

Multivariate measure of niche overlap using canonical correspondence analysis¹

Janice L. PAPPAS & Eugene F. STOERMER, Center for Great Lakes and Aquatic Sciences, University of Michigan, Ann Arbor, Michigan, 48109-2099, U.S.A., e-mail: jlpappas@umich.edu

Abstract: A measure of multidimensional niche overlap based on species scores from canonical correspondence analysis is presented. This measure, N, is the amount of overlap using species scores as approximate weighted averages with respect to multidimensional niche breadth for the community given by $N = 0 - [(u_i - u_j)^t V^{-1} (u_i - u_j) g | V]$. Multidimensional niche overlap was calculated for a representative periphytic community from northern Lake Huron. Classification of species interaction or species similarities was accomplished using cluster analysis of the multidimensional niche overlap values. Two major groups of taxa showing similarities were evident based on niche influence. This influence produced a gradation of species found in clean waters to those found in varying degrees of pollution.

Keywords: niche overlap, canonical correspondence analysis, periphyton, diatoms, multidimensional, cluster analysis.

Résumé: Nous présentons ici une mesure du chevauchement de niches à dimensions multiples reposant sur les coordonnées des points des espèces obtenues d'une analyse canonique des correspondances. Cette mesure, N, est une évaluation du chevauchement à partir du pointage des espèces comme approximation des moyennes pondérées en regard de l'étendue des niches à dimensions multiples pour la communauté. Elle a été obtenue de l'équation suivante : $N = 0 - [(u_i - u_j)^i V^{-1} (u_i - u_j) g IV]$. Le chevauchement de niches à dimensions multiples a été calculé pour une communauté périphytique du nord du lac Huron. Le classification de l'action réciproque ou des similitudes entre espèces a été effectuée par le biais d'une analyse de groupement à partir des valeurs de chevauchement de niches à dimensions multiples. Nous avons pu mettre en évidence deux groupes principaux de taxons montrant des similitudes sur la base de l'influence des niches. Cette influence a généré une gamme d'espèces, depuis celles qui sont associées aux eaux propres aux eaux polluées à divers degrés.

Mots-clés: chevauchement de niche, analyse canonique des correspondances, périphyton, diatomées, dimensions multiples, analyse de groupement.

Introduction

Species interactions and community organization are important determinants of how species coexist, survive and thrive. Community structure can be elucidated by examining niche breadth and overlap in terms of niche dimension and shape. Niche dimensionality refers to the number of environmental factors which separate species rather than all the biologically relevant influences in the environment (Levins, 1968). Niche shape refers to the degree of probability that species will interact in an n-dimensional environment; thus, this metric refers to various degrees of species' similarities.

A niche is all the ways an organism needs to adapt to its environment in order to survive (Pianka, 1978). Adaptation is a dynamic process in terms of resource availability, physical and chemical influences and biological interactions, all of which can be assessed spatially and temporally. To analyze niche on a multidimensional basis, ordination techniques such as principal components analysis (Rotenberry & Wiens, 1980) and discriminant analysis (Green, 1971; 1974; Harner & Whitmore, 1977; Dueser & Shugart, 1979; Van Horne & Ford, 1982) have been used. In both of these techniques, environmental gradients are inferred from analysis.

Other ordination techniques which may seem to be useful have restrictions that make them not entirely applicable to multidimensional niche analysis. To use canonical correlation analysis, the number of species cannot be approximately the same as the number of samples (ter Braak, 1990).

Redundancy analysis (RA) is sensitive to scaling, being dependent on absolute abundance (ter Braak, 1990), and canonical scores are based on weighted sums (Jongman, ter Braak & van Tongeren, 1987). By using correspondence analysis, there is still the result that environmental gradients are inferred and the problem of the arch effect makes interpretation of all but the first eigenvector difficult (ter Braak & Prentice, 1988), whether or not detrending is used (Greenacre, 1984).

Alternatively, canonical correspondence analysis (CCA), based on unimodal species response curves, includes the relation between species and their environment in canonical space (ter Braak, 1986). Species scores, as approximate weighted averages, are useful descriptors of favorable habitat and niche space (Prentice & Cramer, 1990). Each constrained axis is a dimension and can be identified as a gradient influenced by an environmental variable represented as a vector in the ordination diagram. Eigenvalues represent measures of species separation or dispersion for each canonical axis (Jongman, ter Braak & van Tongeren, 1987).

