M, Gese EM. 2003. Coyote (Canis latrans). USDA Natl Wildl Res Cent-Staff Publ.:224.

Berger J. 2007. Fear, human shields and the redistribution of prey and predators in protected areas. Biol Lett. 3(6):620-623.

Breck SW, Poessel SA, Mahoney P, Young JK. 2019. The intrepid urban coyote: a comparison of bold and exploratory behavior in coyotes from urban and rural environments. Sci Rep. 9(1):1–11.

Chitwood MC, Lashley MA, Higdon SD, DePerno CS, Moorman CE. 2020. Raccoon Vigilance and Activity Patterns When Sympatric with Coyotes. Diversity. 12(9):341. doi:10.3390/d12090341.

Ciuti S, Northrup JM, Muhly TB, Simi S, Musiani M, Pitt JA, Boyce MS. 2012. Effects of humans on behaviour of wildlife exceed those of natural predators in a landscape of fear. PloS One. 7(11):e50611.

Clinchy M, Zanette LY, Roberts D, Suraci JP, Buesching CD, Newman C, Macdonald DW. 2016. Fear of the human "super predator" far exceeds the fear of large carnivores in a model mesocarnivore. Behav Ecol. 27(6):1826–1832.

Colborn AS, Kuntze CC, Gadsden GI, Harris NC. 2020. Spatial variation in diet-microbe associations across populations of a generalist North American carnivore. J Anim Ecol. 89(8):1952–1960. doi:10.1111/1365-2656.13266.

Crooks KR. 2002. Relative Sensitivities of Mammalian Carnivores to Habitat Fragmentation. Conserv Biol. 16(2):488–502. doi:10.1046/j.1523-1739.2002.00386.x.

Crooks KR, Soulé ME. 1999. Mesopredator release and avifaunal extinctions in a fragmented system. Nature. 400(6744):563–566.

Demeny K, McLoon M, Winesett B, Fastner J, Hammerer E, Pauli JN. 2019. Food subsidies of raccoons (Procyon lotor) in anthropogenic landscapes. Can J Zool. 97(7):654–657.

Donadio E, Buskirk SW. 2006. Diet, morphology, and interspecific killing in Carnivora. Am Nat. 167(4):524–536.

Drake MD, Peterson MN, Griffith EH, Olfenbuttel C, Moorman CE, Deperno CS. 2019. Hunting interacts with socio-demographic predictors of human perceptions of urban coyotes. Wildl Soc Bull. 43(3):447–454. doi:10.1002/wsb.993.

Elbroch LM, Kusler A. 2018. Are pumas subordinate carnivores, and does it matter? PeerJ. 6:e4293.

Ellington EH, Gehrt SD. 2019. Behavioral responses by an apex predator to urbanization. Behav Ecol. 30(3):821–829.

Elmhagen B, Rushton SP. 2007. Trophic control of mesopredators in terrestrial ecosystems: top-down or bottom-up? Ecol Lett. 10(3):197–206.

Fedriani JM, Fuller TK, Sauvajot RM, York EC. 2000. Competition and intraguild predation among three sympatric carnivores. Oecologia. 125(2):258–270.

Flores-Morales M, Vázquez J, Bautista A, Rodríguez-Martínez L, Monroy-Vilchis O. 2019. Response of two sympatric carnivores to human disturbances of their habitat: the bobcat and coyote. Mammal Res. 64(1):53–62.

Fowler NL, Kautz TM, Petroelje TR, Wilton CM, Kellner KF, O'Brien DJ, Parsons B, Beyer Jr. DE, Belant JL. 2021. Marginal support for a trophic cascade among sympatric canids in peripheral wolf range. Ecology. 102(11):e03494. doi:10.1002/ecy.3494.

Gallo T, Fidino M, Lehrer EW, Magle S. 2019. Urbanization alters predator-avoidance behaviours. Loison A, editor. J Anim Ecol. 88(5):793-803. doi:10.1111/1365-2656.12967.

Gámez S, Potts A, Mills KL, Allen AA, Holman A, Randon PM, Linson O, Harris NC. 2022. Downtown diet: a global meta-analysis of increased urbanization on the diets of vertebrate predators. Proc R Soc B. 289(1970):20212487.