The entire ordination is a representation of a sample of a larger population. Constrained eigenvalues represent fractions of the variance. The first constrained eigenvector determines the direction in which the sample from the population has maximum variation. That is, variances are not independent from eigenvector to eigenvector (Van Valen, 1978).

In the CCA ordination, scores of environmental variables influence the gradient of each canonical axis to a greater or lesser

degree, but not completely. Therefore, each environmental vector is meaningful relative to every other environmental vector. The order of species scores for one environmental variable occurs relative to those for every other environmental variable.

Species axes are sample scores derived from species scores by weighted averaging. Environmental axes are sample scores regressed on environmental variables. In CCA, the weighted correlation matrix of ordination axes shows interdependence between species axes as a positive or negative correlation and independence between environmental axes as a value of 0.000. The result is an interdependence among species such that vectors of species scores over *n*-axes may be quantitatively compared.

The method presented will provide a multidimensional niche overlap measure of which the result will include quantitative representation of the species-environment relation as well as the degree to which species share tolerances to environmental conditions or interspecies relations. These values may be used in additional quantitative methods to assess species interaction or species similarities with respect to environmental tolerances. Multidimensional niche overlap values will be shown to be a useful way to assess relations between dominant as well as rare species in a community.

Methods

MULTIDIMENSIONAL NICHE OVERLAP

Niche space may be described by closed geometric curves or surfaces in *n*-dimensional space. That is, with the assumption of normality on a multivariate basis, niche boundaries may be described as probability ellipses for two dimensions, ellipsoids for three dimensions and hyperellipsoids for more than three dimensions. In CCA, the length of a multivariate probability ellipse in a given dimension is proportional to the standard deviation, which represents relative niche breadth on the given canonical function.

Niche overlap is the actual overlap of probability ellipses for two species. The exponent of the multivariate normal density function measures squared distance. The *n*-dimensional normal density function is:

$$f(x) = \frac{1}{(2\pi)^{n/2} |V|^{1/2}} e^{-(x-\mu)^{\nu} V^{-1}(x-\mu)}$$
[1]

where V is the variance-covariance matrix, x is the canonical score and μ is the mean. Concentration ellipses generated from the estimate of the exponent of this function for two $n \times 1$ vectors (u_1 and u_2) of species scores on several constrained eigenvectors is:

$$(u_1 - u_2)^{*} V^{-1}(u_1 - u_2)$$
 [2]

Vectors of species scores are vectors of approximate weighted averages. Niche distance is obtained by using the generalized distance (Mahalanobis, 1948) between these vectors of weighted means. This modification is used since the distance measured is between species rather than between a species and the overall mean of the population (or habitat). The distance takes into account covariances between species. Generalization of the two species vector case becomes:

$$D_{ij}^{2} = (u_{i} - u_{j})^{*} V^{-1}(u_{i} - u_{j})$$
 [3]

where u_i and u_j represent vectors for species i and j, respectively. Overlap of species tolerance ranges is a probability because the generalized distance is distributed $\Re \chi_n^2(\alpha)$ with n degrees of freedom (Johnson & Wichern, 1992). Probability contours can then be defined as $(1 - \alpha)100$. That is, niche overlap values are probability values with respect to degree of overlap.

Distances calculated between species may not comprise a linear set (Bernstein, 1988). That is, the shortest distance may be factors of ten smaller than the longest distance. To bring the distances into a common scale, a scaling factor, g, is multiplied by the distances. In this way, no one pair of distances will be unduly influential. That is,

$$(u_i - u_j)' V^{-1}(u_i - u_j) g$$
 [4]

For n-dimensions there are n-environmental variables affecting the species from a given number of samples. The variance-covariance matrix, V, of species canonical scores as approximate weighted averages per n-dimensions has n-variances and 1/2n(n-1) potentially different covariances. A single numerical value which describes the variation expressed by this matrix is the determinant or generalized variance. The generalized variance is equal to the product of the individual species variances or squared niche breadths. Geometrically, for n-dimensions, the generalized variance is proportional to the square of the volume of the ellipsoid generated by the n-deviation vectors (Johnson & Wichern, 1992). That is,

$$(u_i - u_i)^{\perp} V^{-1}(u_i - u_i) \alpha |V|$$
 [5]

Conceptually, the generalized variance is a measure of multidimensional niche breadth (Carnes & Slade, 1982) of the community sampled. If the generalized variance is large, species are spread out or highly dispersed; if it is small, species are close together or not very dispersed.

The effect of multidimensional niche breadth on niche distances and the probability that species' tolerance ranges overlap in the *n*-dimensional space is:

$$N = 0 - \left[\left(u_i - u_j \right)^{i} V^{-1} \left(u_i - u_j \right) g |V| \right]$$
 [6]

where N is multidimensional niche overlap between species, 0 is the matrix of 1's, and g is a scaling factor. Each multidimensional niche overlap value, when multiplied by 100, represents the percentage of overlap between species affected by n-environmental variables or species interaction as measured at that time.

APPLICATION AND ILLUSTRATION

A small periphytic community from northern Lake Huron is used for illustration. On August 26, 1968, sixteen species of diatoms were identified from a shallow water sample from a locality known locally as Big Tub, Tobermory Lt., Ontario, Canada, 45° 13' N, 82°, 20' w. The abundances of Cocconeis pediculus Ehrenb., Gomphonema olivaceum var. calcarea (Cl.) Cl., other Gomphonema. spp., Rhoicosphenia curvata (Kütz.) Grun. ex Rabh., Diatoma vulgare Bory., D. tenue var. elongatum Lyngb., Cymbella prostrata var. auerswaldii (Rabh.) Reim. in Patr. & Reim., C. microcephala

Grun., Gomphoneis herculeana (Ehrenb.) Cl., Achnanthes minutissima Kütz., other Achnanthes. spp., Fragilaria vaucheriae var. capitellata (Grun.) Patr. in Patr. & Reim., F. intermedia Grun. in V. H., F. pinnata Ehrenb., Cyclotella comta (Ehrenb.) Kütz., and Navicula sp. #44 were used for analysis. At the time, identification of the morphologically variable Cyclotella comta was made according to Hustedt (1930). Gomphonema spp. and Achnanthes spp. represent groups of diatoms not identifiable at the species level. Navicula sp. #44 is indigenous to the Great Lakes and has been given a numerical designation by the Center for Great Lakes and Aquatic Sciences. A Cocconeis pediculus auxospore was also found. Cocconeis pediculus and Gomphonema olivaceum var. calcarea were the dominant forms accounting for 77.3 and 16.1% of the community, respectively. Abundances of periphytic samples in this location were incorporated into the species used for analysis (Stoermer, unpubl. data).

On July 23, 1969, physical and chemical parameters were measured in surface waters off shore from Tobermory Lt. Of the parameters measured, those used in this analysis were temperature, turbidity, conductivity, pH, alkalinity, total residue, organic nitrogen, ammonia, nitrate, total phosphorus, chloride and iron. During the summer of 1969, other samples at the surface were taken and the data were incorporated in the environmental parameters used for this analysis (unpubl. STORET data, Canada, 1969).

During the time period from summer 1968 to summer 1969, little historical data are available on water quality and algal abundances for northern Lake Huron. However, results of a comprehensive water quality study indicated that in the late 1960's to early 1970's, northern Lake Huron was found to be oligotrophic (Dobson, Gilbertson & Sly, 1974). Values reported for alkalinity and total phosphorus (Dobson, Gilbertson & Sly, 1974) are consistent with those used in this analysis. In addition, dissolved oxygen concentrations (Dobson, Gilbertson & Sly, 1974) are characteristic of water conditions reported for summer 1969 environmental data used in this study (unpubl. STORET data, Canada, 1969).

For the late 1960's, few publications exist on algal abundances in northern Lake Huron. However, in 1969, a listing of planktonic and periphytic algae was compiled of species found in the Great Lakes (Michalski, 1969). Qualitatively, for Lake Huron, those species in this 1969 listing (Michalski, 1969) are consistent with most of those species found in the 1968 periphytic composition and abundance used in our study.

The data sets from northern Lake Huron were chosen to illustrate the method of calculating multidimensional niche overlap values since the resultant matrix could be kept to a minimum. Presumably, any species-environmental data set could be used. However, since the number of species squared determines the size of the multidimensional niche overlap matrix, the need for a small species data set as an example of implementing this method should be evident.