Gaynor KM, Hojnowski CE, Carter NH, Brashares JS. 2018. The influence of human disturbance on wildlife nocturnality. Science. 360(6394):1232–1235. doi:10.1126/science.aar7121.

Gehrt SD, Prange S. 2007. Interference competition between coyotes and raccoons: a test of the mesopredator release hypothesis. Behav Ecol. 18(1):204–214. doi:10.1093/beheco/arl075.

Gordon CE, Feit A, Grüber J, Letnic M. 2015. Mesopredator suppression by an apex predator alleviates the risk of predation perceived by small prey. Proc R Soc B Biol Sci. 282(1802):20142870.

Green AM, Barnick KA, Pendergast ME, Şekercioğlu ÇH. 2022. Species differences in temporal response to urbanization alters predator-prey and human overlap in northern Utah. Glob Ecol Conserv. 36:e02127. doi:10.1016/j.gecco.2022.e02127.

Inouye DW. 1978. Resource partitioning in bumblebees: experimental studies of foraging behavior. Ecology. 59(4):672–678.

Kays R. 2018. Canis latrans. IUCN Red List Threat Species E T3745A103893556 Doi. 10:2018-2.

Lesmeister DB, Nielsen CK, Schauber EM, Hellgren EC. 2015. Spatial and temporal structure of a mesocarnivore guild in midwestern north America. Wildl Monogr. 191(1):1–61. doi:10.1002/wmon.1015.

Lewis JS, Bailey LL, VandeWoude S, Crooks KR. 2015. Interspecific interactions between wild felids vary across scales and levels of urbanization. Ecol Evol. 5(24):5946–5961. doi:10.1002/ece3.1812.

Magle SB, Hunt VM, Vernon M, Crooks KR. 2012. Urban wildlife research: past, present, and future. Biol Conserv. 155:23–32.

Magle SB, Simoni LS, Lehrer EW, Brown JS. 2014. Urban predator-prey association: coyote and deer distributions in the Chicago metropolitan area. Urban Ecosyst. 17(4):875–891. doi:10.1007/s11252-014-0389-5.

Manlick PJ, Pauli JN. 2020. Human disturbance increases trophic niche overlap in terrestrial carnivore communities. Proc Natl Acad Sci. 117(43):26842–26848. doi:10.1073/pnas.2012774117.

Marcus Rowcliffe J, Carbone C, Jansen PA, Kays R, Kranstauber B. 2011. Quantifying the sensitivity of camera traps: an adapted distance sampling approach. Methods Ecol Evol. 2(5):464–476.

Marzluff JM. 2001. Worldwide urbanization and its effects on birds. In: Avian ecology and conservation in an urbanizing world. Springer. p. 19-47.

McClennen N, Wigglesworth RR, Anderson SH, Wachob DG. 2001. The Effect of Suburban and Agricultural Development on the Activity Patterns of Coyotes (Canis Latrans). Am Midl Nat. 146(1):27–36.

McDonnell MJ, Pickett ST. 1990. Ecosystem structure and function along urban-rural gradients: an unexploited opportunity for ecology. Ecology. 71(4):1232–1237.

Mormile JE, Hill CM. 2017. Living with urban baboons: Exploring attitudes and their implications for local baboon conservation and management in Knysna, South Africa. Hum Dimens Wildl. 22(2):99–109.

Newsome TM, Dellinger JA, Pavey CR, Ripple WJ, Shores CR, Wirsing AJ, Dickman CR. 2015. The ecological effects of providing resource subsidies to predators. Glob Ecol Biogeogr. 24(1):1–11. doi:10.1111/geb.12236.

Newsome TM, Howden C, Wirsing AJ. 2019. Restriction of anthropogenic foods alters a top predator's diet and intraspecific interactions. J Mammal. 100(5):1522–1532.

Nix JH, Howell RG, Hall LK, McMillan BR. 2018. The influence of periodic increases of human activity on crepuscular and nocturnal mammals: Testing the weekend effect. Behav Processes. 146:16–21. doi:10.1016/j.beproc.2017.11.002.

Oro D, Genovart M, Tavecchia G, Fowler MS, Martínez-Abraín A. 2013. Ecological and evolutionary implications of food subsidies from humans. Ecol Lett. 16(12):1501–1514.