To maintain a 16 by 16 multidimensional niche overlap matrix of 256 elements, from the data sets, a simulation was performed to generate species and environmental data for fourteen samples with data values randomly obtained. For each species abundance value, a range was determined based on relative abundance. For each environmental variable measurement, a range was determined based on variation for the season. In the past, simulations have been used in developing ecological methods (see, for example, LaFrance, 1972; Gauch & Whittaker, 1972).

Both RA and CCA were performed on the data. RA was used to determine the general relation between species and environmental variables. From CCA, species scores as approximate weighted averages were used to calculate multidimensional niche overlap. For illustration, to indicate interspecies relation or species similarity with respect to environmental tolerances, cluster analysis was used to classify multidimensional niche overlap values. However, other quantitative methods (such as multidimensional scaling) may be used to further characterize multidimensional niche overlap values.

Results

Since it was known that the assemblage occurred in a narrow range of environmental conditions, RA was performed to assess the relation between species and environmental variables. Constrained eigenvalues were 0.899, 0.040, 0.002 and 0.001 for constrained eigenvectors one, two, three and four, respectively. This represented 95.4, 99.6, 99.8 and 99.9 cumulative percent of the species-environment variation explained for the first four constrained eigenvectors. For RA performed with correction for species error variance, eigenvalues for the first four canonical axes were 0.210, 0.179, 0.138 and 0.116. Cumulative percentage of the species-environment variation explained was 21.0, 39.0, 52.7, and 64.3 for these first four contrained axes. The preliminary analyses indicate an important relation between species and environmental variables.

In CCA, Cocconeis pediculus and Gomphonema olivaceum var. calcarea were analyzed as passive, and therefore downweighted to zero to offset their high degree of influence. These two species were added to the ordination after CCA was performed using transition formulae (ter Braak, 1988). Constrained eigenvalues were 0.078, 0.064, 0.036 and 0.022 for axes 1, 2, 3 and 4, respectively. These eigenvalues represent the variation in the species data explained by the environmental variables and are considered to be informative; with eigenvalues numerically close, those which are less than 0.02 are not reliable (ter Braak, 1988). The cumulative percentage of species-environment variation explained was 32.2, 58.3, 73.1 and 82.0 for the first four canonical axes which was better than the results for RA with correction by the species error variance. Variation not explained by the environmental variables may be spatial since growth habit of these species involves variation of substrate type (Stoermer, 1980).

Species axes were interdependent and environmental axes were independent (Table I). Multidimensional niche overlap values were calculated (Table II). Most of the overlap values were high. In general, Cocconeis pediculus, Gomphonema olivaceum var. calcarea, Cymbella microcephala, Achnanthes minutissima and Fragilaria intermedia were found to overlap least with Rhoicosphenia curvata and

TABLE I. Weighted correlation matrix of species and environmental axes. Species axis-species axis correlations have small values; therefore, interdependence is indicated. Environmental axis-environmental axis correlations have zero values; therefore, independence is indicated. In addition, species axis-environment axis correlations are given

SPEC AX1	1.0000							
SPEC AX2	-0.0060	1.0000						
SPEC AX3	0.0026	-0.0179	1.0000					
SPEC AX4	-0.0004	0.0028	-0.0012	1.0000				
ENVI AX1	0.9996	0.0000	0.0000	0.0000	1.0000			
ENVI AX2	0.0000	0.9790	0.0000	0.0000	0.0000	1.0000		
ENVI AX3	0.0000	0.0000	0.9962	0.0000	0.0000	0.0000	1.0000	
ENVI AX4	0.0000	0.0000	0.0000	0.9999	0.0000	0.0000	0.0000	1.0000
	SPEC AX1	SPEC AX2	SPEC AX3	SPEC AX4	ENVI AX1	ENVI AX2	ENVI AX3	ENVI AX4

Gomphoneis herculeana. Cyclotella comta was found to overlap least only with Rhoicosphenia curvata. Fragilaria vaucheriae var. capitellata, Fragilaria pinnata and Achnanthes spp. were found to overlap least only with Gomphoneis herculeana. Cluster analyses were performed on multidimensional niche overlap values to classifiy similar species in a hierarchical fashion. Although results were the same using different algorithms, Euclidean distance, complete linkage clustering (farthest neighbor) was chosen for its clarity (Figure 1). Elucidation of niche influence on this classification of species will provide an understanding of the relation between similar species with respect to their tolerance of environmental conditions.