Prange S, Gehrt SD, Wiggers EP. 2003. Demographic Factors Contributing to High Raccoon Densities in Urban Landscapes. J Wildl Manag. 67(2):324–333. doi:10.2307/3802774.

Prugh LR, Stoner CJ, Epps CW, Bean WT, Ripple WJ, Laliberte AS, Brashares JS. 2009. The rise of the mesopredator. Bioscience. 59(9):779–791.

Randa LA, Yunger JA. 2006. Carnivore occurrence along an urban-rural gradient: a landscape-level analysis. J Mammal. 87(6):1154–1164.

Roemer GW, Gompper ME, Van Valkenburgh B. 2009. The Ecological Role of the Mammalian Mesocarnivore. BioScience. 59(2):165–173. doi:10.1525/bio.2009.59.2.9.

Ruprecht J, Eriksson CE, Forrester TD, Spitz DB, Clark DA, Wisdom MJ, Bianco M, Rowland MM, Smith JB, Johnson BK. 2021. Variable strategies to solve risk–reward tradeoffs in carnivore communities. Proc Natl Acad Sci. 118(35):e2101614118.

Santos F, Carbone C, Wearn OR, Rowcliffe JM, Espinosa S, Lima MGM, Ahumada JA, Gonçalves ALS, Trevelin LC, Alvarez-Loayza P, et al. 2019. Prey availability and temporal partitioning modulate felid coexistence in Neotropical forests. PloS One. 14(3):e0213671.

Schuette P, Wagner AP, Wagner ME, Creel S. 2013. Occupancy patterns and niche partitioning within a diverse carnivore community exposed to anthropogenic pressures. Biol Conserv. 158:301–312.

Sévêque A, Gentle LK, Vicente López-Bao J, Yarnell RW, Uzal A. 2022. Impact of human disturbance on temporal partitioning within carnivore communities. Mammal Rev. 52(1):67–81.

Shedden JM, Bucklin DM, Quinn NM, Stapp P. 2020. Do Coyotes Eat Mesocarnivores in Southern California? A Molecular Genetic Analysis. Proc Vertebr Pest Conf. 29(29). [accessed 2021 Oct 15]. https://escholarship.org/uc/item/1023p03m.

Shores CR, Dellinger JA, Newkirk ES, Kachel SM, Wirsing AJ. 2019. Mesopredators change temporal activity in response to a recolonizing apex predator. Behav Ecol. 30(5):1324–1335. doi:10.1093/beheco/arz080.

Smith JA, Suraci JP, Clinchy M, Crawford A, Roberts D, Zanette LY, Wilmers CC. 2017. Fear of the human 'super predator'reduces feeding time in large carnivores. Proc R Soc B Biol Sci. 284(1857):20170433.

Sovie AR, Greene DU, Frock CF, Potash AD, McCleery RA. 2019. Ephemeral temporal partitioning may facilitate coexistence in competing species. Anim Behav. 150:87–96.

Stark JR, Aiello-Lammens M, Grigione MM. 2020. The effects of urbanization on carnivores in the New York metropolitan area. Urban Ecosyst. 23(2):215–225. doi:10.1007/s11252-019-00923-0.

Suraci JP, Clinchy M, Dill LM, Roberts D, Zanette LY. 2016. Fear of large carnivores causes a trophic cascade. Nat Commun. 7(1):1–7.

Suraci JP, Clinchy M, Zanette LY, Wilmers CC. 2019. Fear of humans as apex predators has landscape-scale impacts from mountain lions to mice. Ecol Lett. 22(10):1578–1586.

Timm R, Cuarón A, Reid F, Helgen K, González-Maya J. 2017. Procyon lotor. The IUCN Red List of Threatened Species 2016: e. T41686A45216638.

Tucker MA, Böhning-Gaese K, Fagan WF, Fryxell JM, Moorter BV, Alberts SC, Ali AH, Allen AM, Attias N, Avgar T, et al. 2018. Moving in the Anthropocene: Global reductions in terrestrial mammalian movements. Science. 359(6374):466–469. doi:10.1126/science.aam9712.

Urban MC, Skelly DK, Burchsted D, Price W, Lowry S. 2006. Stream communities across a rural–urban landscape gradient. Divers Distrib. 12(4):337–350.

Van Donselaar JL, Atma JL, Kruyf ZA, LaCroix HN, Proppe DS. 2018. Urbanization alters fear behavior in black-capped chickadees. Urban Ecosyst. 21(6):1043–1051. doi:10.1007/s11252-018-0783-5.

Way JG, Auger PJ, Ortega IM, Strauss EG. 2001. Eastern coyote denning behavior in an anthropogenic environment. Northeast Wildl. 56:18–30.

Wirsing AJ, Heithaus MR, Brown JS, Kotler BP, Schmitz OJ. 2021. The context dependence of non-consumptive predator effects. Ecol Lett. 24(1):113–129. doi:10.1111/ele.13614.

Table 1. Temporal activity between study areas using Mardia-Watson-Wheeler test for raccoons (a) and coyotes (b). W is the test statistic (approximately Chi-sq distributed), and associated degrees of freedom and p value are included. Temporal overlap (Δ) coefficients and 95% confidence intervals are also included to assess the overlap in activity patterns between study areas. Temporal activity at each study area was based off the aggregated triggers for all surveys with that study area.

(a) Raccoon

	Site Comparison	W	df	pvalue	Δ_{overlap}	Δ (CI)
	HMC vs UMBS	50.001	2	<0.001	0.791	0.75-0.83
	HMC vs SNWR	40.358	2	<0.001	0.836	0.79-0.87
	HMC vs DMP	80.085	2	<0.001	0.764	0.72-0.80
Г	UMBS vs SNWR	9.654	2	0.008	0.903	0.88-0.93
	UMBS vs DMP	9.54	2	0.008	0.937	0.91-0.96
	SNWR vs DMP	63.218	2	<0.001	0.883	0.87-0.90

(b) Coyote

Site Comparison	W	df	pvalue	Δ_{overlap}	Δ (CI)
HMC vs UMBS HMC vs SNWR HMC vs DMP UMBS vs SNWR UMBS vs DMP	29.793 17.087 1.771 4.679 22.872	2 2 2 2 2	<0.001 <0.001 0.412 0.096 <0.001	0.781 0.830 0.938 0.898 0.815	0.71-0.85 0.77-0.88 0.90-0.97 0.84-0.94 0.75-0.88
SNWR vs DMP	10.963	2	0.004	0.849	0.79-0.90

Table 2. Mardia-Watson-Wheeler test results comparing temporal activity for raccoons (a) and coyotes (b) at each study area between each survey year. W is the test statistic (approximately Chi-sq distributed), and associated degrees of freedom and p value are included.

Years	Site	W	df	pvalue
2019 vs 2018	HMC	3.030	2	0.220
2019 vs 2017	HMC	5.826	2	0.054
2019 vs 2016	HMC	0.228	2	0.892
2018 vs 2017	HMC	22.99	2	<0.001
2018 vs 2016	HMC	1.927	2	0.381
2017 vs 2016	HMC	6.77	2	0.034
2016 vs 2015	UMBS	5.533	2	0.063
2018 vs 2017	SNWR	35.319	2	<0.001
2018 vs 2016	SNWR	61.836	2	<0.001
2017 vs 2016	SNWR	26.202	2	<0.001
2020 vs 2018	DMP	7.948	2	0.018
2020 vs 2017	DMP	6.5761	2	0.037
2018 vs 2017	DMP	9.884	2	0.007

Years	Site	W	df	pvalue
2019 vs 2018	HMC	4.436	2	0.109
2019 vs 2017	HMC	4.836	2	0.891
2019 vs 2016	HMC	11.043	2	0.004
2018 vs 2017	HMC	0.975	2	0.614
2018 vs 2016	HMC	2.543	2	0.281
2017 vs 2016	HMC	3.884	2	0.143
2016 vs 2015	UMBS	5.471	2	0.649
2018 vs 2017	SNWR	0.098	2	0.952
2018 vs 2016	SNWR	1.665	2	0.435
2017 vs 2016	SNWR	1.214	2	0.545
2020 vs 2018	DMP	15.187	2	<0.001
2020 vs 2017	DMP	11.27	2	0.004
2018 vs 2017	DMP	0.741	2	0.690