Discussion

From cluster analysis, niche influences based on tolerances to environmental conditions were evident in producing separation of species groups. The first major division assigns Gomphoneis herculeana and Rhoicosphenia curvata to one group and all other species to another group. Further division indicates that Gomphoneis herculeana and Rhoicosphenia curvata are joined by Fragilaria vaucheriae var. capitellata, Gomphonema spp., and Diatoma vulgare as a definable group. Of the other group, Navicula sp. #44, Fragilaria intermedia, F. pinnata, Gomphonema olivaceum var. calcarea, Cocconeis pediculus, Cymbella microcephala, Achnanthes spp., and Diatoma tenue var. elongatum form a subgroup. Achnanthes minutissima, Cymbella prostrata var. auerswaldii, and Cyclotella comta form the other subgroup with C. comta as an outlier.

The only centric diatom present, Cyclotella comta, is usually planktonic. This taxon, along with most of the pennates, is characteristically tolerant of pH levels greater than 7 and tolerate small amounts of salt (Patrick & Reimer, 1966; Lowe, 1974; Patrick & Reimer, 1975; Beaver, 1981). Only Achnanthes minutissima has a pH tolerance of 7 (Lowe, 1974). The presence of Achnanthes minutissima may be an indicator of high oxygen content in alkaline waters and is indifferent to calcium and iron (Beaver, 1981).

In addition, Cocconeis pediculus, Cymbella microcephala and Cymbella prostrata var. auerswaldii are found in oxygen rich waters and are calcium indifferent (Beaver, 1981). Fragilaria pinnata and Fragilaria vaucheriae are found in oxygen rich waters (Beaver, 1981). Gomphonema olivaceum var. calcarea and Rhoicosphenia curvata prefer alkaline waters (Patrick & Reimer, 1966; 1975; Beaver, 1981). Gomphonema spp. may be found in waters with low

conductivity and low hardness (Patrick & Reimer, 1975). Some species of *Navicula* prefer iron-rich water (Patrick & Reimer, 1966). Alkalinity, pH, conductivity and chloride

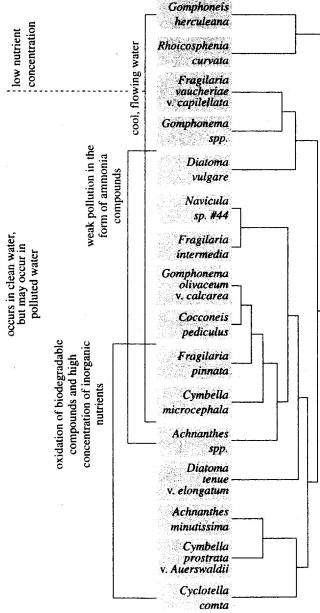


FIGURE 1. Complete linkage clustering of multidimensional niche overlap values for species. Largest group divisions indicate niche influences as tolerance of environmental conditions.

values
overlap
niche
ensiona
ıltidim
Ę
2
Ξ
Щ
ABL
.≺

Diatoma tenue v. elongatum Fragilaria intermedia Fragilaria pinnata	85 0.8825 0.9802 0.9493 31 0.8460 0.9368 0.9812 0.9414 65 0.9125 0.7594 0.7835 0.7255 33 0.7336 0.9294 0.8415 0.7309 91 0.7804 0.6844 0.8165 0.8568 27 0.8627 0.7213 0.5668 0.7048 29 0.8846 0.8901 0.9519 0.8782 23 0.8949 0.8539 0.9543 0.9002 80 0.8225 0.6870 0.7328 0.6351 94 0.7877 0.7601 0.7817 96 0.7479 0.8093 0.9126 0.9029 90 0.7007 0.8808 0.8517 0.8346 75 0.8808 1.0000 0.9431 0.9260 96 0.8517 0.9431 1.0000 0.9411	0.8346 0.9260 0.9411
Achnanthes minutissima Cyclotella comta	0.8733 0.8385 0.9050 0.7931 0.5635 0.6133 0.9181 0.6991 0.5462 0.7727 0.8506 0.6929 0.9142 0.6623 0.5897 0.4580 0.7131 0.7493 1.0000 0.8196 0.7131 0.7037 0.8196 1.0000 0.7479 0.7037	_
Fragilaria vaucheriae v. capitellata	0.8577 0.8706 0.4256 0.6669 0.8088 0.6913 0.8915 0.8915 1.0000 0.7131 0.7493 0.7587	0.7817
dds vuveuoyduo9	0.7826 0.8090 0.5817 0.6279 0.7706 0.6058 0.9085 0.7469 1.0000 0.8915 0.8915 0.4580 0.8225 0.6870	0.6315
dds səyнивинэү	0.8977 0.9177 0.8896 0.6793 0.8848 0.5954 0.9270 0.7469 0.6922 0.6922 0.6623 0.8539 0.8539	0.9002
Cymbella microcephala	0.9555 0.9777 0.7355 0.8046 0.8722 0.6242 1.0000 0.9270 0.8891 0.8891 0.8896 0.8896 0.8896 0.8896 0.8896 0.8896 0.8896 0.8896 0.8896 0.8896	0.8782
Gomphoneis herculeana	0.6967 0.6116 0.6553 0.4980 0.6020 1.0000 0.6242 0.5954 0.6913 0.5953 0.7727 0.8627 0.7727	0.7048
Cymbella prostrata v. auerswaldii	0.7972 0.8521 0.6481 0.6481 1.0000 0.6020 0.8722 0.8722 0.7706 0.8088 0.9181 0.6991 0.7864	0.8568
Diatoma vulgare	0.8950 0.8424 0.5743 1.0000 0.4173 0.6793 0.6659 0.6669 0.6669 0.6669 0.6669 0.7336 0.7336	0.7309
Rhoicosphenia curvata	0.7344 0.6991 1.0000 0.5743 0.6481 0.6553 0.7355 0.8896 0.6789 0.6789 0.6789 0.6789	0.7255
Gomphonema olivaceum v. calcarea	0.9866 1.0000 0.6991 0.8424 0.8521 0.6116 0.9777 0.9177 0.9050 0.7731 0.8090 0.8706 0.9050 0.7931 0.8460	0.9414
Socconeis sulucibad	1.0000 0.9866 0.7344 0.8950 0.7972 0.6967 0.8977 0.8577 0.8773 0.8733 0.8733 0.8733 0.8733 0.8733	0.0405
	Cocconeis pediculus Gomphonema olivaceum v. calcarea Rhoicosphenia curvata Diatoma vulgare Cymbella prostrata v. auerswaldii Gomphoneis herculeana Cymbella microcephala Achnanthes spp. Gomphonema spp. Fragilaria vaucheriae v. capitellata Achnanthes minutissima Cyclotella comta Diatoma tenue v. elongatum Fragilaria intermedia	Navicula en #44

have approximately the same niche influence for all species and were not instrumental in defining niche differences.

The species in this assemblage were found in shallow water and subjected to some wave action. Those species which prefer cool, flowing waters are Gomphonema olivaceum and Gomphoneis herculeana (Patrick & Reimer, 1975) as well as Rhoicosphenia curvata, Diatoma vulgare and Fragilaria vaucheriae (Patrick & Reimer, 1966). Fragilaria vaucheriae var. capitellata (Patrick & Reimer, 1966) and Gomphonema olivaceum var. calcarea (Patrick & Reimer, 1975; Beaver, 1981) have been found in both standing and flowing waters. Fragilaria intermedia has been found in the plankton of river assemblages and is therefore tolerant of both standing and flowing waters (Krieger, 1927). From cluster analysis, these species form a group in which temperature and turbidity (as a proxy for flowing water) are the major niche influences. Alternatively, many of these species may be classified as shallow, littoral, wave-zone taxa (Kingston et al., 1983). All of these species occur in temperatures from 0 to 15°C (Lowe, 1974). Diatoma vulgare does occur as a winter dominant (Beaver, 1981). Although Cymbella prostrata var. auerswaldii is found in streams (Lowe, 1974), cluster analysis reveals that this is not its primary niche influence on a multidimensional basis in this analysis (Figure 1).