Table 3. Temporal overlap (Δ) coefficients and 95% confidence intervals for raccoon and coyote activity in low and high coyote zones within each camera survey in Michigan from 2016-2020. The overlap coefficients for raccoons and coyotes here should be interpreted with caution, as splitting the coyote detections into low and high zones resulted in low numbers of detections in the low zones (and is reflected in the wide confidence intervals). Mardia-Watson-Wheeler test results comparing raccoon activity between the top quartile and the bottom three quartiles of raccoon activity for each survey are contained in the last two columns, where W is the test statistic (approximately Chi-sq distributed), and *p* value are included.

Survey period	Site/Year	Δ (CI) High	Δ (CI) Low	W	pvalue
May-Aug	HMC'19	0.32-0.72	0.28-0.85	10.024	<0.001
Jun-Aug	HMC'18	0.49-0.76	0.68-0.91	15.122	0.007
Jul-Jun	HMC'17	0.65-0.83	0.56-0.80	3.841	0.147
Jun-Oct	HMC'16	0.59-0.85	0.45-0.83	0.918	0.632
Jul-Nov	UMBS'16	0.65-0.85	0.71-0.87	9.631	0.008
Oct-Dec	UMBS'15	0.42-0.73	0.66-0.86	7.392	0.025
Sep-Dec	SNWR'18	0.59-0.78	0.66-0.89	10.458	0.005
May-Aug	SNWR'17	0.60-0.84	0.54-0.88	3.647	0.162
Feb-May	SNWR'16	0.63-0.79	0.61-0.81	6.086	0.048
Jan-Sep	DMP'20	0.46-0.58	0.61-0.80	3.302	0.192
Oct-Feb	DMP'18	0.52-0.75	0.61-0.85	0.376	0.829
Nov-Mar	DMP'17	0.59-0.83	0.57-0.84	0.692	0.708

FIGURE LEGENDS

Fig. 1 The three comparisons considered within our study: a) Raccoon temporal activity was compared between study areas; b) raccoon temporal activity was compared between years, and across zones of coyote intensity of use within each study area; c) raccoon temporal activity results from the within study area comparisons in response to coyotes were compared across study areas.

Fig. 2 Study sites across Michigan. From north to south, the Huron Mountain Club (HMC), the University of Michigan Biological Station (UMBS), the Shiawassee National Wildlife Refuge (SNWR), and the Detroit Metroparks (DMP). Example coyote spatial activity kernel density hotspots are included for each site; hotspots in coyote detections varied by year, and KD maps were generated for each survey.

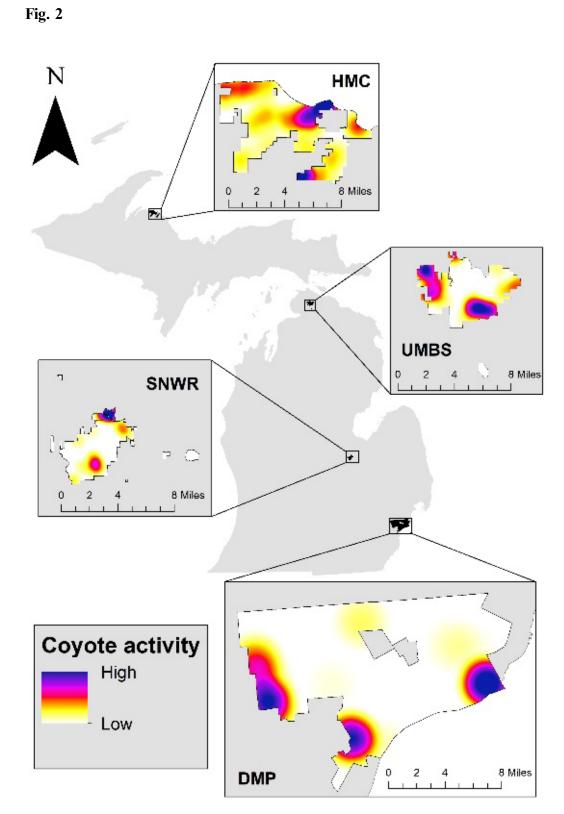
Fig. 3 a) Raccoon activity across all four study areas. Time use of raccoons was summed for all surveys within a study area; b) Coyote activity across all four study areas. Time use of coyotes was summed for all surveys within a study area.

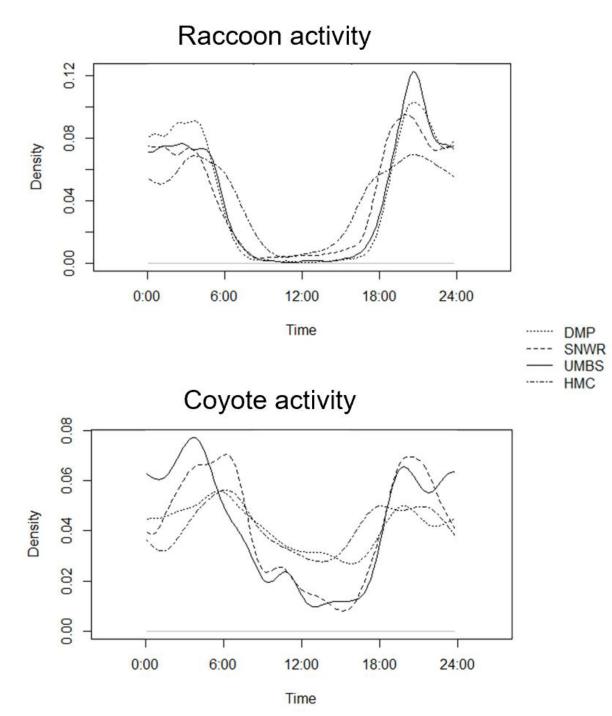
Fig. 4 Overlap plots with raccoon temporal activity in high and low coyote zones plotted against aggregated coyote temporal activity for the survey at the opposite ends of the urban-rural gradient for 2017 and 2018 (the two years during which both DMP and HMC were surveyed).

Fig. 5 Mean temporal overlap ($\Delta_{temporal}$) between raccoons and coyotes in high and low spatial zones of coyote activity with 95% confidence intervals.

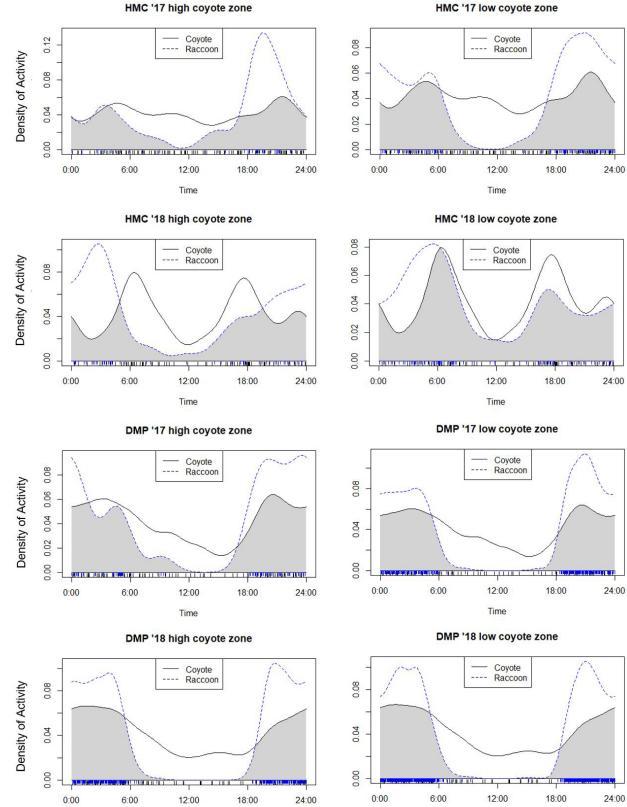
С а High Coyote Zone Low Coyote Zone Urban Density of Activity Density of Activity Rural 03 12:00 24:00 0:00 0:00 12:00 24:00 Within site b Seasonality/ Biotic Interannual Density of Activity ĸ 0:00 12:00 24:00

Between sites





r Manuscrip utho



Time

HMC '17 low coyote zone

Time

Fig. 4