Total phosphorus, ammonia, organic nitrogen and nitrate have varying degrees of niche influence on Diatoma vulgare, Navicula sp. #44, Fragilaria intermedia, Gomphonema olivaceum var. calcarea, Cocconeis pediculus, Fragilaria pinnata, Cymbella microcephala, Achnanthes spp., Diatoma tenue var. elongatum, Achnanthes minutissima, Cymbella prostrata var. auerswaldii, and Cyclotella comta (Figure 1). Various Navicula species, Fragilaria intermedia, F. pinnata, Gomphonema olivaceum var. calcarea, Cocconeis pediculus, Diatoma vulgare, and Cymbella microcephala (Lowe, 1974; Beaver, 1981) are influenced by weak pollution in the form of ammonia compounds. Cocconeis pediculus does occur in oligotrophic habitats (Stoermer, 1980).

Fragilaria pinnata, Cymbella microcephala, Achnanthes spp., Diatoma tenue var. elongatum, Achnanthes minutissima, Cymbella prostrata var. auerswaldii, and Cyclotella comta are influenced by complete oxidation of biodegradable compounds and found in habitats with high concentrations of inorganic nutrients (Lowe, 1974). Achnanthes minutissma is tolerant of nutrient enrichment, but found in oligotrophic habitats as well (Stoermer, 1980). Many of these taxa may be found in deep, cold, dark, more stable areas (Kingston et al., 1983).

Fragilaria vaucheriae var. capitellata may be found in oligotrophic waters, and Gomphoneis herculeana may be found in undisturbed areas (Stoermer, 1980). Cocconeis pediculus, Cyclotella comta, Diatoma tenue var. elongatum, various Navicula species (Lowe, 1974), and Diatoma vulgare (Beaver, 1981) are found in both clean and polluted waters. Species tolerant of eutrophy are Cymbella prostrata (Stoermer, 1980), Rhoicosphenia curvata, Diatoma vulgare, Fragilaria vaucheriae, Gomphomena olivaceum, Fragilaria pinnata, and various Navicula species (Lowe, 1974; Beaver, 1981). Diatoma tenue var. elongatum is tolerant of heavy

nutrient and chloride loadings (Stoermer, 1980). Given the environmental conditions, all the species have broad tolerance to pollution. However, there is a general gradation of species found in clean waters to those tolerant of varying degrees of pollution as indicated in cluster analysis (Figure 1).

The species in the assemblage analyzed are more similar than different since they were found in a narrow range of environmental conditions. From species scores in CCA, multidimensional niche overlap values were calculated. Cluster analysis of those values were useful in determining species similarities and exploring the fine differences among the interspecies relation with respect to environmental tolerances in this community. This multivariate method provides a way to include both dominant and rare species in analyses and determines species similarities with to respect to the environmental conditions that are influential. Essentially, by using species scores as "data", calculating covaried distance between these scores and rescaling of this to result in a probability value, this multidimensional niche overlap method is a useful means for analyzing interspecies relations.

Acknowledgements

We thank D. Dolan for providing STORET data of physical and chemical parameters for Tobermory Lt. We thank M. B. Edlund who was most helpful in providing some of the information used in this study. We also thank B. McClellan for making the table of multidimensional niche overlap values.

Literature cited

- Beaver, J., 1981. Apparent Ecological Characteristics of Some Common Freshwater Diatoms. Ontario Ministry of the Environment, Technical Support Section, Central Region, Ontario.
- Bernstein, I. H., 1988. Applied Multivariate Analysis. Springer-Verlag, New York.
- Carnes, B. A. & N. A. Slade, 1982. Some comments on niche analysis in canonical space. Ecology, 63: 888-893.
- Dobson, H. F. H., M. Gilbertson & P. G. Sly, 1974. A summary and comparison of nutrients and related water quality in lakes Erie, Ontario, Huron, and Superior. Journal of the Fisheries Research Board Canada, 31: 731-738.
- Dueser, R. D. & H. H. Shugart, 1979. Niche pattern in a forest-floor small-mammal fauna. Ecology, 60: 108-118.
- Gauch, H. G., Jr. & R. H. Whittaker, 1972. Coenocline simulation. Ecology, 53: 446-451.
- Green, R. H., 1971. A multivariate statistical approach to the Hutchinsonian niche: Bivalve molluscs of central Canada. Ecology, 52: 534-556.
- Green, R. H., 1974. Multivariate niche analysis with temporally varying environmental factors. Ecology, 55: 73-83.
- Greenacre, M. J., 1984. Theory and Applications of Correspondence Analysis. Academic Press, London.
- Harner, E. J. & R. C. Whitmore, 1977. Multivariate measures of niche overlap using discriminant analysis. Theoretical Population Biology, 12: 21-36.
- Hustedt, F., 1930. Die Süsswasserflora Mitteleuropas. Heft 10. 2nd edition. Bacillariophyta (Diatomeae). A. Pascher (ed.). Verlag von Gustav Fischer, Bremen.
- Johnson, R. A. & D. W. Wichern, 1992. Applied Multivariate

- Statistical Analysis. 3rd edition, Prentice Hall, Englewood Cliffs, New Jersey.
- Jongman, R. H. G., C. J. F. ter Braak & O. F. R. van Tongeren, 1987. Data Analysis in Community and Landscape Ecology. Pudoc, Wageningen.
- Kingston, J. C., R. L. Lowe, E. F. Stoermer & T. B. Ladewski, 1983. Spatial and temporal distribution of benthic diatoms in northern Lake Michigan. Ecology, 64: 1566-1580.
- Krieger, W., 1927. Zur Biologie des Flussplanktons. Pflanzenforschung 10: 1-64.
- LaFrance, C. R., 1972. Sampling and ordination characteristics of computer-simulated individualistic communities. Ecology, 53: 387-397.
- Levins, R., 1968. Evolution in Changing Environments: Some Theoretical Explorations. Princeton University Press, Princeton, New Jersey.
- Lowe, R. L., 1974. Environmental requirements and pollution tolerance of freshwater diatoms. EPA-670/4-74-005. U.S. Environmental Protection Agency, National Environmental Research Center, Office of Research and Development, Cincinnati, Ohio.
- Mahalanobis, P. C., 1948. Historical note on the D2-statistic. Sankhya, 9: 237-240.
- Michalski, M. F. P., 1969. Planktonic and periphytic algae of the Great Lakes: A list of recorded species and their distribution, part VI.2. Pages 1-66 in D. V. Anderson (ed.). The Great Lakes as an Environment. Great Lakes Institute. University of Toronto, Toronto, Ontario.
- Patrick, R. M. & C. W. Reimer, 1966. The Diatoms of the United States. Vol. I. Academy of Natural Sciences. Philadelphia Monograph no. 13, Philadelphia, Pennsylvania.
- Patrick, R. M. & C. W. Reimer, 1975. The Diatoms of the United States. Vol. II, part 1. Academy of Natural Sciences.
 Philadelphia Monograph no. 13, Philadelphia, Pennsylvania.
- Pianka, E. R., 1978. Evolutionary Ecology. 2nd edition, Harper & Row, New York.
- Prentice, H. C. & W. Cramer, 1990. The plant community as a niche bioassay: Environmental correlates of local variation in *Gypsophila fastigiata*. Journal of Ecology, 78: 313-325.
- Rotenberry, J. T. & J. A. Wiens, 1980. Habitat structure, patchiness, and avian communities in North American steppe vegetation: A multivariate analysis. Ecology, 61: 1228-1250.
- Stoermer, E. F., 1980. Characteristics of benthic algal communities in the Upper Great Lakes. EPA-600/3-80-073. U.S. Environmental Protection Agency, Large Lakes Research Station, Grosse Ile, Michigan.
- ter Braak, C. J. F., 1986. Canonical correspondence analysis: A new eigenvector technique for multivariate direct gradient analysis. Ecology, 67: 1167-1179.
- ter Braak, C. J. F., 1988. CANOCO: A FORTRAN program for canonical community ordination by [partial] [detrended] [canonical] correspondence analysis, principal components analysis and redundancy analysis (version 2.1). TNO Institute of Applied Computer Science, Wageningen.
- ter Braak, C. J. F., 1990. Update notes: CANOCO version 3.10. Agricultural Mathematics Group, Wageningen.
- ter Braak, C. J. F. & I. C. Prentice, 1988. A theory of gradient analysis. Advances in Ecological Research, 18: 271-317.
- Van Horne, B. & R. G. Ford, 1982. Niche breadth calculation based on discriminant analysis. Ecology, 63: 1172-1174.
- Van Valen, L., 1978. The statistics of variation. Evolutionary Theory, 4: 33-43.